The health-related microbial quality of drinking water from ground tanks, standpipes and community tankers at source and point-of-use in eThekwini Municipality:

Implications of storage containers, household demographics, socio-economic issues, hygiene and sanitation practices on drinking water quality and health.

by

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As the candidate's supervisor I have/have not approved this thesis/dissertation for submission

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Abstract

The aim of this study was to investigate the microbiological quality of drinking water at the source (taps at eThekwini laboratories, standpipes and mobile community tankers) and corresponding point-of-use (storage containers and ground tanks) supplied to peri-urban areas in Durban by eThekwini Municipality. It also aimed to identify factors associated with deterioration in water quality such as storage of water, household demographics, hygiene and sanitation practices. In order to determine the microbial quality of drinking water, the pour plate method (for enumeration of heterotrophic organisms) and the membrane filtration technique (for total coliforms and *E. coli* enumeration) were used. Conductivity, turbidity, pH and total and residual chlorine levels of drinking water were measured. Microbial and physico-chemical data was collated and statistically analysed with epidemiological data from an associated study to determine the link between microbial quality of drinking water, household demographics, health outcomes, socio-economic status, hygiene and sanitation practices. Findings showed that all point-of-use water was unsafe for human consumption as a result of either poor source water quality, in the case of standpipes, and microbial contamination at the point-of-use, in the case of ground tanks and community tankers. The latter could be attributed to unsanitary environments, poor hygiene practices or poor wateruse behaviour. Households which included children aged 0-5 years and in which open-top containers were used for water storage had the highest rates of diarrhoea and vomiting. Water from ground tanks had the best microbial quality but people in households using this water presented with the highest rate of diarrhoea. Therefore provision of microbially safe drinking water will not reduce the rate of health outcomes if addressed in isolation. In order to reduce water-associated illness, provision of safe and adequate amounts of water, hygiene and sanitation education and education on water-use behaviour should be provided as a package. The provision of improved water delivery systems does not ensure that drinking water is safe for human consumption. Measures, such as point-of-use water treatment should be considered to ensure that drinking water provided at the source and point-of-use is microbially safe for human consumption.

Preface

The experimental work described in this dissertation was carried out in the School of Biological and Conservational Sciences, University of KwaZulu-Natal, Durban, from February 2005 to April 2007, under the supervision of Professor Michael Smith and Dr Nicola Rodda.

This study represents the original work by the author, Ms Urisha Singh, student number 202501371, and has not been submitted in any form for any degree or diploma to any tertiary institution. Where use has been made of the work of others, it is duly acknowledged in the text.

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Declaration 1 – Plagiarism

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Abbreviations

Abbreviation	Definition of abbreviations	Abbreviation	Definition of abbreviations	
ANOVA	Analysis of Variance	LSD	Lowest Significant Difference	
AVG	Average	MDG	Millennium Development Goal	
CFU	Colony forming unit	NR	Negligible Risk	
Cl	Chlorine	NRF	National Research Foundation	
DWAF	Department of Water Affairs and Forestry	OCL.	Hypochlorite ions	
E. coli	Escherichia coli	PR	Potential Risk	
ECOSAN	Ecological Sanitation	PVC	Polyvinyl Chloride	
EHEC	Enterohaemorrhagic E. coli	SR	Substantial Risk	
EIEC	Enteroinvasive E. coli	SD	Standard deviation	
EPEC	Enteropathogenic E. coli	STEC	Shiga toxigenic E. coli	
ETEC	Enterotoxigenic E. coli	UD	Urine Diversion	
EWS	eThekwini Water Services	UNICEF	United Nations Children's Foundation	
GM	Geometric mean	VIP	Ventilated Improved Pit latrine	
HCL	Hydrochloric Acid	WASH	Water and Sanitation Hygiene	
HOCL	Hypochloric acid	WHO	World Health Organisation	
НРС	Heterotrophic plate count			

List of SI units

List of Units

mg/L	Milligrams per litre
mS/m	Millisiemens per metre
NTU	Nephelometric Turbidity Unit
Cfu/mL	Colony Forming Unit per millilitre

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CHAPTER 1: RATIONALE OF THE STUDY

1.1 Context

South Africa is a developing country which lacks adequate supplies of potable water and sanitation for its population. The lack of infrastructure coupled with rapid population growth in certain areas is a major contributing factor to this problem. Areas with rapid population growth include the low-income rural and peri-urban settlements in South Africa. Where improved drinking water ¹ has been provided to such communities, contamination of water during storage, collection and transportation is of concern since contamination would result in water that is unsafe for human consumption (Jagals *et al.* 1997; Jagals *et al.* 2004).

The World Health Organisation (WHO) and United Nations Children's Foundation (UNICEF) addressed issues pertaining to health, poverty, lack of sanitation and potable water in the Millennium Development Goals (MDGs). These goals, which serve to address the world's main development challenges, have been agreed upon by 189 nations and signed off by 147 heads of state and governments as part of the Millennium Declaration. The seventh goal, to ensure environmental sustainability, has as one of its objectives, to reduce by half the proportion of people, worldwide, without sustainable access to an improved safe water source, in urban and rural areas (UNICEF, 2007).

In order to address one of South Africa's main development challenges, government introduced the South African Free Basic Water Policy (DWAF, 2002). This policy aimed to reduce the number of South Africans without access to safe drinking water (Mosdell and Leatte, 2005). The policy states that every household is entitled to 6000 L of free water per month (25 L per person per day in a household of 8 people). Through provision of safe drinking water, government also aimed to reduce child mortality and waterborne infections. Therefore an array of South Africa's development challenges was aimed at being achieved through addressing the challenge of lack of adequate supplies of drinking water to the population (Mosdell and Leatte, 2005).

¹ Described as water that is supplied from a protected source, *i.e.* a source that receives and supplies drinking water from a municipal drinking water treatment facility (WHO, 2005).

Having the responsibility to serve the city of Durban and its surrounding areas with drinking water, eThekwini Municipality undertook to increase the access to safe drinking water and hence address the South African Free Basic Water Policy in Durban, South Africa (eThekwini Municipality, 2005). The eThekwini Municipality covers an area of approximately 2300 km² and serves approximately 3.1 million people with drinking water (KwaZulu-Natal Municipalities, 2007). Low-income areas, such as rural and peri-urban areas, unlike urban areas, lack infrastructure for waterborne sewerage and high-pressure water delivery systems (in-house taps). To differentiate between areas with and without proper infrastructure for waterborne sewerage and high-pressure water delivery systems the Municipality drew a waterborne edge around the city known as the Waterborne Sewered Ring. Areas beyond the Waterborne Sewered Ring would not receive waterborne sewerage or in-house taps. Instead they are provided with Ventilated improved pit latrines (VIP)² or Urine Diversion (UD)³ toilets for sanitation purposes. To address the MDG of reducing by half the proportion of people, without sustainable access to an improved safe water source, in rural and peri-urban areas in Durban, South Africa, eThekwini Municipality is supplying water to these areas via three main systems; ground tanks⁴, standpipes⁵ and mobile community water tankers⁶ (KwaZulu-Natal Municipalities, 2007).

Water in communal standpipes and mobile community water tankers require users to collect, transport and store water prior to use whilst ground tanks themselves serve as a storage container. In the case of standpipes and mobile community water tankers, water is collected,

 $^{^{2}}$ A form of on-site sanitation in which urine and faecal waste collect into an underground pit. This waste is broken down by bacterial interactions within the pit (WHO, 1997).

 $^{^{3}}$ A form of on-site sanitation in which, urine is kept separate from faecal waste. Faecal waste is kept dry and allowed to decompose. The decomposed material and collected urine can later be used for soil conditioning (Esrey *et al.* 1998).

⁴ An on-site low-pressure water delivery system comprised of a storage tank that holds 200L of drinking water. Water is supplied via the municipal water distribution system. Tanks are placed outside households and serve individual households (eThekwini Municipality, 2005).

⁵ A semi-pressure water delivery system comprised of a free standing pipe which is fitted with a tap. These structures are located outside households and serve several households, within a radius of 200m, with drinking water. It is a communal water delivery system and requires that water be collected, transported and stored in households for use. (eThekwini Municipality, 2005).

⁶ A low-pressure, communal water delivery system comprising a specialized water storage vessel attached to a vehicle. The vessel is designed for transporting water from a water treatment plant to peri-urban and rural areas. Water is collected in containers, transported and stored in households till required for use.

transported to and stored in households in either open-top⁷ or closed-top⁸ portable storage containers. Numerous studies have demonstrated the deterioration in microbiological quality of drinking water as it moves from source⁹ to point-of-use¹⁰ (Moyo *et al.* 2004; Trevett *et al.* 2005; Gundry *et al.* 2006). Such deterioration in water quality can cause point-of-use water to be unsafe for human consumption (Jagals *et al.* 1997; Mirza *et al.* 1997; Momba and Notshe, 2003; Moyo *et al.* 2004; Trevett *et al.* 2005; Gundry *et al.* 2004; Trevett *et al.* 2005; Gundry *et al.* 2004; Trevett *et al.* 2005; Gundry *et al.* 2006). Factors contributing to deterioration in water quality include poor hygiene and sanitation practices, the use of contaminated containers to store and transport water, insertion of dirty hands into water, contact of water with particulate matter, animals and insects as a result of openings in containers, and poor environment surrounding source water (Verweij *et al.* 1991; Jagals *et al.* 1997; Hoque *et al.* 1999; Roberts *et al.* 2001; Momba and Notshe, 2003; Trevett *et al.* 2004; Trevett *et al.* 2001; Momba and Notshe, 2003; Trevett *et al.* 2004;

Water stored in open-top containers is more prone to faecal contamination than water stored in closed-top containers and may contain faecal pathogens that could result in illness such as diarrhoea (Hammad and Dirar, 1982; Deb *et al.* 1986; Empereur-Bissonnet *et al.* 1992; Jagals *et al.* 1997; Mirza *et al.* 1997; Roberts *et al.* 2001). Dipping utensils and hands of users may be faecally contaminated and may thus result in faecal contamination of water (Echeverria *et al.* 1987; Pinfold, 1990; Hoque *et al.* 1995; Islam *et al.* 2001; Trevett, 2003; Trevett *et al.* 2005). Water stored in closed-top containers has also been showed to become faecally contaminated as a result of techniques used to remove water from containers. Such techniques include dipping of utensils into water or direct insertion of hands into water (Swerdlow *et al.* 1992; Swerdlow *et al.* 1997).

⁷ A storage container which does not have a lid. It is therefore left open. Water that is stored in such containers is in direct contact with environmental factors and is therefore easily contaminated than water stored in containers with a cover (eThekwini Municipality, 2005).

⁸ Containers which can be closed *i.e.* it has a lid. Water stored in these containers are less prone to contamination as they are more protected from environmental contaminants than water stored in open-top containers (Jagals *et al.* 1997; Trevett *et al.* 2004; eThekwini Municipality, 2005).

⁹ The origin or starting point from which water supplied to a specific water delivery system or storage container originates or stems from. This is representative of water from the water treatment works (Jagals *et al.* 1997; Trevett *et al.* 2004; eThekwini Municipality, 2005).

¹⁰ The actual point at which consumers use water from. This point is supplied with drinking water by the source (WHO, 1997).

Deterioration of point-of-use drinking water quality has also in the past been associated with the age of household members (Deb *et al.* 1986; Yeager *et al.* 1991; Qadri *et al.* 1992; Trevett *et al.* 2004). Households including children aged 5 years or younger have been shown to have higher counts of faecal pathogens and other microbes in point-of-use drinking water than households with all other age groups (Roberts *et al.* 2001). This has been shown to be especially prevalent in areas where open defaecation is practiced, since children usually have direct access to these areas. Insertion of contaminated hands or dipping utensils into water storage vessels when removing water for use can result in faecal contamination of household drinking water associated with the presence of children in a household (Hoque *et al.* 1995; Roberts *et al.* 2001; Trevett, 2003; Trevett *et al.* 2005). The presence of elderly people (greater than 50 years of age) in households has also been associated with increased microbial content of point-of-use drinking water possibly due to poor hygiene practises (Eisenberg *et al.* 2001).

Poor microbial quality of drinking water is linked to various health conditions, most typically manifesting as diarrhoea, vomiting and gastroenteritis (Chanlett, 1992; American Society for Microbiology, 2002). Diarrhoeal disease has been documented to account for 4.3% of the total global disease burden, in which 88% of cases is caused by poor quality drinking water, poor hygiene and inadequate sanitation. In South Africa, death due to diarrhoeal disease claims the fourth highest number of infant lives, only exceeded by HIV/AIDS, low birth weight and perinatal complications (Bradshaw *et al.* 2003). Thus microbial contamination of drinking water poses a risk of infection to users and needs to be controlled effectively.

The link between microbial drinking water quality and human health has been questioned. Whilst some studies, such as those by, Payment *et al.* (1991), Payment *et al.* (1993), Pinfold *et al.* (1991), Quick, (1997) and Chidavaenzi *et al.* (1998), have shown that good microbial quality of drinking water is related to a reduction in health outcomes, other studies by Esrey *et al.* (1985), Esrey *et al.* (1991) and Payment *et al.* (1993) suggest that a reduction in health outcomes is more likely to be achieved through the provision of good quality and quantity of water in conjunction with proper hygiene practices and good sanitation, rather than through the provision of microbially safe drinking water alone.

1.2 Aims of this study

The main objective of this study was to investigate the microbiological quality of stored water in eThekwini Municipality supplied by ground tanks, standpipes and community tankers. The study aimed to identify some of the factors associated with deterioration in water quality.

The specific objectives were to:

- Determine the microbial quality of drinking water from ground tanks, communal standpipes and mobile community water tankers at the source and at the corresponding point-of-use.
- Determine the implications of type of storage container used (open-top or closed-top) on the microbial quality of drinking water from communal standpipes and community water tankers.
- Determine the impact of household demographics (*viz.* age distribution of household members) on microbial quality of drinking water at the point-of-use (water stored in open-top containers, closed-top containers or ground tanks)
- Evaluate the microbiological quality of drinking water in light of water quality, sanitation and hygiene education provision.

1.3 Structure of dissertation

This dissertation comprises 6 chapters. In addition to the current chapter a review of literature in the field is presented in Chapter 2. Chapter 3 describes methodology followed by individual chapters on results (Chapter 4), discussion (Chapter 5) and conclusions and recommendations for future research (Chapter 6). References and appendices follow Chapter 6. A study on the effects of seasonality on water quality is presented in the Appendix A. This has been included as an appendix due to the use of different sampling sites for different seasons. It is acknowledged that the comparison of different communities impacts the quality of the study presented on seasonality hence this work is presented in the appendix and not in the main body of this dissertation.

1.4 Acknowledgement of data sources

This dissertation is a microbial water quality study. It included the use of data from two previous studies, with due permission of the owners of each database, as well as data collected by the author. Firstly an Honours degree study in the field of microbiology, by Swasti Maraj, completed in 2005 at the University of KwaZulu-Natal. The information used from this study is part of a database owned by the University of KwaZulu-Natal and is available to the public domain, therefore written consent for use of this database has not been included in the appendices. Secondly, a Masters of Science study in the field of epidemiology completed by Renuka Lutchminarayan in 2007. The epidemiological data used was from the Ecological Sanitation (ECOSAN) database owned by the Department of Public Health Medicine. Permission for use of this database was received from Dr Stephen Knight (Appendix B). All microbial information pertaining to Cato Manor ground tanks in this study was provided by Ms Swasti Maraj. All statistical analysis performed on these combined data sets were conducted by the author in conjunction with Ms Tonya Esterhuizen, a statistician at the School of Health Sciences, Nelson R Mandela School of Medicine and Ms Jaclyn Kelly Wright, a postgraduate student with experience in biostatistics at the Nelson R Mandela School of Medicine. All microbial and physico-chemical data for Sawpitts and Mtamuntengayo was obtained and statistically analysed by the author.

CHAPTER 2: LITERATURE REVIEW

Access to safe and reliable drinking water supplies is one of the key factors for determining health status. In 2002, diarrhoeal disease due to unpotable water and poor sanitation accounted for the second largest number of deaths in children aged 1-5 years (Mara, 2006). At that time, 1.1 billion people worldwide still required access to safe drinking water. According to more recent estimates, 25% of people lacking safe drinking water are found in sub-Saharan Africa, 40% in East Asia and 19% in South Asia (Warby, 2007).

The WHO, UNICEF and other leading organisations worldwide introduced the MDGs in an attempt to address the world's main development challenges. The MDGs have been extracted from the actions and targets of the Millennium Declaration which was approved by 189 nations and signed by 147 governments and heads of state in September 2002 at the UN Millennium Summit (UNICEF, 2007). There are 8 MDGs, each aimed at improving the quality of life of the poorest people worldwide. These goals are:

- 1: elimination of extreme poverty and hunger;
- 2: attainment of universal primary education;
- 3: encouragement of gender equality and empowerment of women;
- 4: reduction of child mortality;
- 5: improvement of maternal health;
- 6: elimination of HIV/AIDS, malaria and other diseases;
- 7: improvement of environmental sustainability; and
- 8: introduction of a Global Partnership for development.

The MDGs consist of 18 quantifiable targets that are measured by 48 indicators. These goals are required to show improvement from statistics in 1990 and are to be reached by 2015. The focus in this study is goal 7 and goal 4. The targets of goal 7 are to reduce by half the proportion of the population living without sustainable access to safe drinking water and sanitation. An indicator for this target is the proportion of people who have access to a safe water supply and sanitation facilities (UNICEF, 2007). The target of goal 4 is to reduce by two thirds the mortality rate among children less than five years of age. Diarrhoea has been demonstrated to be one of the primary causes of death in children less than five years of age.

Diarrhoea in children less than five years of age is commonly caused by poor microbial quality of drinking water, inadequate sanitation, poor hygiene, and poor water-use behaviour at the point-of-use (Deb *et al.* 1986; Empereur-Bissonet *et al.* 1992; Eisenberg *et al.* 2001; Trevett *et al.* 2005). The present study focuses on the microbial quality of drinking water being supplied to communities by a range of water delivery systems, within the eThekwini Municipal area. It also investigates the subsequent impact of water quality on health in children aged 0-5 years, children aged >5-18 years and in adults (*i.e.* household occupants >18 years of age).

South Africa is continuing to address its development challenge regarding lack of adequate water supplies through the Free Basic Water Policy which was introduced in year 2000 (DWAF, 2001). This policy was introduced to address the government's constitutional obligation to ensure provision of basic services (drinking water and sanitation) to all South Africans. Its main objective was to deal with poor service delivery levels of drinking water which contributed significantly to disease burden in poor communities (Brocklehurst, 2005). It therefore aimed to reduce the number of people without access to safe drinking water and thus reduce waterborne disease and child mortality (Mosdell and Leatte, 2005). The policy stated that every household in South Africa is entitled to 6000 L of free water per month (25 L per person per day in a household of eight people) (DWAF, 2001). Through the introduction of this policy the number of South Africans with access to safe drinking water, dwelling in rural areas, has increased from 8.6 million in 1999 to 15.7 million in 2008 (DWAF, 2008). The total population of rural dwellers in 1999 was approximately 17.9 million whilst in 2008 it was approximately 19.9 million (DWAF, 2008). Therefore, the aim to reduce by half the proportion of people without access to safe drinking water, in the context of rural areas in South Africa, has been achieved. The 2008 Millennium Development Goal progress report for South Africa shows that the proportion of urban populations with access to an improved water source¹¹ has exceeded the MDG target of 85.2% and currently stands at 94.8%. The proportion of rural populations with access to safe drinking water has increased from 44% in 1994 to 77.8% in 2008 and has also exceeded the MDG target of 72.2% (Lehohla, 2007; UNICEF, 2007; DWAF, 2008). Therefore the South African government has thus far been successful in implementing and receiving good service delivery results through the execution of the South African Free Basic Water Policy in South Africa.

¹¹ A source that receives and supplies drinking water from a municipal drinking water treatment facility. An improved water source may also be referred to as an improved water supply (WHO, 2005).

In Durban, South Africa, eThekwini Municipality addressed the Free Basic Water Policy by providing drinking water to rural and peri-urban areas via standpipes, ground tanks and mobile community tankers. To further address this policy and the MDG number 7, eThekwini Municipality introduced a 3-pronged intervention programme. This programme ensured that every household which received a ground tank also received a UD toilet and hygiene education through the Water and Sanitation Hygiene program (WASH). During the course of this programme individuals were educated on care and use of UD toilets and ground tanks. They were also educated on the importance of good hygiene practices such as washing of hands with soap after using toilets and before cooking and eating. Water supply systems and sanitation systems used in Durban, South Africa are explained below (eThekwini Municipality, 2005).

2.1 Water supply systems in eThekwini Municipality, Durban: South Africa

Consumers provided with water by eThekwini Municipality, are supplied via mobile community water tankers (community tankers), communal standpipes (standpipes), ground tanks, semi-pressured roof tanks and high-pressure taps (eThekwini Municipality, 2005). In this study only ground tanks, communal standpipes and mobile community water tankers have been used. Brief explanations on all the water delivery systems are however included in the literature. Mobile community water tankers, communal standpipes, ground tanks, semi-pressured roof tanks and high-pressure taps are divided into three types of delivery systems: high-, semi- and low-pressure water delivery systems (Figure 1).

The high-pressure system consists of high pressure taps and communal standpipes (Figure 1). Unlike communal standpipes, high-pressure taps is least susceptible to external contamination as water is not stored or transported outside the distribution system and hence introduction of external contaminants is minimised. Semi-pressure systems including roof tanks are more susceptible to external contamination than high-pressure systems since water from these sources are usually stored, transported or exposed to environmental factors that impact on microbial water quality. Low-pressure systems are comprised of mobile community water tankers and ground tanks (Figure 1). Low-pressure systems are generally most prone to microbial contamination due the storage periods of water either in storage containers or within ground tanks themselves (eThekwini Municipality, 2005).

Water from the treatment plant is distributed to mobile community tankers, communal standpipes and on-site ground tanks (Figure 1). Mobile community tankers and communal standpipes are sources of drinking water and the corresponding point-of-use for these delivery systems is household storage containers which may be either open-top or closed-top. With regard to the ground tank delivery system, the on-site ground tank is the point-of-use as it is found in the yard of individual households and hence does not require collection of drinking water. The source for this delivery system would be any point in the municipal distribution system (explained further in Figure 2, Chapter 3).

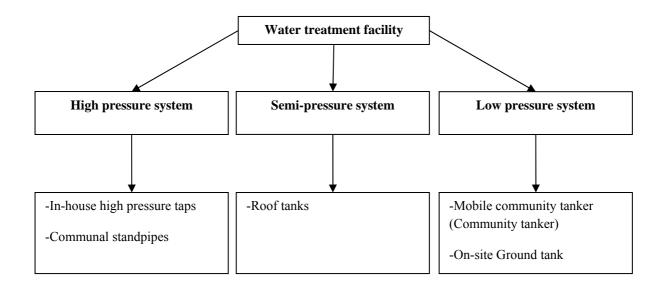


Figure 1 Schematic diagram representing the three types of water delivery systems which receive water from treatment facilities. The water delivery systems include the high-pressure, semi-pressure and low-pressure water delivery systems.

2.1.1 High-pressure taps

These water delivery systems are found in fully developed areas, where a functioning piped distribution system and waterborne sanitation exist. (Durban Metro Water, 2002; eThekwini Municipality, 2005). Contamination of water in this system is not as widespread as that of the low-pressure delivery systems as water is delivered directly from the water treatment plant via pipelines to taps in the house. No storage of water is involved, therefore chances of contamination through transportation and storage of water is minimised. This system was not included in the focus of this study, so will not be considered further.

2.1.2 Roof tank water delivery system

Roof tanks can supply in excess of the 200 L of free basic water per day, *i.e.* after supplying 6000 L (200 L x 30 days = 6000 L) of free basic water for the month all other water used is billed for. It works on a meter basis whereby, once the 6000 L of free water is drawn, all further water used is billed for. Tanks are continually filled with water, which they receive from a water treatment plant via pipelines. (Durban Metro Water, 2002, eThekwini Municipality, 2005). This system is not included in the focus of this study, so will not be considered further.

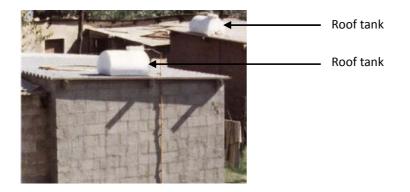


Figure 2 Illustration of a water storage tank (roof tank) located on the roof of a household.

2.1.3 Communal Standpipes (Standpipes)

Standpipes are a communal high-pressure water delivery system which is typically used in informal and peri-urban settlements (DWAF, 1994). A single standpipe can serve several hundred people with water on a daily basis. The Department of Water Affairs and Forestry Water Supply and Sanitation Policy (1994) states that standpipes should be located no further than 200 m from a household, although it is recognized that there are still areas where this is far from being achieved (DWAF, 1994; Jagals *et al.* 1997). Water is collected from standpipes in containers and transported to households, where it is stored and used. Water from standpipes is drawn from a distribution pipe with a tap (Figure 3). This water is prone to contamination because of collection, transport and storage methods used at the household level (Jagals *et al.* 1997). Storage and collection containers are often contaminated with microorganisms which adhere to the container surfaces. In this way bacteria can enter water used for drinking purposes, amongst others, and can represent a health hazard. The organic compounds in water and material used to manufacture storage and collection containers, such as polyethylene, may serve as a source of nutrients for bacterial growth and proliferation,

thereby contaminating water and placing consumers at risk (Momba and Mnquemvu, 2000; Momba and Kaleni, 2002; Momba and Notshe, 2003).

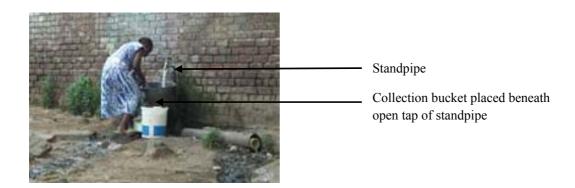


Figure 3 A resident filling a bucket with water from a communal standpipe.

2.1.4 Ground tanks

Ground tanks are located in the yard of households. Houses supplied with ground tanks are equipped with on-site sanitation services such as VIP or UD toilets. Ground tanks are only used in areas where the plot size is large enough to accommodate a hydraulic load of 200 L per day. It is found mainly in rural and peri-urban areas. Ground tanks are made of polyvinyl chloride (PVC) and can hold up to 200 L of water. They are re-filled with treated water every evening through a piping system from the treatment works operated via an electronic switch mechanism. A tap is present at the bottom of each tank and serves to release water from the tank when required (Figure 4) (Durban Metro Water, 2002; eThekwini Municipality, 2005). Water in ground tanks is prone to contamination since tanks are used as the container in which water is stored in for 24 hours. Water and the container surface can provide nutrients for growth and proliferation of bacteria, which may either be found floating in water (planktonic bacteria) or attached to the inner surface of containers (biofilm) (Lehtola et al. 2004). Bacteria from biofilm may be released into water when shear forces are applied to biofilm layers, such as when the tank is refilled (LeChevallier et al. 1980; LeChevallier et al. 1991; Block et al. 1993). This bacterially contaminated water may then be used by the consumer for drinking purposes and could represent a potential health hazard if bacteria are pathogenic. Since the recent increase in free basic water provision to 300 L per day, ground tanks will gradually be phased out and replaced with electronic bailiffs (devices attached to distribution pipes to limit water supply to free basic water provision). It may be expected that this will decrease opportunities for contamination. The use of UD toilets in combination with ground tanks, as implemented by eThekwini Municipality to some peri-urban areas in Durban, has been shown to reduce incidences of diarrhoea, vomiting and worm infestations in a study by Lutchminarayan in 2007. This is thought to be largely due to the better hygiene practices associated with the use of this sanitation system as a result of the provision of sanitation, water and hygiene education provision as a package.



200 L Ground tank placed on an elevated surface

Pipe through which ground tank is refilled with treated water

Tap on ground tank through which water is released

Buckets in the vicinity of ground tanks.

Figure 4 Ground tanks located in the yard of a house. A pipe allows for the tank to be re-filled with water from the municipal distribution system every 24 hours. A tap at the lower end of the tank allows for water to be released from the tank as required.

2.1.5 Mobile Community water tankers (Community tankers)

A mobile community water tanker (hereafter referred to as community tanker) is a lowpressure communal water delivery system which consists of a large tanker attached to a vehicle (Figure 5) (eThekwini Municipality, 2005).

This tanker is filled with treated water from the water treatment works. It transports water to target areas, typically those not yet supplied with standpipes or ground tanks. Community members then collect water directly from the tanker via a tap on the tanker. Water is then transported to and stored in households for use (eThekwini Municipality, 2005).

Water from community tankers is also susceptible to contamination during collection, transportation or storage, or as a result of poor hygiene within the tanker since the interior is difficult to access and hence clean (eThekwini Municipality, 2005).



Mobile community tanker filled with drinking water from a collection point at the municipal treatment plant.

Pipe system through which the tanker is emptied and refilled.

Figure 5 Illustration of a mobile community tanker. Tankers are filled with treated water at the treatment plant. Thereafter it is transported to target areas were community members collect water from tankers via a tap. This water is then transported to and stored in households for use.

2.2 Sanitation systems in eThekwini Municipality, Durban: South Africa

Ventilation improved pit latrine systems and UD toilets are the two major types of sanitation used by rural and peri-urban dwellers in Durban. The UD toilets were introduced at the same time as ground tanks for water supply. eThekwini Municipality provided ground tanks and UD toilets as a package in order to promote good sanitation practices and hence maintain the microbial and sanitary quality of drinking water. Pit latrines are being replaced by UD toilets due to the contamination of ground water and accessibility problems for emptying VIP toilets (eThekwini Municipality, 2005)

2.2.1 Ventilation improved pit latrines (VIP)

In pit latrines, there is presently a mix of older pit latrines of various constructions and of VIP latrines since being implemented in 1994. Faeces and urine collects into a common reinforced pit (Figure 6). The waste material in this pit is broken down by natural bacterial reactions. The additional features of a VIP latrine includes a ventilation pipe which extracts air from the toilet when wind blows in order to prevent odour, it also includes an enclosed pit (Buckley *et al.* 2008). The problem with these toilets however, is that the wet system results in odour which is not controlled solely by the ventilation pipe. This attracts flies to the area and is a health concern (WHO, 1997). It is due to such issues with this system that the UD toilet system was developed and implemented (Winblad, 1996).



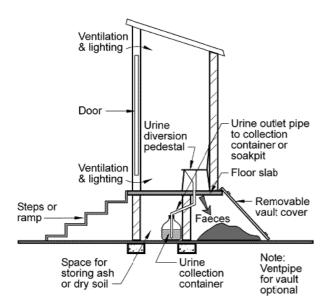
Figure 6 Illustration of a ventilated improved pit latrine.

2.2.2 Urine diversion toilets (UD toilets)

Urine diversion toilets make use of technology in which urine is separated from faecal waste so as to keep faecal waste dry (Esrey et al. 1998). This is facilitated by a special design of the toilet (Figure. 7) in which urine is diverted via a pipe / tube into a collection container or soak away and faecal waste is collected into a vault immediately beneath the toilet (Esrey et al. 1998). These toilets may have a single vault or a double vault. In Durban UD toilets are double vaulted. This allows for the decomposition of the contents of the first vault whilst the second vault is being used, and vice versa. After defaecation sand is spread over faeces to absorb moisture and control odour. It is important to keep faecal waste dry to promote dessication of faecal material and to control odour and flies. The dehydration of faecal waste helps in the destruction of harmful bacteria and viruses over time (Esrey et al. 1998). This dehydrated, decomposed faecal material can then be used for soil conditioning, although eThekwini Municipality instructs users to bury the vault contents on site. Included in the design of the UD toilet structure is also a sink, which can be used for washing of hands thus promoting hygiene practices in an attempt to reduce incidences of diarrhoea and vomiting caused by ingestion of faecal bacteria through the oral-faecal route. Found on the door of each UD toilet in Durban is a flyer with instructions on the proper use of UD toilets and associated hygiene practices, thus encouraging and educating users on proper handling and use of UD toilets (eThekwini Municipality, 2005; Lutchminarayan, 2007).

There are several advantages linked to the use of UD toilets in comparison to the use of VIP latrines. Firstly, UD toilets, if used correctly, have fewer odours and flies than VIP toilets and

are thus conducive to installation within homes. Secondly the vault can be constructed above ground and hence will not affect soil quality or pollute groundwater. Thirdly, the faecal compost and urine generated from these toilets may be used in rural and urban agriculture. Fourthly, no water is required to use this system and lastly, no expensive treatment of sewage is required (Morgan, 1999; SUDEA, 2000).



*(Source: www.cemforce.co.za)

Figure 7 Illustration of the structure of a urine diversion toilet.

2.3 Contamination of drinking water

The problem with water supplied by standpipes and community tankers however, lies in potential contamination of water during collection, storage and transportation, since these delivery systems are communal. For users who do not have continuous in-house private water services, these users are at risk of consuming microbially contaminated water because of contamination during transport and storage of drinking water (Momba and Notshe, 2003; Moyo *et al.* 2004; Trevett *et al.* 2005; Gundry *et al.* 2006). Inconsistency in hygiene and proper sanitation practices is a major contributing factor to such contamination (Echeverria *et al.* 1987; Pinfold, 1990; Swerdlow *et al.* 1992; Hoque *et al.* 1995; Tuttle *et al.* 1995; Roberts *et al.* 2001; Trevett *et al.* 2004).

Collection and storage of large quantities of water over a long period of time occur with accompanying potential for deterioration in water quality (Bailey and Archer, 2004; Genthe

et al. 1997; Gundry *et al.* 2004; Jagals *et al.* 1997; Jagals *et al.* 1999; Jagals *et al.* 2004; Wright *et al.* 2004). Studies conducted in the Ncera and Ntselamanzi rural villages in Harare, Zimbabwe, demonstrated that water supplied to communities via standpipes was contaminated by large numbers of pathogenic microorganisms (Momba and Kaleni, 2002). Contamination was attributed to collection, transportation, storage and water-use practices (Momba and Kaleni, 2002). In rural and peri-urban areas a single standpipe may serve hundreds of households. The location of standpipes may be viewed as being remote to certain users. Regulations state that standpipes should not be more than 200 m away from households using the water supply, but in some instances people travel up to 750 m for water (Momba and Kaleni, 2002).

Drinking water from communal delivery systems, such as standpipes and community tankers, is collected and stored in either open-top containers or closed-top containers. The type of storage container and the handling thereof impacts on the microbial quality of point-of-use drinking water (Shiffman *et al.* 1978; Jagals *et al.* 1997; Hoque *et al.* 1999; Trevett *et al.* 2004). Studies have shown that contributors to drinking water contamination include insects, and airborne and particulate matter that enters drinking water through openings in containers. Contaminated cloths used to wipe storage containers or taps during the collection and transfer of water, introduction of dirty hands into water, poor hygiene and sanitation practices and presence of animals also contribute to microbial contamination of collected drinking water (Feachem *et al.* 1978; Shiffman *et al.* 1978; Khairy *et al.* 1982; Heinanen *et al.* 1988; Verweij *et al.* 1991; Hoque *et al.* 1999; Roberts *et al.* 2001; Momba and Notshe, 2003; Trevett *et al.* 2005).

Utensils used to remove water from storage containers can also cause contamination of drinking water. Bacteria or particulate matter adhering to utensils can enter drinking water (Jagals *et al.* 1997). Storage vessels also offer a route for contamination as their inner surfaces may have microbes attached to it as a result of biofilm¹² development or poor hygiene of containers (Donlan and Pipes, 1988; LeChevallier *et al.* 1991; Tokajian *et al.* 2000; Arjun *et al.* 2004). Water in ground tanks is also susceptible to contamination through the formation of biofilm. Biofilm microorganisms utilise carbon and other organic nutrients

¹² A biofilm can be described as a structured community of microorganisms which are encapsulated in an extracellular polymeric matrix which they produce. Biofilms are found attached to surfaces that are in contact with water (Xavier *et al.* 2005).

from water and container surfaces for growth and proliferation, hence contamination of water occurs (LeChevallier *et al.* 1990; Mittleman, 1995). Insufficient hygiene education of water users is a major contributing factor for poor water quality since storage and collection containers are seldom washed and hence microorganism growth is not hindered (Swerdlow *et al.* 1992; Hoque *et al.* 1995., Swerdlow *et al.* 1997; Jagals *et al.* 1999; Medical Research Council, 1999; Coulson, 2000; Nala *et al.* 2003; Trevett and Carter, 2008).

The deterioration of water quality between the source and point-of-use is thus a serious concern (Clasen and Barstable, 2003; Trevett *et al.* 2004; Wright *et al.* 2004; Maraj *et al.* 2005; Trevett *et al.* 2005). Such microbial deterioration can result in increased incidence of adverse health outcomes such as diarrhoea and vomiting in children less than five years of age (Yeager *et al.* 1991; Qadri *et al.* 1992; Mirza *et al.* 1997; Mahmud *et al.* 2001).

The microbial quality of drinking water is established by testing for certain microorganisms in drinking water samples. These organisms are known as indicator organisms and they assist in determining the quality of drinking water eThekwini Municipality, 2005).

2.4 Water quality with microorganisms as indicators

Indicator microorganisms are used to evaluate the microbial quality of drinking water and its safety for human consumption. Safe drinking water can be defined as drinking water which is devoid of harmful chemicals or microorganisms that can cause illness in humans if present in certain concentrations (McFeters, 1990; Prescott *et al.* 1993; WHO, 2001). Due to the large number of microbes that can contaminate drinking water, it is impossible to analyse for every type of microbe. Therefore, a few microorganisms have been selected as indicators of water quality (WHO, 1993; DWAF, 1996; Water Research Commission, 1998; WHO, 2001). According to the South African Water Quality Guidelines (DWAF, 1996), for an organism to be classified as an indicator it needs to satisfy the following requirements:

- It has to be suitable for all water types.
- It must be present in polluted waters including sewage.
- It must be absent in unpolluted water.
- It must be present in numbers that correlate with the degree of pollution.
- It must be present in higher numbers than the pathogen.

- The survival of the organism in water must at least be as long as the pathogen's survival time.
- It must not be pathogenic or unsafe to work with in a laboratory and
- It must be detectable by practical and reliable methods (DWAF, 1996).

The most common indicators used in the analysis of drinking water quality include heterotrophic organisms, total coliforms, *E. coli* and somatic coliphages. The presence of indicator organisms in drinking water could be indicative of poor sanitary or general quality of water or it could indicate inefficiency in treatment processes (WHO, 2005). Indicator organisms used in this study included heterotrophic organisms, total coliforms and *E. coli*.

2.4.1 Heterotrophic organisms

Heterotrophic organisms consist of ubiquitously present microbial flora including *Acinetobacter, Aeromonas, Flavobacterium, Klebsiella, Moraxella, Serratia, Pseudomonas* and *Xanthomonas* species (Davies and McFeters, 1988). Heterotrophic organisms can proliferate rapidly and successfully in water or in biofilm. They are used as an indicator of general water quality in terms of increased general organic load. Following water treatment, the number of heterotrophic organisms should be low (between 0 and100 colony forming units per millilitre (cfu/mL) (DWAF, 2005; WHO, 2005). If heterotrophic organism counts are high after treatment (>100 cfu/mL) the inference is that treatment processes are inefficient (DWAF, 1996; DWAF, 2005). If heterotrophic organism counts are high in water from distribution system or that biofilm is present in either the distribution system or the storage containers. Environmental contamination of water could also have occurred. This leads to questioning of general water quality and to testing for more specific indicators (DWAF, 1994; DWAF, 1996; WHO, 2005) (Table 1).

2.4.2 Total coliforms

Total coliforms is a group of bacteria that are found in the environment, natural waters and in the faeces of all warm-blooded animals including humans. These bacteria are able to survive and grow in water and are hence used as indicators of water quality (Rompre *et al.* 2002). The presence of total coliforms in drinking water is indicative of treatment efficacy, regrowth and general hygiene status of water in distribution systems. The presence of total coliforms is

also used as an indicator of the potential presence of disease-causing microbes (pathogens) in drinking water. Following water treatment procedures, total coliforms should be absent in drinking water. Their presence points toward treatment inefficiency if found after treatment. If detected in stored water and distribution systems, biofilm presence, regrowth in distribution systems or contamination of water with faecal waste, soil or plant debris are suspected (DWAF, 1996; Momba and Kaleni, 2002; WHO, 2005).

2.4.3 Escherichia coli (E. coli)

E. coli is a species of bacteria found in human and animal faecal matter (Feng *et al.* 2002). Their presence in drinking water is indicative of faecal pollution of water and the potential presence of faecal pathogens. Presence could also indicate inefficient water treatment since water treatment processes (*e.g.* chlorination) are intended to remove faecally-derived bacterial pathogens. *E. coli* is sensitive to disinfection processes using chlorine (Edberg *et al.* 2008).

The WHO guidelines for drinking water state that for water to be safe for human consumption, heterotrophic organism counts must be between 0 and 100 cfu/mL whilst *E. coli* must be absent and total coliforms should be between 0 and 5 cfu/100mL (WHO, 1993 and WHO, 2005). According to South African Water Quality Guidelines, for drinking water to be safe for human consumption, heterotrophic bacteria must be present in counts lower than 100 cfu/ml, total coliforms must not exceed 5 cfu/100ml and *E. coli* should be absent from water (Table 1).

The table derived from the South African Water Quality Guidelines, given below, also indicates the level of risk of microbial infection posed to users of drinking water based on the specific counts of indicator microorganisms as per South African Water Quality Guidelines (DWAF,1996).

Table 1 Summary of South African target water Quality Guidelines for domestic water quality, acceptable counts of heterotrophic organisms, total coliforms and *E. coli* in drinking water.

Heterotrophic organisms (HPC cfu/mL)	Total coliforms (cfu/mL	E. coli (cfu/100mL)	Description
0-100	0-5	0	Negligible risk (NR) of microbial infection
100-1000	5-100	0-10	Potential risk (PR) of microbial infection. Indication of inadequate water treatment and possible post- treatment contamination and/or regrowth in the water system
>1000	>100	>20	Substantial risk (SR) of microbial infection. Indication of inadequate water treatment and possible post-treatment contamination and/or regrowth in the water system.

*Table derived from DWAF, 1996.

2.5 Water Quality - physico-chemical parameters

The physico-chemical parameters of drinking water, including the pH, temperature, total and residual chlorine levels, turbidity and conductivity, are important indicators of water quality. They can also influence the microbial quality of drinking water at the source and corresponding point-of-use (Momba and Kaleni, 2002; Momba and Notshe, 2003). Physico-chemical parameters used to determine the quality of drinking water are explained below.

2.5.1 pH

The pH of a substance can be described as the acidity or basicity of a solution (Norby, 2000). The pH of water has an effect on microorganism growth and also affects biofilm development. Guffanti *et al.* (1984) and Mayo and Noike (1994) showed that pH of water influences ion transport and biomass regulation in microorganisms. A pH closer to neutrality favours microorganism proliferation to a greater extent than a more basic pH. A pH closer to neutrality results in an increase in the metabolism of microorganisms and hence an increase in microbial growth and proliferation.

2.5.2 Temperature

One of the most important factors influencing bacterial growth in drinking water is temperature (LeChevallier et al. 1980; Donlan and Pipes, 1988; Donlan et al. 1994; LeChevallier et al. 1996; Momba and Notshe, 2003). The higher the temperature, the faster the regrowth of microorganisms (Tokajian et al. 2000). Microorganisms are incapable of regulating their own internal temperature and are dependent on ambient temperature and pH to influence their biomass composition, nutrient requirements, nature of metabolism and their rate of metabolic reactions (Pirt, 1971; Novak, 1974; Esener et al. 1981). An increase in temperature causes pH of water to approach neutrality and hence favours microorganism growth. Without the correct temperature, the rate at which organisms utilise substrates could be compromised and hence their growth and metabolism reduced. Microorganism growth is increased when temperatures reach 15°C or more (LeChevallier et al. 1996; Power and Naggy, 1999). Lund and Ormerod (1995) have also shown that biofilm formation in different drinking water systems are closely influenced by temperatures above 5°C. High temperatures also cause a reduction in residual chlorine and total chlorine levels, since chlorine volatilizes as Cl₂ gas more readily at higher temperatures. Therefore microorganism growth increases since the high temperature increases the metabolism of the organism and the low residual chlorine levels are not enough to destroy microbes (LeChevallier et al. 1980; Donlan and Pipes, 1988; LeChevallier et al. 1996; Donlan et al. 1994; Momba and Notshe, 2003).

2.5.3 Turbidity

Turbidity can be described as a measure of the amount of light that is scattered and absorbed by water as a result of suspended matter found in the water. It is used to quantify the amount of suspended solids in water (Allen *et al.* 2008). Turbidity in drinking water can be caused by suspended or colloidal matter including, silt, soil, clay, organic and inorganic matter and microorganisms (APHA, 2005). In drinking water systems where chemical disinfection (such as chlorine disinfection) is used, the type of turbidity rather than the amount of turbidity is more important. This is because organic matter, a source of turbidity, in drinking water can react with chemical disinfectants resulting in the production of disinfectant by-products which could potentially have long-term health effects on humans (Edberg *et al.* 2000). Turbid water is often brown and may have unpleasing aesthetic qualities (appearance, odour, taste). According to drinking water guidelines by DWAF (1996, 2005) and WHO (2005) the turbidity of drinking water should be below one Nephelometric Turbidity Unit (NTU) and must not exceed five NTU (Table 2). Above five NTU a discoloration in drinking water is noticeable. Higher values imply that water may be unsafe for human consumption. High values could be attributed to rusting pipes or storage vessels, or to inefficient treatment procedures. (Muyima and Ngcakani, 1997). Turbid water could favour microbial growth since suspended particles provide surfaces for attachment of microorganisms and higher nutrient content to support microbial proliferation and biofilm development (McCoy and Olsen, 1986; Miettinen *et al.* 1997; Sathasivan *et al.* 1997; Percival *et al.* 2000; Lehtola *et al.* 2004; Allen *et al.* 2008).

2.5.4 Free and residual chlorine

Chlorine is the most commonly used disinfectant, the purpose of which is to eliminate microbes from drinking water and prevent regrowth of microorganisms in drinking water systems and storage vessels (Gorchev, 1996). If chlorine levels are too low, proliferation of microorganisms will occur. According to WHO guidelines, each litre of drinking water should be treated with 2.5 mg of chlorine as the last stage of treatment (WHO, 2005). If temperatures are above 18°C, water should be allowed at least 30 minutes to react with chlorine. If temperatures are lower, a shorter reaction time is acceptable. Only 2 mg/L of chlorine is required to remove most bacterial contaminants from drinking water, the remaining 0.5 mg/L serving as residual chlorine (free chlorine) (WHO, 2005). The purpose of residual chlorine is to eradicate any contaminants that may enter drinking water during storage or transportation and to prevent regrowth of microbes in the distribution system and storage vessel. Therefore if chlorine levels are below the recommended levels at the source or point-of-use, microbial regrowth may not be controlled. The pH and turbidity of water affects the efficacy of chlorine (Gorchev, 1996). When chlorine is added to water it hydrolyses and yields hypochloric acid (HOCl) and hydrochloric acid (HCl). The hydrolysis to produce hypochloric acid is completed at a pH that is greater than 4. Since hypochloric acid is a weak acid, it partially dissociates into hypochlorite ions (OCI). In comparison to hypochlorite ions, hypochloric acid is a more efficient disinfectant and therefore a pH which favours a higher ratio of hypochloric acid to hypochlorite ions favours better disinfection of drinking water. At pH 6.5, approximately 90% of free chlorine is present as hypochloric acid whilst at a pH 9 the hypochlorite ions are more dominant. Therefore better disinfection occurs at a lower pH (Morris, 1982). Turbidity affects the disinfection efficacy of chlorine by providing a form of protection for microorganisms from chlorine. By doing so, an increase in oxygen and chlorine demand is created (WHO, 1996). The ideal conditions recommended by WHO for efficient chlorination is a pH of less than eight, turbidity between one and five NTU, a residual chlorine level of no less than 0.5 mg/L and a contact time of water with chlorine for at least 30 minutes (WHO, 1993; WHO, 1996)

2.5.5 Conductivity

Conductivity is measured as the electrical current which can be conducted by a water sample. Charged ions in water allow for the creation of an electric current. Conductivity increases as the concentration of ions increases. Hence conductivity is used as a measure of dissolved ions in water. Water with a high conductivity generally has a higher dissolved material content than water of low conductivity. The dissolved material could include contaminants or nutrients and therefore high conductivity values in drinking water are of concern since either microbes or substances that support microbial growth or microbial attachment could be present. Also, high conductivity could indicate the presence of salts or other ions which could cause encrustation or corrosion of distribution systems (WHO, 1979; Henley, 1995). According to WHO and DWAF (South Africa) guidelines for drinking water to be considered safe for human consumption, the physico-chemical parameters set out in Table 2 must be met. The failure of drinking water to meet these criteria implies that water may be unsafe for human consumption and hence can possibly cause adverse effects, particularly microbial infection.

Physico-chemical parameter	WHO guidelines 2005	DWAF guidelines 1996
Total Chlorine	2.5 mg/l	2 mg/l
Residual chlorine	0.2-0.5 mg/l	0.2-1.5 mg/l
Temperature	25°C	25°C
рН	7 - 8.2	5-9.4
Turbidity	Less than 1 NTU and	Less than 1 NTU and not
Turbiany	not greater than 5 NTU	greater than 5 NTU
Conductivity	< 100 mS/m	< 100 mS/m

Table 2	Guidelines	for 1	recommended	ph	vsico-che	mical	parameters in	drinking wa	iter.

*Table modified from: DWAF, 1996, DWAF, 2005, WHO, 1993 and WHO, 2005.

Microbial contamination of drinking water coupled with poor physico-chemical properties of water may result in the presence of microbes or pathogenic organisms in drinking water which could cause illness in humans. Vectors of water-associated diseases and health outcomes associated with the presence of such vectors are discussed below.

2.6 Waterborne diseases and health outcomes considered in this study

Water-related disease is a problem experienced worldwide. It is caused by contamination of water by animal, human or chemical wastes (WHO, 1993; WHO, 2005). There are four types of water-associated diseases: waterborne, water-washed, water-based and water-related diseases (Bradley, 1977). Waterborne diseases are caused by consuming food or drinking water contaminated with pollutants. The most common waterborne diseases include gastroenteritis, diarrhoea, typhoid fever, cholera, hepatitis and shigellosis. Pathogenic protozoa such as Giardia lamblia and Cryptosporidium parvum can also contaminate drinking water and cause diseases such as cryptosporidiosis and giardiasis, which can be lethal in severe cases (Bradley, 1977; Ford, 1999; Payment and Hunter, 2001). Among helminthic parasites, Ascaris lumbricoides is an intestinal parasite that infects approximately 25% of the world's population annually (Crompton, 1988). Water-washed diseases occur as a result of poor personal hygiene and skin contact with contaminated water. Examples are ascariasis, scabies and skin sores (Bradley, 1977; WHO, 1993). Water-based diseases are caused by parasites that are found in organisms such as bilharzias living in water. Waterrelated diseases occur as a result of insect vectors that breed in water. Malaria is an example of this type of disease which is transmitted by a mosquito vector (WHO, 1993; Bradley, 1977; Peterson et al. 1998).

The current study uses occurrences of diarrhoea and vomiting to illustrate a link, if present, between microbial quality of drinking water, hygiene, sanitation, socio-economic status, poverty index and health. Diarrhoea and vomiting are symptoms of several water-associated diseases, including those caused by vectors as described below.

2.6.1 Vectors of water-associated disease

Diarrhoea is not in itself a disease; it is rather a syndrome/symptom of several diseases. Diarrhoea can occur as a result of gastroenteritis which is commonly caused by viruses, bacteria or protozoa (Wilson, 2005). Examples of diarrhoeal diseases¹³ include; cholera, typhoid, bacillary dysentery (shigellosis), giardiasis and cryptosporidiosis (Water Aid, 2008). Diarrhoea is also a manifestation of enteric viruses such as norovirus. Diarrhoea can be defined as the release of three or more loose or watery stools in a 24 hour period (Baqui *et al.*, 1991). Individuals with diarrhoea suffer rapid depletion of water, sodium and ions from their bodies. If more than 10% of the body fluid is lost per day the individual dies (Water Aid, 2008). If diarrhoea persists for more than 3 days dehydration may occur, thus resulting in severe health impacts and even death in extreme cases (Baqui *et al.* 1991; Water Aid, 2008). Several studies have used the incidence of diarrhoea as an indicator of poor hygiene practices and environmental conditions including poor water quality and sanitation.

2.6.1.1 *E. coli*

There are several different strains of *E. coli* that cause diarrhoea. The principle subgroups are Enteropathogenic E. coli (EPEC), Enteroinvasive E. coli (EIEC), Enterotoxigenic E. coli. (ETEC), Enterohaemorrhagic E. coli (EHEC) and Shiga toxigenic E. coli (STEC) (Centre for Disease Control, 2008a). The EPEC subgroup is a common cause of diarrhoea in children whilst EIEC causes illness similar to shigellosis in humans. The ETEC subgroup causes travellers' disease, in which bacteria enter the cells of the small intestine and release enterotoxin. This results in abdominal cramps, vomiting and diarrhoea. The STEC subgroup of E. coli produces shiga toxins whilst the EHEC strain is associated with enterohaemorrhagic colitis in humans. Both these subgroups have associated virulence factors. E. coli O157 is the most common serotype of E. coli that is associated with EHEC. It is also the most dangerous and causes severe diarrhoea. In some cases it causes haemolytic uraemic syndrome and death (American Society for Microbiology, 2002; Centre for Disease Control, 2008d). Haemolytic uraemic syndrome is a disease which results in acute renal failure as a result of an inflammatory response that occurs upon exposure of the renal endothelium to shiga toxin. Disseminated intravascular coagulation occurs and the fibrin mesh that is formed captures thrombocytes and destroys red blood cells leading to a reduction in both counts. Its peak incidence is in children between 4 months to six years of age (Corrigan and Boineau, 2001).

¹³ Diarrhoeal disease refers to a disease in which diarrhoea is a symptom or manifestation (Wilson, 2005).

2.6.1.2 Vibrio cholerae

Vibrio cholerae is the causative agent of cholera. Cholera can be described as an acute diarrhoeal illness which occurs upon consumption of drinking water or food contaminated with faecal waste containing cholera-causing bacteria, *Vibrio cholerae* (Centre for Disease Control, 2008b). Symptoms occur after ingestion of the bacterium contained in contaminated water or food. The bacterium enters the small intestine and produces a toxin which is a potent stimulator of adenylate cyclase. This causes secretion of watery fluids, rich in sodium, potassium and bicarbonate from the intestine. In severe cases the rate of loss of these nutrients exceeds the absorptive capacity of the intestine. Cholera presents with vomiting, abdominal cramps, leg cramps, diarrhoea, dehydration and death. Diarrhoea in cholera can be very severe and watery. *Vibrio cholerae* serogroup O1, biotype El Tor and serogroup O139 are responsible for causing cholera with severe symptoms and result in high mortality rates (Sack *et al.* 2003; Centre for Disease Control, 2008b).

2.6.1.3 Salmonella typhi

Salmonella typhi is a foodborne bacterial pathogen which causes typhoid fever. People infected with typhoid fever carry the bacteria in the bloodstream and intestinal tract. In some instances, individuals may recover from typhoid fever but may still carry the bacteria. Carriers and infected people shed *Salmonella typhi* in their stool. Typhoid fever is a potentially fatal disease that results in extreme fever, abdominal cramps, flat red spots, headaches and appetite suppression (Centre for Disease Control, 2005).

2.6.1.4 Shigella species

Shigella species causes shigellosis. Shigellosis is an infectious disease also known as bacillary dysentery. Four *Shigella* species cause shigellosis: *Shigella dysenteriae*, *Shigella*. *flexneri*, *Shigella boydii*, *and Shigella sonnei*. Most *Shigella* infections occur via the faecal-oral route. People are infected with foodborne *Shigella* by consuming water or food contaminated by faeces from infected people, eating vegetables grown in fertilizers containing sewage, consuming food contaminated by flies that were bred in *Shigella*-infected faeces and by drinking and swimming in contaminated water. Manifestations of shigellosis are tiredness, fever, abdominal cramps, nausea, vomiting and watery or bloody diarrhoea.

Shigella flexneri is similar to EHEC in that it produces shiga toxin that result in haemolytic uraemic syndrome (Sack, 1997; Centre for Disease Control, 2008c).

2.6.1.5 Giardia lamblia

Giardia lamblia is the causative agent of Giardiasis. Giardiasis is a diarrhoeal illness caused by the parasitic protozoan *Giardia lamblia*. This parasite dwells in the intestine of humans and animals. It is passed in the stool of infected individuals. It is transmitted via the faecaloral route. Upon ingestion the cysts pass into the intestine were they develop and multiply. Eggs are released in the stool of the infected person. These eggs have an outer covering and can remain in the environment for long periods of time. They can also be spread easily in areas were proper hygiene is not practiced. Symptoms of giardiasis are diarrhoea, flatulence, greasy stool that tends to float, stomach cramps and nausea (Centre for Disease Control, 2008d).

2.6.1.6 Cryptosporidium species

Cryptosporidium parvum is the causative agent of cryptosporidiosis which can be described as a diarrhoeal disease. Recently it has been shown that *Cryptosporidium intestinalis* also causes cryptosporidosis. The oocysts of this parasite have a hard outer covering thus allowing them to survive outside the body for a long time. This outer shell also makes the parasitic oocyst resistant to chlorine disinfection. *Cryptosporidium* can be transmitted through the oral-faecal route. The most common method of transfer is via the consumption of contaminated drinking water. The oocysts travels to the small intestine were they develop and multiply and new oocysts are then released in the stool. Manifestations of this infection are nausea, watery diarrhoea, malaise, fever, weight loss and abdominal cramps (Centre for Disease Control, 2008e).

2.6.1.7 Ascaris lumbricoides

Ascaris lumbricoides (round worm) is a parasitic worm that causes ascariasis. Ascariasis is an infection of the small intestine caused by the roundworm. Roundworm is most common in developing countries where drinking water supplies are limited and of poor quality and sanitation, health and hygiene practices are inadequate. The worm is often transmitted via the faecal-oral route, where hand-washing after defaecation is not a common practice and

contamination of soil is common due to the practice of open-defaecation. Drinking water is also easily contaminated when dipping utensils are used to remove water from storage containers. *Ascaris lumbricoides* is a human intestinal parasite which enters the small intestine through ingestion of food or water containing *Ascaris* eggs, from contaminated crops, soil or faecal waste. The eggs form larva in the intestine which are released into circulation and travel to the lung. They leave the lung after three weeks of molting. Thereafter they are coughed up, swallowed and travel via the oesaphagus to the stomach and intestine. There they develop into adult male and female worms. Fertilization occurs and the female can produce approximately 200 000 eggs per day for a year. The eggs are passed out in stool and become infectious in soil in two weeks. Eggs can remain in the soil for three years. In severe cases of ascariasis, intestinal blockages may occur resulting in severe abdominal pain. Fever, wheezing and difficulty in breathing may also occur (Murray *et al.* 2005; Centre for Disease Control, 2008f).

2.7 Relevance to the present study

It is therefore clear that contamination of drinking water with a range of microorganisms can result in ill health and even death. The importance of a study, such as the current study, wherein, the microbial quality of drinking water and its associated physico-chemical properties are analysed and correlated to epidemiological data, serves as a tool to determine if the health outcomes tested (such as diarrhoea and vomiting) are related to drinking water quality solely or if there are any other confounding factors.

In several instances infections caused by the above-mentioned vectors results in vomiting and diarrhoea. It is for this reason that in the epidemiological study, results which are referenced as part of the present study, diarrhoea and vomiting have been used as measures of health with regard to microbial quality of drinking water, sanitation, health and hygiene practices.

The current microbial water quality study considered the microbial and physico-chemical quality of drinking water. It also made use of information from a study by Lutchminarayan, 2007 on health outcomes, socio-economic status and hygiene and sanitation practices of individuals dwelling in the same households from which drinking water for microbial and physico-chemical analysis was sampled. This information was collated and used to understand if any links existed between the quality of drinking water supplied by eThekwini

Municipality, incidences of health outcomes and hygiene and sanitation practices of individuals dwelling in low-income areas in Durban, South Africa. The current study may also provide insights into the issue of quantity of water versus quality of water and hygiene practices. The microbial and physico-chemical quality of drinking water supplied by eThekwini Municipality was also investigated in order to provide a guide of the current water quality status and to allow for improvements in point-of-use water quality if indicated.

CHAPTER 3: METHODOLOGY

3.1 Background to study sites and water systems analysed

The current study made use of three peri-urban areas as study sites, namely Cato Manor Sawpitts and Mtamuntengayo. The Cato Manor peri-urban settlement is situated approximately 5 km from the Durban city centre. Sawpitts and Mtamuntengayo are approximately 35 km from the Durban city centre. These two latter areas are found approximately 7 km from the Durban International Airport. (Figure 8).



*(Source: www.savenues.com).

Figure 8 Map of Durban illustrating the relative locations of Cato Manor, Mtamuntengayo and Sawpitts.

Each of these areas is located in Durban, South Africa and receives drinking water via semipressure and low-pressure water delivery systems (Chapter 2, Figure 1). Below are the areas and the corresponding water delivery systems from which water was analysed (Table 3).

- In Cato Manor drinking water was analysed from on-site ground tanks (ground tanks).
- In Sawpitts drinking water was analysed from ground tanks and community tankers.
- In Mtamuntengayo drinking water was analysed from community tankers (Table 3).

Table 3 Types of water delivery systems sampled in this study and the corresponding area from which samples were taken.

	Cato Manor	Sawpitts	Mtamuntengayo
Water delivery	-Ground tanks	-Ground tanks	-Community tankers
system used	-Standpipes		

This study compared the microbial and physico-chemical properties of drinking water at the source and its corresponding point-of-use (Figure 9). The source of water in this instance was defined as the origin or starting point from which water supplied to a specific water delivery system originated or stemmed from. It was a sample of water taken from that point in the distribution system that represented treated municipal water (Chapter 1). The point-of-use was defined as the point at which consumers use water from (Chapter 1).

Water from taps at eThekwini Municipality laboratories was taken as representative of source water samples, whilst water from ground tanks in the yard of individual households was sampled as representative of point-of-use water (Figure 9). Taps at eThekwini Municipality were taken as representative of source water since ground tanks were supplied with water from the same municipal distribution system. Since ground tanks serve as both a storage vessel and a dispensing vessel, it was regarded as the point-of-use. Community tankers and standpipes differ in this aspect. This is because these water delivery systems are communal and require that water be collected from the standpipe or tanker, transported to the household and stored until used (Figure 10). Therefore the point-of-use is the storage vessel from which water is used. The source/supply that represents the source of treated municipal water for this storage vessel is the community tankers and standpipes (Figure 9 and Figure 10).

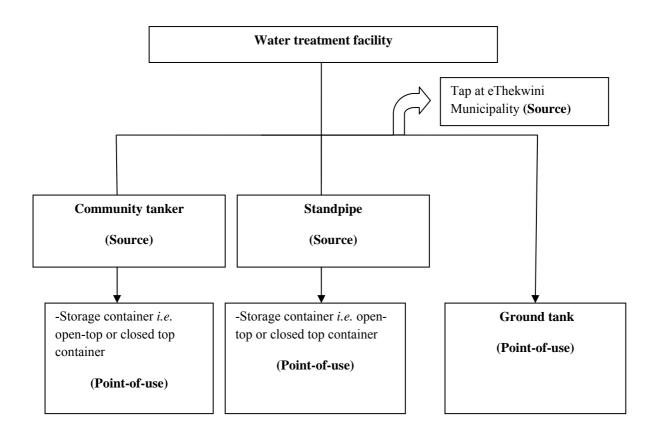


Figure 9 Schematic diagram representing the point-of-use and corresponding source for mobile community tankers, communal standpipes and on-site ground tanks.

Figure 10 explains the flow of water as occurs when water is obtained from communal water sources. Standpipes and community tankers received water from a municipal water treatment plant. This water was then collected from these sources, by users, in either open-top or closed-top storage containers. For purposes of this study, closed-top containers were defined as containers which had a lid that served as a seal from the environment when water was not being used, whilst open-top containers, conversely, were defined as containers which were devoid of a lid and those which were left open during the entire period in which water was stored and used (Chapter 1). The containers filled with water were transported to households after collection. Here, it is stored and used as required (Figure 10)

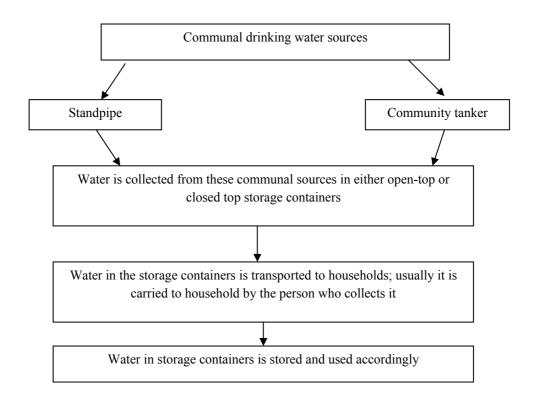


Figure 10 Schematic diagram representing the pathway of water supplied by communal water delivery systems such as standpipes and community tankers.

For the purposes of this study, comparisons were as follows:

- The microbial and physico-chemical quality of drinking water from ground tanks in Sawpitts and ground tanks in Cato Manor were compared. Drinking water from Cato Manor was sampled in 2004 by Maraj (2005). This sampling was done during winter months. Sampling of drinking water from Sawpitts was done by the author and a research team in 2006 during the summer months. Sampling data from both areas were used to compare the microbial and physico-chemical quality of drinking water in these two areas during different seasons. However, a shortfall in the seasonal study was the difference in locations; this is acknowledged and hence this piece of work is presented in the Appendix A.
- The microbial and physico-chemical quality of drinking water from communal water sources *viz*, standpipes in Cato Manor and community tankers in Mtamuntengayo were compared. The reason that standpipes were compared to community tankers rather than to corresponding standpipes was because the same households in Cato Manor could not be sampled due to high mobility of household occupants. A large database of information on Mtamuntengayo was already established at the time of this study (Lutchminarayan, 2007). This allowed for better selection of households with

similar socio-economic and demographic properties than those selected in the 2004 study by Maraj *et al.* (2005), in which water from standpipes and storage vessels in Cato Manor were sampled and analysed. Therefore it was decided that Lutchminarayan's (2007) study site would be used and households were selected accordingly. The use of this study site also allowed for addition of information on water quality in these areas, into the existing database. The use of different areas was therefore not viewed as a shortfall of this comparison, since both water sources were communal and household selections in each area was based on exactly the same criteria.

- The microbial and physico-chemical properties of drinking water stored in open-top and closed-top containers in Cato Manor and Mtamuntengayo were compared. These vessels were supplied by either standpipes or community tankers.
- The microbial quality of drinking water at the point-of-use in households comprising adults only (*i.e.* occupants which are > 18 years of age), households including children aged 0-5 years and households including children aged >5-18 years was compared. The purpose of this comparison was to establish if household demographics impacted on the microbial quality of drinking water at the point-of-use.
- The microbial quality of drinking water, health outcomes (diarrhoea and vomiting) and socio-economic status of individuals supplied drinking water by each of the above-mentioned water delivery systems was compared in order to establish if there is a link between water quality, health outcomes, socio-economic status and water-use-behaviour. Data on health outcomes and socio-economic status was obtained from the ECOSAN database.

3.2 Ethics approval

The current study is a water quality study. It aimed to supplement the existing information of an epidemiological study by Lutchminarayan (2007), but it is not itself an epidemiological study. Therefore the ethical approval of the epidemiological study was used for the current study with permission of the owners. The same sites which were used in the epidemiological study were used for the current study. To gain entrance into these areas and hence sample household drinking water, the author and members of the project team went through the eThekwini Municipality community liaison structures. The study was conducted with the permission of the municipal manager (Appendix B) and the community was approached through the eThekwini Water and Sanitation (EWS) community liaison staff and the ward councillor, using members of the communities as facilitators and interpreters. The purpose of the study was explained to each householder and verbal informed consent was obtained before samples were collected at each household. Permission to conduct this study was granted by the eThekwini Municipality manager, this letter is included in Appendix B.

3.3 Household selection

Households were selected based on the type of water delivery system (ground tanks, standpipes, community tankers), household demographics (households including children aged 0-5 years, households including children aged >5-18 years and households comprising adults only) and type of storage containers used to store drinking water in (open-top or closed-top storage containers).

3.3.1 Household selection in Cato Manor

All work in Cato Manor was conducted by Maraj *et al.*, 2005 as part of an Honours dissertation at the University of KwaZulu-Natal. For this study, questionnaires were developed in conjunction with social workers at the University to determine the age of people dwelling in households and to determine the type of water delivery system used. The questionnaire comprised of two sections, a general information section and a section referred to as a diarrhoea diary. The household questionnaire aimed at obtaining general information regarding socio-economic status, type of water delivery system used, type of sanitation practised, type of water storage vessels used, hygiene practices, poverty indices and demographic distributions. The diarrhoea diary was used to record incidences of diarrhoea and duration of diarrhoea in members dwelling in each household. Based on the answers to these questions households were selected as explained above in 3.3.

Maraj (2005) obtained informed consent from every household that agreed to participate in this study. All microbial and physico-chemical analysis was performed by Maraj in conjunction with the staff of eThekwini Municipality Laboratories.

3.3.2 Household selection in Sawpitts and Mtamuntengayo

Household selection in Sawpitts and Mtamuntengayo included the use of the ECOSAN database containing information gathered for a completed Master of Science degree in Epidemiology by Lutchminarayan in 2007.

Lutchminarayan and colleagues performed comparative evaluations on the health impacts of ecological sanitation interventions, water services and hygiene education programmes, individually and in combinations, in eThekwini Municipality, Durban, South Africa. The research aimed to determine whether providing sanitation, safe water and health and hygiene interventions in peri-urban households improved health outcomes.

The database for this study included information from 1350 households. The study design was an observational analytic prospective cohort study between intervention and control groups (Lutchminarayan, 2007). Intervention groups comprised households in which members were given health and hygiene education and which were provided with UD toilets and ground tanks by eThekwini Municipality. Household members were also educated on health and hygiene practices and how to implement proper use and care of UD toilets and ground tanks by eThekwini Municipality staff. The control group comprised households that did not have UD toilets and ground tanks. These households were not educated on health and hygiene practices. They often made use of communal water supplies and stored their drinking water in open-top or closed-top containers. These households used either VIP toilets or practice open defaecation (the bush) for sanitation purposes.

All data in the epidemiological study were collected by means of questionnaire surveys (Lutchminarayan, 2007). Three surveys were used:

 Household questionnaire – The aim of this survey was to gather general information on the socio-economic status of household members, the type of water and sanitation facilities, health and hygiene practices and education levels in households. Details on occupants of the household, number of members in households, age distributions, and possession checklists (to indicate socio-economic status and levels of education) were also recorded.

- Health outcomes questionnaire The aim of this survey was to determine the rates of diarrhoea, vomiting, skin sores and worms and the duration of such health outcomes in household members participating in the study.
- Observational questionnaire The aim of this survey was to determine if information given in the household questionnaire corresponded to what was observed in the surroundings. For example, if a respondent indicated that they washed their hands with soap after using a toilet, the observational team member would check if soap was present in the wash area. This served as a method of confirmation of the preceding two questions.

All questionnaire data were captured using EpiData version 2.2. Data were combined and stored on a database program, for statistical analysis.

For the current study, a sample population of 72 households was selected as a subset from the database consisting of 1350 households used in the epidemiological study by Lutchminarayan (2007).

Households were selected according to type of water delivery system used, age distribution within households, and type of storage containers used to store drinking water.

The microbial study was blinded¹⁴, in terms of epidemiological study outcomes. The same methods used in the microbial study by Maraj (2005) were implemented in the current microbial study.

Nine groups were compared to determine if the quality of domestic water supplies was related to health outcomes and socio-economic status (Table 4). Sampling of drinking water and its subsequent microbial and physico-chemical analysis was performed by the author in conjunction with the staff of eThekwini Municipality Laboratories.

¹⁴ The blind method is a scientific method used to prevent results from being influenced by observer bias or the placebo effect as a result of conscious or unconscious bias (Freund *et al.* 1988; Bacchieri and Cioppa, 2007; Kaptchuk, 2000).

Table 4 Criteria used for sample selection based on demographic composition of household members in Sawpitts and Mtamuntengayo which store drinking water in ground tanks or in open-top or closed-top storage containers respectively. The water source and corresponding point-of-use is also given.

Household demographic group	Area	Water source	Point-of-use samples
Households with children aged 0-5 years Households with children aged >5-18 years Households comprising adults only (>18 years)	Sawpitts	Water samples from eThekwini Municipality taps were taken as representative of source water since ground tanks are supplied with water from the municipal distribution system.	Water stored in ground tanks for 24 hours
Households with children aged 0-5 years Households with children aged >5-18 years Households comprising adults only (>18 years)	Mtamuntengayo	Community tankers	Water stored in open-top containers
Households with children aged 0-5 years Households with children aged >5-18 years Households comprising adults only (>18 years)	Mtamuntengayo	Community tankers	Water stored in closed- top containers

3.4 Sampling

Drinking water was sampled once a day for a total of 10 days at each site. Sampling was done from Monday to Thursday each week, since eThekwini Municipal Laboratories closed on Fridays and samples requiring overnight incubation could not be analysed on weekends. The source and corresponding point-of-use water from ground tank, standpipe and community tanker-supplied households were sampled (Figure 9). Samples were analysed by eThekwini Municipality Laboratories to allow more samples to be analysed than could be handled by the experimenter alone (greater statistical strength of data and the lab followed full quality control procedures including blinding of samples thereby removing any chance of bias).

3.4.1 Sampling from ground tanks

Taps on ground tanks were wiped down with 90% ethanol in order to destroy any bacteria present on taps which could enter drinking water samples and hence provide biased results. Taps were then left to run (flushed) for 1-2 minutes. The purpose of flushing taps was to remove bacteria present in water that was stagnant in the tap region. It is important to practice

flushing of taps when collecting drinking water samples for microbial analysis in order to make sure that microbial constituents detected in water samples are representative only of actual water found in the tank rather than the water left to stagnate at the tap region. Water was collected into a 200 ml sterilized Schott bottle. All collection bottles were sealed in foil and sterilized in an autoclave before collection of samples. This served to ensure that all bottles were clean and hence to limit information bias which could occur if sample bottles were contaminated. The foil layer of each bottle was removed by the experimenter, who wore a clean pair of sterile gloves at each sample point, only when samples were ready to be collected. Once the foil layer was removed the neck of the bottle was wiped down with 90% ethanol to remove any contaminants present on the bottle and hence to limit bias with regard to microbial results. A sample was collected from the running tap, the cap of the bottle was closed and the bottle neck was wiped down with 90% ethanol before the sample bottle was stored on ice. It is important to note that at no point during the collection of samples did the gloves of the experimenter touch the tap mouth of the ground tank or any other surrounding region. Following the collection of each sample, gloves were discarded. A new pair of gloves was used in the collection of each sample. All 200 ml Schott bottles contained 3-4 drops of sodium thiosulphate. Sodium thiosulphate served to neutralise residual chlorine in samples to be used for microbial analysis. Removal of residual chlorine in these samples was important so that microbes would not be destroyed by the residual chlorine during transportation to labs and hence a true representation of microbial content in the drinking water sample could be attained. A second sample of drinking water from ground tanks was collected in 100 ml McCartney bottles, in the same manner as explained above. These samples were used to measure the turbidity, conductivity and pH of samples at the eThekwini Municipality laboratory. All samples were placed on ice and transported to the eThekwini Municipality laboratory for analysis. The purpose of ice was to maintain the water temperature in order to prevent changes in pH and, in McCartney bottles, to prevent changes in chlorine levels since chlorine volatilizes at high temperature. A change in temperature would also affect the microbial content, the turbidity and conductivity of samples (Chapter 2).

3.4.2 Sampling from standpipes

The methodology used to sample water from standpipes was the same as methodology used to sample water from ground tanks. Firstly, taps on standpipes were wiped down with 90% ethanol. Thereafter taps were flushed for 1-2 minutes. Water samples were then collected into

200 ml sterilized Schott bottles and 100 ml McCartney bottles for the relevant analysis, as described in 3.4.1.

3.4.3 Sampling from community tankers

Water from community tankers is dispensed into water collection vessels via a tap found on the tanker. Therefore the same sampling methodology was used to collect water samples from community tankers as described in 3.4.1 and 3.4.2.

3.4.4 Sampling of water from storage vessels

After collection, water was stored in either open-top or closed-top storage vessels. Water samples from open-top containers were collected by pouring out drinking water directly from the storage container into the sample collection bottles. Note that, sterile gloves were worn by the sampler at all times. Only sterilized sample bottles were used. Sample bottles were wiped down with ethanol before and after samples were collected. Contact of the neck of the collection bottle with the rim of the container and the hands of the sampler was always avoided. The neck of the sterilized sample bottles was wiped down with 90% ethanol before and after collection. A similar procedure was followed for sampling of water from closed-top containers, the only difference being that the lid of the container was removed, using sterile gloves before water samples were taken.

3.5 Analysis of water samples

All samples were analysed in batches. Technicians were blinded with regard to the origin of samples. Each sample was labelled with a code which could only be tracked by the experimenter.

All water samples were analysed for the following microbes:

- *E. coli*,
- total coliforms and
- heterotrophic organisms.

The following physico-chemical properties of water samples were also measured:

- pH
- water temperature
- turbidity
- conductivity
- total chlorine and
- residual chlorine

3.5.1 Methodology used to enumerate E. coli and total coliforms

Microbial analysis was conducted by the experimenter and eThekwini Municipality laboratory staff. The experimenter assisted in inoculation of agar whilst the eThekwini Municipality laboratory staff prepared agar and conducted colony counts following incubation. Membrane filtration, combined with incubation on specific agar, was used to detect E. coli and total coliform bacteria (eThekwini Municipality Test Method Number MM002, 2004). The filtration apparatus consisted of three plastic funnels which fitted onto three filtration stands placed upon a filter manifold. Funnels and filtration stands were sterilized by being boiled in water. The filtration rack was in turn connected via plastic piping to a vacuum pump. Filtrate was collected in a sterile glass bottle. One hundred millilitres of each sample was filtered through a sterile membrane filter with a 0.45µm pore size. Membrane filters were dipped in boiling water prior to placement on the filter units. After each sample was filtered through the apparatus, the apparatus was disassembled, and the funnels sterilized as described. Membrane filters were placed grid upwards on Chromocult® coliform agar (Merck) in a 60 mm plastic petri dish, and incubated in an inverted position at 37°C for 12 hours. This was done in triplicate for all samples. Following incubation, plates were removed and the colonies on the surface of the filter paper were counted. Blue colonies represented E. coli colonies, and pink colonies represented total coliform colonies.

The chromogenic substances in Chromocult coliform agar are the cause for the colour differentiation between bacterial colonies formed by *E. coli* and total coliforms. Coliform bacteria contain β-D-glucuronidase, which cleaves the Salmon-GAL substrate contained within the agar. This results in the formation of a salmon to red colour when observing the coliform colonies. In the case of *E. coli*, the indole reaction, improved by the addition of tryptophane in the medium, is used as a method of confirmation of *E. coli* presence. β-D-

glucuronidase is identified within *E. coli* through use of X-glucuronide. This, together with Salmon-GAL is cleaved by *E. coli*, resulting in the formation of a dark blue to violet colour (eThekwini Municipality, 2004; Satory and Howard, 2008).

Colony numbers were recorded and reported as colony forming units (CFU) per 100 millilitres of sample filtered.

3.5.2 Methodology used to enumerate heterotrophic organisms

The standard pour plate method was used to enumerate heterotrophic organisms (eThekwini Municipality Test Method Number MM007, 2004). Heterotrophic plate count agar (Merck) was prepared prior to use. Agar was sterilised by autoclaving at 120°C for 15 minutes. Agar was melted using a water bath, and removed from heat once it could be held comfortably. Liquid agar was poured into a 90 mm sterile Petri dish, which was inoculated with one millilitre of water sample. The petri dish containing the liquid agar and water sample was swirled to facilitate mixing. Thereafter the agar was allowed to set. Plates were placed in an incubator at a temperature of 37°C for 24 hours. Following incubation, plates were removed from the incubator, and the number of colonies enumerated. These numbers were reported as cfu/ml.

3.5.3 Measurement of physico-chemical parameters

3.5.3.1 Measurement of pH

The pH of all water samples, collected in McCartney bottles, was measured at the eThekwini Municipality laboratory, by the experimenter, using a Metrohm 691 pH meter [®].

3.5.3.2 Measurement of water temperature

Water temperature of all samples was measured on site, by the experimenter, using a thermometer.

3.5.3.3 Measurement of conductivity

Conductivity was measured at the eThekwini Municipality laboratory, by the experimenter, using a Mettler Toledo MC226 conductivity meter [®].

3.5.3.4 Measurement of turbidity

Turbidity was measured at the eThekwini Municipality laboratory, by the experimenter, using a Hach 2100N Turbidimetre [®].

3.5.3.5 Measurement of free and total chlorine

Free and total chlorine levels were measured on site, by the experimenter. This was done by combining either total chlorine or free chlorine reagent (Hanna Instruments) with 10 ml of water sample, using the low and medium range free and total chlorine meter (Hanna Instruments).

3.6 Statistical analysis

All data was analysed using the statistical package SPSS version 15. Ms Tonya Esterhuizen, a qualified statistician from the School of Health Sciences, was consulted for guidance in the selection of appropriate statistical procedures and tests. All statistical analyses were performed by the author in consultation with a biostatistics postgraduate student, Ms Jaclyn Wright.

3.6.1 Statistical analysis of microbial parameters

Three microbial parameters (*E. coli*, total coliforms and heterotrophic organisms) were measured in drinking water.

All data was tested for normality of distribution using the Kolmogrov-Smirnov test. Microbial data was found to be non-normally distributed. Transforming of microbial data was unsuccessful in obtaining normal distribution due to the extreme variability in microbial counts. Microbial data was therefore ranked and subject to non-parametric analysis of variance using the Kruskall-Wallis test with the Tukey post-hoc testing in order to determine:

- The relationship between the microbial quality of drinking water at the source and corresponding point-of-use.
- The relationship between the microbial quality of drinking water from open-top and closed-top storage containers and
- The relationship of the microbial quality of drinking water sampled in summer compared to the quality of drinking water sampled in the winter (Appendix A).

Because distribution of microbial data was typically skewed, one was added to all data in order to be presented in Log form on graphs. Without logging data, graphs would not have been able to be reported as log graphs and would hence be difficult to read since there were numerous low values but a few very high data values which would have masked other values on a non-log scaled graph.

Results include the arithmetic mean, geometric mean and 95th percentile for all microbial data. The reason that each of these statistics was included is explained below.

South African Water quality guidelines (DWAF, 2005) specify frequency of sampling. For communities with a population of less than 2 500, the sampling frequency of once a month is recommended. The communities sampled in this study were in all cases smaller than 2 500 people. Since more than twenty samples were taken per day over a two week period, for each sample group tested, the probability of finding a microbially non-compliant sample was much higher. For this reason, means (arithmetic means and geometric means) were compared to South African Water Quality guidelines and WHO water quality guideline values (DWAF, 1996; DWAF, 2001; WHO, 2005). The value most commonly used for comparison to WHO guidelines is the arithmetic mean (Langmark pers.com), although evaluation of WHO guideline limits is also linked to frequency of sampling.

The arithmetic and geometric means were calculated without adding 1 to the data, as was done during log transformation of data to draw graphs. The geometric mean differs from the arithmetic mean in that whilst the calculation of the arithmetic mean involves the addition of a list of numbers and division of this sum by the total count of numbers in the data set (n), the geometric mean is calculated by multiplying the set of numbers and taking the *n*th root of the

product. The reason for reporting on both the geometric and arithmetic means in the results chapter (Chapter 4) is that whilst the WHO guideline is often compared to arithmetic mean microbial counts, the geometric mean microbial counts offers better indication of central tendency (Eaton *et al.* 1995; Jagals *et al.* 1999). Hence to develop a holistic view on the risk posed by drinking water, to cause infection in users, the arithmetic and geometric mean microbial counts were compared to guideline values. To further supplement the measure of risk, the 95th percentile of each microbial count was calculated and plotted on graphs, in order to determine and hence report on the upper limit of risk (Kay *et al.* 2003). The 95th percentile values were also compared to guideline values in order to avoid gross underestimation of the risk posed by drinking water to humans in terms of microbial infections, as would be expected from mean values.

3.6.2 Statistical analysis of physico-chemical parameters

All physico-chemical data was tested for normality of distribution using the Kolmogrov-Smirnov test. Data was normally distributed and was tested for significance of differences by one-way-Analysis Of Variance (ANOVA) coupled with least significant difference (LSD) post-hoc tests. Average values for all physico-chemical data are reported.

3.6.3 Statistical analysis to determine the relationship between microbial water quality, health outcomes and socio-economic factors.

All microbial data for each household sampled in Sawpitts and Mtamuntengayo was averaged and merged with the epidemiological data set in order to allow for statistical analysis between microbial and epidemiological data sets. Analysis was performed to define relationships, if any, between:

- Microbial water quality and health outcomes (diarrhoea and vomiting).
- Microbial water quality and social factors.
- Health outcomes and social factors.
- Place (*i.e.* Sawpitts and Mtamuntengayo) versus age of household members and socio-economic factors.

3.6.3.1 Assessing the statistical relationships between microbial water quality and health outcomes.

Four health outcomes (diarrhoea, vomiting, skin sores and worm infestations) were assessed by Lutchminarayan (2007). Only two of these health outcomes (diarrhoea and vomiting) were used in this study. These health outcomes were assessed by recording the number of occurrences of diarrhoea and vomiting in each household. The rate of diarrhoea and vomiting was recorded based on the observation and subsequent reported incidences of these health outcomes by the mother of the household or the main interviewee of a household.

The incidence of each health outcome was calculated as follows and was used in all statistical tests:

Incidence of health outcome = <u>Number of people in each household ×follow up time ×1000</u>

To investigate relationships between microbial water quality, diarrhoea and vomiting, 3 microbial parameters (*E. coli*, total coliforms and heterotrophic organisms) were used.

To test for significant differences in water quality and in health outcomes among the various points-of-use, water quality data and health outcomes data, were ranked and the Kruskall-Wallis test was conducted on the ranked data, with post hoc testing using the Tukey HSD test. Similarly, to test for significant differences in water quality and in health outcomes among the different age groups, the Kruskall-Wallis test was performed. In addition, to take into account both the effect of age group and of point-of-use water quality simultaneously on health outcomes and the effect of age group on microbial water quality, the Kruskall-Wallis test was performed, using nine groups (Table 3). In order to test for direct relationships between water quality and health, Spearman's rank correlation tests were performed.

Table 5 Categorical division of groups based on age distribution in households and type of storage container used to store drinking water.

Category number	Age distribution	Type of storage container		
1	0-5 years	Open-top		
2	>5–18 years	Open-top		
3	Adults only (>18 yrs)	Open-top		
4	0-5 years	Closed-top		
5	>5 – 18 years	Closed-top		
6	Adults only (>18 yrs)	Closed-top		
7	0-5 years	Ground tanks		
8	>5 – 18 years	Ground tanks		
9	Adults only (>18 yrs)	Ground tanks		

3.6.3.2 Assessing the statistical relationship between socio-economic factors and health outcomes and between socio-economic factors and microbial water quality

To determine significant relationships between socio-economic factors and health, and between socio-economic factors and microbial water quality, Spearman's rank correlation, Mann-Whitney and Kruskall-Wallis statistical tests were used as follows. In cases where the socio-economic factors were quantitative, Spearman's rank correlation was used. Where socio-economic factors were categorical (with 2 categories), Mann-Whitney was used. For more than two categorical groups, the Kruskall-Wallis test was used.

In addition, all microbial parameters (heterotrophic organisms, total coliforms and *E. coli*) were categorised into four quartiles to determine the link between hand-washing and water quality.

3.6.3.3 Assessing the statistical relationships between place/age distribution and socio-economic factors

To test for significant relationships between place or age group, and socio-economic factors, the Mann-Whitney test, the Kruskall-Wallis test and cross tabulations were used as follows. For the relationship between place (Mtamuntengayo and Sawpitts) and quantitative socio-economic factors, the Mann-Whitney test was used. Relationships between age group

(households including children aged 0-5 years, >5-18 years and households with adults only), and quantitative socio-economic factors were assessed using the Kruskall-Wallis test. For relationships between place or age group and categorical socio-economic factors, cross tabulations were used. For two by two tables (*i.e.* tables comprising two factors), Fischer's exact statistic was used to determine significant differences, if any, between place, age distribution and socio-economic factors. For larger tables (*i.e.* tables comprising more than two factors), the Chi square statistic was used and assumptions were checked (no less than 20% of expected values <5).

CHAPTER 4: RESULTS

4.1. Point-of-use versus source water quality from ground tanks, standpipes and community tankers

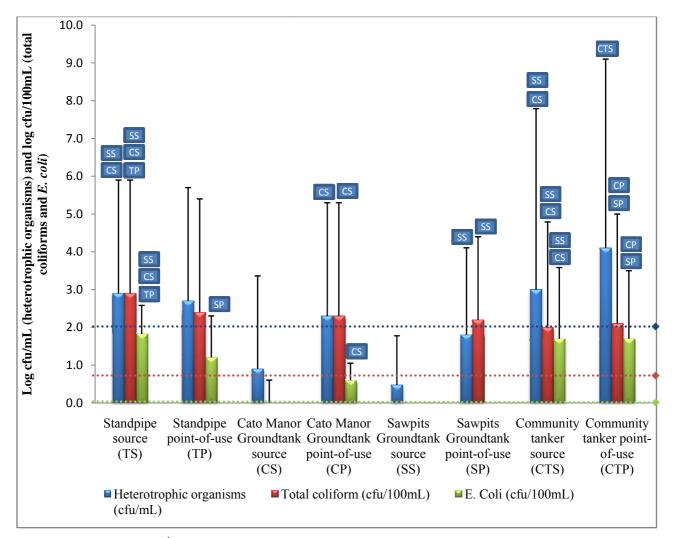
The microbial and physico-chemical quality of drinking water from standpipes, ground tanks and community tankers, at source and at the corresponding point-of-use, supplied to households in Cato Manor and Sawpitts was investigated. The results are presented below.

The DWAF and WHO guidelines for acceptable counts of heterotrophic organisms, total coliforms and *E. coli* and physico-chemical parameters in drinking water are presented in Tables 1 and 2 respectively (Chapter 2).

Measured levels of microbial parameters were compared to guideline values to assess whether water at the point-of-use was fit for consumption, based on risk of microbial infection posed to users as a result of microbe counts in the water.

Water supplied by community water tankers in Mtamuntengayo, had mean *E. coli*, total coliform and heterotrophic organism counts above the acceptable levels for safe drinking water at both source and point-of use (Table 1) (DWAF, 1996; WHO, 2005). Log-transformed counts of heterotrophic organisms, total coliforms and *E. coli* in ground tank households, from Sawpitts and Cato Manor, were significantly higher at the point-of-use compared to the corresponding counts at the source (p<0.001 for all microorganisms) (Figure 11). This indicated that untransformed data are also likely to differ significantly. Standpipe source water had higher microbial counts than the corresponding point-of-use water, however only total coliform and *E. coli* counts were significantly higher (p=0.013 and p=0.010). Community tanker source water had slightly lower, but not significantly so, total coliform counts than the corresponding point-of-use water than in source water from community tankers (p=0.013 for heterotrophic organisms) (Figure 11).

DWAF and WHO maximum levels of heterotrophic organisms, total coliforms and *E. coli*, are given by blue, red and green dotted lines in Figure 11 respectively. These markers show that all water except ground tank source water is unsafe for human consumption and poses a risk of microbial infection to users. If the upper limit of the 95th percentile is taken as the maximum level of risk then maximum risk for all sample groups presented by the 95th percentile is higher than that for the mean risk (Figure 11). Blue boxes indicate where microbial counts are significantly higher in comparison to microbial counts in other source and point-of-use water.



*Error bars represent the 95th percentile. Horizontal lines represent maximum values of heterotrophic organisms (blue), total coliforms (red) and *E. coli* (green) counts allowed in drinking water rendered safe for human consumption. Blue boxes represent significantly higher microbial counts than the sampling point referred to in the box.

Figure 11 Log arithmetic mean microbial counts of source and corresponding point-of-use drinking water from standpipe, ground tank, and community tanker-supplied households.

The 95th percentiles, arithmetic and geometric mean counts of heterotrophic organisms, total coliforms and *E. coli* were compared to South African Water Quality Guidelines and WHO guidelines (Table 1) (reasons for using these statistical measures is discussed in Chapter 3). South African Water Quality Guidelines classify water as posing a negligible risk (NR), potential risk (PR) or substantial risk (SR) of causing infection in users depending on microbial counts of indicator organisms detected for the various source and point-of-use waters (Table 6). It should be noted that a study based purely on counts of indicator microorganisms, such as the present study, strictly only indicates the *risk of contamination*. The necessary factors to determine the *risk of infection* (e.g. presence of pathogens, exposure of consumers to pathogens, infectivity and susceptibility) were not investigated. However, the DWAF guidelines, used to evaluate water quality here, link levels of microorganisms to a probable risk of infection. Comments made on risk of infection or illnesses are based purely on this guideline recommendation.

The upper limit of the 95th percentile for heterotrophic organisms, total coliforms and *E. coli* placed drinking water from standpipe source, standpipe point-of-use, Cato Manor ground tank point-of-use, community tanker source and corresponding point-of-use water in the substantial risk of causing microbial infection to user's category (Figure 11). The upper limit of risk of the 95th percentile for source water from ground tanks in Sawpitts shows that this is the only water which is safe for human consumption with regard to all three microbial parameters (Figure 11).

Arithmetic and geometric mean microbial counts for standpipe source and point-of-use water posed a potential risk of microbial infection to users in one parameter (HPC) and a substantial risk of microbial infection to users in two parameters (total coliforms and *E. coli*) (Table 6). Ground tank source water in Cato Manor and Sawpitts had overall mean microbial counts (*i.e.* both geometric and arithmetic mean counts for HPC, total coliforms and *E. coli*) that posed a negligible risk of microbial infection to users (Table 6). Ground tank-supplied point-of-use drinking water in Cato Manor posed a potential risk of microbial infection to users in two microbial parameters (HPC and *E. coli*)) and a substantial risk of microbial infection to users in two microbial parameters (total coliforms), according to arithmetic and geometric mean microbial counts of total coliforms and *E. coli* posed a substantial risk of microbial infection and a potential risk of total coliforms and *E. coli* posed a substantial risk of microbial infection and geometric mean counts of total coliforms and *E. coli* posed a substantial risk of microbial infection and a potential risk of total coliforms and *E. coli* posed a substantial risk of microbial infection and a potential risk of total coliforms and *E. coli* posed a substantial risk of microbial infection and a potential risk of total coliforms and *E. coli* posed a substantial risk of microbial infection and a potential risk of total coliforms and *E. coli* posed a substantial risk of microbial infection and a potential risk of infection to users respectively, whilst the HPC arithmetic mean value posed a negligible

risk of infection to users and the HPC geometric mean counts posed a potential risk of infection to users (Table 6). Thus far in comparing geometric and arithmetic means, this is the first instance in which the counts display differing results. Source and point-of-use drinking water from community tankers had arithmetic and geometric mean counts of HPC, total coliforms and *E. coli* that posed a substantial risk of causing microbial infection in users (Table 6).

In comparison to ground tank source water, source water from standpipes and community tankers showed a worse quality (Figure 11), falling into the category of substantial risk for *E. coli* and total coliform counts with regard to the geometric and arithmetic mean (Table 6). Water at the point-of-use from ground tanks displayed significant deterioration in microbial quality in comparison to the corresponding source water in all three microbial parameters (p=0.012 for heterotrophic organisms, p<0.001 for total coliforms and p<0.001 for *E. coli*). Water at the point-of-use from standpipes did not display further significant deterioration in microbial quality relative to its source water (p=0.536 for heterotrophic organisms, p=0.153 for total coliforms and p=0.10 for *E. coli*). Point-of-use water from community tankers displayed a significant deterioration in microbial quality, only in terms of heterotrophic organism counts (p=0.013) when compared to its corresponding source water quality (Table 6 and Figure 11).

Overall when comparing risk levels associated with arithmetic mean, geometric mean and 95th percentiles, all three statistical measures rendered the same risk of infection categories for

- total coliform counts in all source and point-of-use water samples;
- *E. coli* in all standpipe source and point-of-use water;
- E. coli in ground tank point-of-use water in Sawpitts;
- *E. coli* in community tanker and source water;
- Heterotrophic organisms in community tanker source and point-of-use water; and
- Heterotrophic organisms in ground tank source water in Sawpitts.

The upper limit of the 95th percentile for heterotrophic organisms differed from the arithmetic and geometric mean risk levels for standpipe and ground tank water in Cato Manor. Here,

drinking water posed a higher level of risk of microbial infection to users in comparison to the level of risk derived from either the geometric or arithmetic means.

Table 6 Arithmetic mean (AM) and geometric mean (GM) microbial counts from source and point of-use water in all sample groups and areas (heterotrophic plate count (HPC) in cfu/mL; total coliforms and E. coli in cfu/100mL.

			HPC		Total coliform		E. Coli	
Area			(cfu/ml)	Risk	(cfu/ml)	Risk	(cfu/ml)	Risk
		AM	794		794		67	
	Source	$n^{d} =$	40	PR ^b	40	SR ^c	40	SR
	(community	GM	902		900		210	
Standpipes	standpipes)	n=	38	PR	34	SR	33	SR
Standpipes	Standp-p-ts)							
		AM	501		251		16	
	Point of use	n=	480	PR	480	SR	480	SR
	(containers in	GM	640		430		95	
	household)	n=	344	PR	300	SR	54	SR
		AM	8		0		0	
	Source	n=	9	NR ^a	9	NR	9	NR
Ground	(laboratory	GM	25		0		0	
tanks	tap)	n=	5	NR	0	NR	0	NR
Cato Manor	Point-of-use		100		100			
	(water	AM	199	DD	199 267	(TD)	4	DD
	dispensed from	n=	267	PR	267	SR	267	PR
	tap on ground	GM	360	DD	290 132	CD	8 25	DD
	tank)	n=	184	PR	132	SR	25	PR
							0	
		AM	3		0		4	
	Source	n=	40	NR	40	NR	4 0	NR
Ground	(laboratory	GM	12		40 0		0	
tanks	(laboratory tap)	n=	5	NR	0	NR	0	NR
Sawpits	Point-of-use		5	1.11		1111		
	(water	AM	64		158		1	
	dispensed from	n=	280	NR	280	SR	280	PR
	tap on ground	GM	160		110		5	
	tank)	n=	131	PR	92	SR	34	PR
	Source	AM	10 000		100		50	
	(water	n=	68	SR	68	SR	68	SR
Community	dispensed from	GM	14 000		250		60	
tankers	tanker)	n=	63	SR	56	SR	43	SR
tunner 5		4.3.7	13 500				5 0	
	Doint of	AM	12 589	CD	126	CD	50	CD
	Point-of-use	n= CM	480 127 000	SR	480	SR	480	SR
	(containers in	GM	231	SR	240	SR	96	SR
aND as ali aible ai	household)	n=	231	SK	147	SK	62	SK

^aNR-negligible risk. ^bPR- potential risk. ^cSR- substantial risk). ^dn represents the number of samples used to derive mean values.

Results for physico-chemical measurements are given in Table 7 below. All standard values for physico-chemical properties are given in Table 2. The pH in all groups was within limits recommended by DWAF and WHO guidelines (Table 2, Chapter 2). Residual chlorine levels in water from ground tanks and standpipes in Cato Manor were within recommended levels of 0.2 to 1.5 mg/L (DWAF, 1996; WHO, 2005). However, residual chlorine levels in point-of-use water from community tankers and Sawpitts ground tank households were below recommended levels in DWAF and WHO guidelines (Table 7).

Total chlorine levels in all groups were below the recommended levels of 2 to 2.5 mg/L. Turbidity in all groups except source water for community tankers was within the recommended levels of 1–5 NTU. Conductivity in all sample groups was within the guideline levels of 100mS/m (DWAF, 1996; DWAF, 2005; WHO, 2005) (Table 7).

High turbidity, in this instance, was associated with low residual and total chlorine levels and lower conductivity levels. The highest turbidity was observed in communal water sources (standpipes and community tankers) and the associated point-of-use water, whether stored in open-top or closed-top containers (Table 7).

The standard deviation and number of data points used for all physico-chemical parameters is also given in Table 7. The standard deviation for pH, residual and total chlorine in all sample groups was below one. This indicated low level of dispersion of data. Turbidity in water samples from all sample groups except, standpipe and community tanker point-of-use water had a standard deviation less than one. The higher standard deviation as occurred in standpipe and community tanker households represents the high level of dispersion of this data. It implies that few households may have had very high turbidity values whilst a larger number of these households may have had low turbidity values.

The standard deviation for conductivity values for water sampled from all point-of-use and sources was above one except in ground tank point-of-use water.

Sample groups		рН	Residual chlorine (mg/L)	Total Chlorine (mg/L)	Turbidity (NTU)	Conductivity (mS/m)
	Source					
	(community standpipes)	7.88	0.37	0.51	0.74	14.74
	n ^a =	40	40	40	40	40
Standpipes	SD ^b =	0.10	0.11	0.17	0.52	2.7
	Point of use					
	(containers in household)	7.92	0.23	0.36	1.1	15.34
	<u>n</u> =	480	480	480	480	480
	SD =	0.14	0.14	0.20	2.2	2.7
	Source (laboratory tap)	8.13	0.28	0.49	0.51	25.84
	n =	9	9	9	9	9
Ground	SD =	0.14	0.17	0.19	0.64	1.8
tanks Cato Manor	Point-of-use (water dispensed from tap on ground tank)	7.99	0.38	0.57	0.66	14.91
		1				
	<u>n =</u>	267	267	267	267	267
	SD =	0.20	0.41	0.35	0.29	1.23
Ground tanks Sawpitts	Source (laboratory tap)	7.56	0.16	0.57	0.03	10.24
	n =	40	40	40	40	40
	SD =	0.30	0.08	0.09	0.9	1.27
	Point-of-use (water dispensed from tap on ground tank)	8.08	0.07	0.1	0.53	12.41
	n =	280	280	280	280	280
	SD =	0.11	0.52	0.08	1.0	0.82
	Source (water dispensed from tanker)	8	0	0.06	7.22	11.4
Community	n =	68	68	68	68	68
tankers	SD =	0.12	0.51	0.48	1	1.1
white b	Point-of-use (containers in household)	8	0	0.13	1.49	6.62
	<u>n</u> =	480	480	480	480	480
	SD =	0.15	0.15	0.47	5.4	6.2

Table 7 Average pH, residual Cl (mg/L), total Cl (mg/L), turbidity (NTU) and conductivity (mS/m) for all source and point-of-use water samples.

^an represents the sample size. ^b SD represents the standard deviation.

4.2. Open-top storage containers versus closed-top storage containers

Water from communal sources (standpipes and community tankers) was collected in either open-top or closed-top containers. Once filled, containers were transported to households where they were stored till ready for use. Presented below are microbial and physico-chemical results and analysis of water collected from open-top or closed-top storage containers (defined in Chapter 1), at the household level, which were supplied drinking water by communal water delivery systems (standpipes and community tankers) (Figure 12 and Table 8).

The overall microbial quality of water from standpipes (source water) was worse than the microbial quality of water from open-top and closed-top containers which they supplied. Heterotrophic organisms counts were slightly, but not significantly, higher in standpipe source water in comparison to water from standpipe-supplied open-top and closed-top containers (p=0.387 and p=0.556 for HPC in closed-top and open-top containers, respectively). Standpipe source water had significantly higher total coliform counts than water from open-top and closed-top storage containers which were supplied by standpipes (p=0.035 and P=0.032 respectively). Standpipe source water had significantly higher *E. coli* counts than water from open-top and closed-top storage containers which were supplied by standpipes (P=0.025 and P=0.021 respectively). Since no further deterioration in water quality between the source and point-of-use, for standpipe-supplied households, was noted it can be implied that the poor microbial quality of drinking water from standpipes is a major contributing factor to the poor microbial quality of point-of-use water (open-top and closed-top containers) (Figure 12).

Standpipe-supplied households displayed slightly, but not significantly higher, *E. coli* and total coliform counts in water stored in open-top containers than in water from closed-top containers (p=0.730 and p=0.215 respectively). However, standpipe-supplied households using closed-top storage containers had slightly, but not significantly, higher HPC counts in water at the point-of-use than households using open-top storage containers (p=0.601) (Figure 12).

As described in the methods chapter (Chapter 3), different containers were sampled, as representatives of source water, for open-top and closed-top storage tankers. Therefore when

referring to community tanker source (closed-top) the inference is that these tankers were sampled as the source water on the day that water from closed-top containers was sampled and *vice versa* for open-top containers. Community tankers which served as the source for closed-top water storage containers showed similar overall microbial counts to water from closed-top containers (Figure 12). There was no increase in microbial counts noted in water from closed-top storage containers supplied by community tankers when compared to the source. This again suggests that, the poor microbial quality of point-of-use water was the result of poor source water quality. This also indicates that the use of closed-top storage vessels may limit further microbial deterioration of drinking water at the point-of-use and during transportation of water to households (Figure 12).

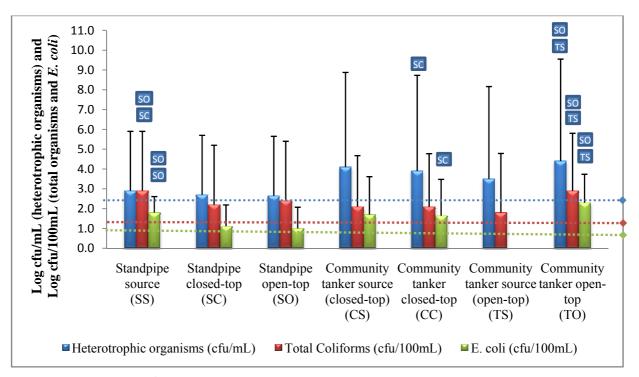
According to statistical testing, water from community tankers which served as the source for open-top storage containers had significantly lower HPC, total coliform and *E. coli* counts than water from the corresponding open-top storage containers which they supplied (p<0.001, p=0.003 and p<0.001 respectively). Water sampled from these open-top storage containers showed significant deterioration in microbial water quality in comparison to its corresponding source water. This suggests that microbial contamination of point-of-use water samples stored in open-top containers may be due to contamination during transport or storage of water in addition to poor quality of source water.

Generally, in community tanker-supplied households, water from open-top containers showed higher microbial counts than water from closed-top containers (Figure 12), with total coliform and *E. coli* counts being significantly higher in water from open-top containers in comparison to water from closed-top containers (p=0.001 and p=0.031 respectively).

Water from community tanker-supplied households using open-top storage containers showed significantly poorer water quality than did water from standpipe-supplied households using open-top water storage containers (p<0.001 for HPC, total coliforms and *E. coli*) (Figure 12 and Table 8). Water from community tanker-supplied households using closed-top storage containers showed significantly higher HPC and *E. coli* counts in comparison to water from closed-top containers supplied by standpipes (p<0.001 for HPC and *E. coli*). Therefore the microbial quality of community tanker-supplied point-of-use water was worse than that of standpipe-supplied point-of-use water. Overall, it has been shown that the quality

of source water and the use of open-top or closed-top water storage containers both contribute towards poor microbial quality of drinking water at the point-of-use.

DWAF guideline values for maximum number of heterotrophic organisms (blue), total coliforms (red) and *E. coli* (green) allowed in drinking water are presented in Figure 12 below. The markers on the graph indicate that all communal source (standpipe and community tanker) and point-of-use (open-top and closed-top containers) water samples are unsafe for human consumption and pose a level of risk of microbial infection to users due to the high microbial counts, which exceeded guideline maximum values, in water tested. The upper limit of all 95th percentiles suggests that the risk of infection posed to users by drinking water in this instance is substantial. Blue boxes indicate where microbial counts are significantly higher in water sampled from either the source, closed-top containers or opentop containers supplied by either standpipes or community tankers.



*Error bars represent the 95th percentile. Horizontal lines represent maximum values of heterotrophic organisms (blue), total coliforms (red) and *E. coli* (green) counts allowed in drinking water rendered safe for human consumption. Blue boxes represent significantly higher microbial counts than the sampling point referred to in the box.

Figure 12 Log arithmetic mean microbial counts in water from standpipe and community tankersupplied households using either open-top or closed-top containers for storage of drinking water. Table 8 gives the comparison of guideline values to arithmetic and geometric mean microbial counts as discussed previously (Chapter 3). Both, arithmetic and geometric mean microbial counts for standpipe source and point-of-use samples posed a potential risk of microbial infection to users in one parameter (HPC) and a substantial risk of microbial infection to users in two parameters (total coliforms and *E. coli*). All community tanker source and point-of-use water posed a substantial risk of microbial infection to users in three parameters (HPC, total coliforms and *E. coli*) with the exception of community tankers supplying open-top storage containers which had overall geometric and arithmetic mean counts of *E. coli* which placed it in the category for negligible risk of causing microbial infection (Table 8). No differences were noted between risk categories when comparing geometric and arithmetic means (Table 8).

The level of risk associated with the upper limit of all 95th percentile values for total coliforms and *E. coli* counts in water from all communal sources and points-of-use posed the same risk of microbial infection to users as did the arithmetic and geometric mean microbial counts in the same water (Figure 12, Table 8). The upper limit of 95th percentile values for heterotrophic organism counts also posed the same level of risk of infection in users as did arithmetic and geometric mean microbial counts for water from all community tanker and corresponding source samples. The only instance in which the upper limit of risk associated with 95th percentile values differed from risk associated with arithmetic and geometric mean microbial counts in the case of standpipe source and point-of-use samples. Here, heterotrophic organism counts in water posed a substantial risk of infection to users according to the upper limit of risk associated with the 95th percentile whilst arithmetic and geometric means suggested that heterotrophic organism counts only posed a potential risk of infection to users.

Table 9 shows the physico-chemical analysis for standpipe source and point-of-use samples and for community tanker source and point-of-use samples. Measured values for pH, conductivity and turbidity for all sample groups were within limits specified in DWAF and WHO guidelines (Chapter 2, Tables 1 and 2) (Table 9). Total chlorine levels were below the recommended levels of 2 to 2.5 mg/L for all sample groups. Residual chlorine levels were below recommended levels of 0.2 to 1.5 mg/L in all community tanker sample groups. Standpipe source and point-of-use sample groups had residual chlorine levels just above the recommended levels (Table 9).

Table 8 Arithmetic mean (AM) and geometric mean (GM) microbial counts in water from standpipe and community tanker supplied households using open-top or closed-top containers for water storage (heterotrophic plate counts (HPC) in cfu/mL; coliforms and *E. coli* in cfu/100mL) as compared to DWAF guidelines for water quality for domestic use.

Sample group		HPC (cfu/mL)	Risk	Total Coliforms (cfu/100mL)	Risk	<i>E. coli</i> (cfu/100mL)	Risk
Standpipe	AM n ^d =	794 40	PR ^b	794 40	SR ^c	63 40	SR
source	GM n=	902 38	PR	900 34	SR	95 23	SR
Standpipe	AM n=	501 240	PR	158 240	SR	13 240	SR
closed-top	GM n=	620 238	PR	430 56	SR	71 53	SR
Standpipe	AM n=	446 240	PR	251 240	SR	10 240	SR
open-top	GM n=	650 240	PR	440 59	SR	110 62	SR
Community tanker	AM n=	12 589 34	SR	125 34	SR	50 34	SR
source (closed-top)	GM n=	15 000 34	SR	130 29	SR	60 30	SR
Community tanker	AM n=	7 943 240	SR	125 240	SR	44 240	SR
(closed-top)	GM n=	17 000 239	SR	150 97	SR	57 60	SR
Community tanker	AM n=	3 162 34	SR	63 34	SR	0 34	NR ^a
source (open-top)	GM n=	13 000 34	SR	510 21	SR	0 0	NR
Community	AM n=	25 118 240	SR	794 240	SR	200 240	SR
tanker (open-top)	GM n=	255 000 240	SR	902 212	SR	250 192	SR

^aNR-negligible risk.

^bPR- potential risk.

^cSR- substantial risk).

^dn represents the number of samples used to derive mean values.

The standard deviation and number of data points for all physico-chemical properties is illustrated below in Table 9. Standard deviation for pH, residual chlorine and total chlorine for all sample groups was below the recommended level of one, thus indicating a good level of dispersion of data. The standard deviation for turbidity was above one for water from open-top containers supplied by standpipes and ground tanks. The standard deviation for conductivity was above one for water from all standpipe sample groups and for water from open-top containers supplied by community tankers.

Table 9 Average pH, free Cl (mg/L), residual	Cl (mg/L), turbidity (NTU) and conductivit	y (mS/m) for all
source and point-of-use water samples.		

Sample groups		рН	Residual chlorine (mg/L)	Total Chlorine (mg/L)	Turbidity (NTU)	Conductivity (mS/m)
		7.9	0.4	0.5	0.7	15
Standpipe source	n ^a	40	40	40	40	40
-	SD ^b	0.1	0.11	0.17	0.52	2.7
Standpipe container		7.9	0.2	0.3	1.1	15
(Closed-top)	n	240	240	240	240	240
` '	SD	0.13	0.14	0.19	1.0	2.8
Standpipe container		7.9	0.2	0.4	1.2	16
(Open-top)	n	240	240	240	240	240
	SD	0.13	0.15	0.21	2.4	2.8
Tanker source		8.1	0.03	0.07	1.7	12
(for samples from	n	34	34	34	34	34
closed-top containers)	SD	0.08	0.04	0.07	0.22	0.56
Tanker container		8.1	0.07	0.2	0.78	13
(Closed-top)	n	240	240	240	240	240
	SD	0.1	0.16	0.64	0.21	0.73
Tanker source		8.1	0.04	0.06	2.6	11
(for samples from	n	34	34	34	34	34
open-top containers)	SD	0.14	0.63	0.98	0.37	0.18
Tanker container		8	0.07	0.1	1.9	12
(Open-top)	n	240	240	240	240	240
	SD	0.19	0.14	0.19	1.41	3.1

^an represents the number of samples used to derive mean values. ^bSD represents the standard deviation.

4.3. Distribution of household demographics and its impact on the microbial quality of drinking water

The following results comprise information on the microbial and physico-chemical properties of drinking water sampled from three sample groups:

Households including children aged 0-5 years •

- Households including children aged >5-18 years and
- Households comprising adults only (>18 years).

Each of these groups was supplied with drinking water by standpipes, ground tanks or community tankers. The purpose of this was to establish if household demographics had an impact on the microbial quality of drinking water. The microbial content of drinking water sampled from households comprising each of the three groups is given in Figure 13.

Community tanker-supplied drinking water sampled from open-top and closed-top containers from all three sample groups (households including children aged 0-5 or >5-18 years and households comprising adults only) had significantly higher HPC counts than drinking water supplied by standpipes or ground tanks to households comprising any of the 3 sample groups (p<0.001 for all).

Drinking water supplied by standpipes and Cato Manor ground tanks had higher, but not significantly higher, HPC counts in water sampled from households comprising adults only than from households including children (Figure 13). This however, differed for community tanker and Sawpitts ground tank-supplied drinking water. In Sawpitts it was found that the highest HPC counts occurred in drinking water sampled from households including children aged 0-5 years, whilst in community tanker-supplied households the highest HPC counts occurred in drinking water sampled from households the highest HPC counts occurred in drinking water sampled from households the highest HPC counts occurred in drinking water sampled from households including occupants aged >5-18 years. Age distribution in this instance showed no specific association with the presence of heterotrophic organisms. Microbial content of drinking water varied with the age of household occupants and with the type of water supply system used (Figure 13).

Standpipe-supplied households had the highest, but not significantly so, *E. coli* and total coliform counts in water sampled from households including children aged >5-18 years in comparison to water from homes having adults only or children aged 0-5 years (Figure 13). In Cato Manor, the highest *E. coli* and total coliform counts in drinking water occurred in samples taken from households including children aged 0-5 years. These values were not significantly higher than *E. coli* and total coliform counts in drinking water samples from households including occupants aged >5-18 years, however *E. coli* counts in water from households including children aged 0-5 years.

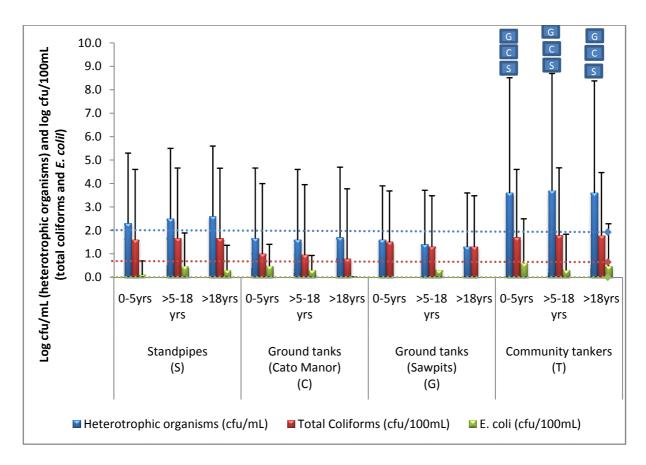
18 years was significantly higher than *E. coli* counts in water sampled from households comprising adults only.

Ground tank-supplied households in Sawpitts had the highest, but not significantly so, *E. coli* counts in drinking water samples from households including children aged >5-18 years and the highest total coliform counts in drinking water samples from households including children aged 0-5 years in comparison to drinking water samples from homes comprising adults only (*i.e.* households having no members under the age of 18 years). The converse applied for drinking water samples from households supplied by community tankers. Households supplied by community tankers had the highest, but not significantly so, *E. coli* counts in water samples from homes including children aged 0-5 years and the highest total coliform counts in water samples from homes including children aged 0-5 years and the highest total coliform counts in water samples from homes including children aged >5-18 years (Figure 13).

The trend for faecal contaminants (*E. coli* and total coliforms) was that water from households including children (either aged 0-5 or >5-18 years) generally had higher, but not significantly higher, *E. coli* and total coliform counts than water from households comprising adults only (>18 years of age).

This implies that faecal contamination of drinking water may be associated with the presence of children in a household. Microbial counts in water sampled from Cato Manor ground tank and community tanker-supplied households including the 0-5 year age group had the highest overall *E. coli* counts (Figure 13).

According to the guideline markers on Figure 13, heterotrophic organisms (blue) and *E. coli* (green) counts exceed maximum allowed levels in DWAF guidelines for water from standpipes and community tankers in samples taken from homes including all three age groups. Total coliform counts in water from any of the delivery systems, and for any age group, tested exceeded guideline values (Figure 13). Therefore water from standpipes, ground tanks and community tankers posed a risk of microbial infection to users with regard to heterotrophic bacteria and total coliform counts. Community tanker water posed the highest risk of microbial infection to users as presented by the upper limit of the 95th percentile shown as error bars on the graph (Figure 13). Blue boxes on the graph indicate where microbial counts are significantly higher than sampling points referred to in the box.



*Error bars represent the 95th percentile. Horizontal lines represent maximum values of heterotrophic organisms (blue), total coliforms (red) and *E. coli* (green) counts allowed in drinking water rendered safe for human consumption. Blue boxes represent significantly higher microbial counts than the sampling point referred to in the box.

Figure 13 Log arithmetic mean microbial counts of heterotrophic organisms, total coliforms and *E. coli* in drinking water supplied by community tankers, standpipes and ground tanks (ground tanks in Cato Manor and Sawpitts). Households are distinguished by age distribution of household members (*i.e.* households including children aged 0-5 years, households including children aged 5-18 years and households consisting of individuals over the age of 18 (adults).

The upper limit of risk associated with 95th percentile values differed from arithmetic mean and geometric mean microbial risk levels in the following way:

• Drinking water samples from standpipe-supplied households including occupants of all age groups tested had heterotrophic organism counts that posed a substantial risk of microbial infection to users whilst arithmetic means posed a potential risk of microbial infection to users in all three microbial parameters. Geometric mean microbial counts, however posed a potential risk of infection to users in one parameter (HPC) and a substantial risk of infection to users in two parameters (total coliforms and *E. coli*) (Figure 13 and Table 10).

• Heterotrophic organism and total coliform counts in drinking water from ground tanks in Cato Manor and Sawpitts posed a substantial risk of infection to users, with regard to the upper limit of risk represented by the 95th percentile, whilst arithmetic means posed a negligible risk and geometric means posed a potential risk of microbial infection to users (Figure 13 and Table 10).

All geometric mean and 95th percentile values representing all three microbial parameters in community tanker water, for all age groups, posed a substantial risk of infection to users (Figure 13, Table 10).

Table 10 comprises arithmetic and geometric mean microbial counts in drinking water sampled from standpipe, ground tank and community tanker-supplied households including children aged 0-5 years, children aged >5-18 years and households including occupants aged >18 years (also referred to in text as adults). The arithmetic and geometric means were used to compare point-of-use microbial drinking water quality to the South African and WHO water quality guidelines for drinking water (DWAF, 1996; WHO, 2005). The WHO and DWAF microbial guideline values are presented in Chapter 2 (Table 1).

The risk of microbial infection associated with HPC counts in standpipe-supplied households, according to the arithmetic and geometric mean, showed that water posed a potential risk of microbial infection to users in all three age groups. The arithmetic and geometric mean results however differed in terms of the risk category for total coliforms and *E. coli*. Whilst the arithmetic mean of total coliform and *E. coli* counts showed that water from standpipe-supplied households posed a potential risk of microbial infection to users, in all three demographic groups, geometric mean microbial counts of total coliform and *E. coli* showed that water posed a substantial risk of microbial infection to users in all three demographic groups (Table 10).

Drinking water samples from ground tank-supplied households in Cato Manor and Sawpitts had differing arithmetic and geometric mean microbial counts for HPC, total coliforms and *E. coli*, thus resulting in different categories of risk of microbial infection in users being associated with each mean type (*i.e.* arithmetic and geometric mean) (Table 10).

Arithmetic mean counts showed that occupants from ground tank-supplied households, in Sawpitts and Cato Manor, in all three sample demographic groups, were at a negligible risk of microbial infection as a result of the HPC counts and at a potential risk of microbial infection as a result of total coliform and E. coli counts. Geometric mean microbial counts of HPC and total coliforms in ground tank-supplied water placed users, in all three demographic groups, at a potential risk and substantial risk of microbial infection respectively, with the exception of water sampled from ground tank-supplied homes in Sawpitts which included children aged >5-18 years which had total coliform counts that posed a potential risk of microbial infection to users. Geometric mean microbial counts in water sampled from ground tank-supplied households in Cato Manor had E. coli counts that posed a substantial risk of microbial infection to users in all age groups except the >18 year group, in which E. coli posed a potential risk of infection to users. (Table 10). According to arithmetic mean microbial counts, community tanker-supplied drinking water posed a substantial risk of microbial infection to users, in all three demographic groups, in one parameter (HPC) and a potential risk of microbial infection to users in two parameters (total coliforms and E. coli). This differed slightly for geometric mean microbial counts, in which case, community tankersupplied drinking water posed a substantial risk of microbial infection to users, in all demographic groups, in all three microbial parameters (HPC, total coliforms and E. coli) (Table 10).

People from households supplied drinking water by community tankers in any of the three demographic groups were at the highest risk of microbial infection since the microbial counts in water at point-of-use exceeded recommended levels in DWAF and WHO guidelines to the greatest extent (Table 1). Microbial counts in point-of-use water supplied by ground tanks in Cato Manor, standpipes and community tankers had the highest *E. coli* and total coliform counts in households including children aged 0-5 or >5-18 years and fell in the category of potential risk, according to the arithmetic mean and the category of substantial risk according to the geometric mean, according to DWAF and WHO guidelines (Table 1) for safe drinking water. Therefore faecal contamination of water is more prominent in households including children than households comprising adults only (Table 10). Where the geometric mean differed from the arithmetic mean, the risk of microbial infection associated with geometric mean values were always higher than the risk of microbial infection associated with arithmetic mean measures central tendency whilst the arithmetic mean measures an average.

Table 10 Arithmetic mean (AM) and geometric mean (GM) microbial counts for point of-use water discriminated by age distribution (heterotrophic plate counts (HPC) in cfu/mL, total coliforms and *E. coli* in cfu/100mL).

Sample group	Age		HPC (cfu/ml)	Risk	Total Coliforms (cfu/ 100mL)	Risk	E. Coli (cfu/100mL)	Risk
	0-5 yrs	$ \begin{array}{c} \mathbf{AM}\\ \mathbf{n}^{\mathbf{d}} = \\ \mathbf{GM}\\ \mathbf{n} = \end{array} $	200 160 560 160	PR ^b PR	40 160 410 86	PR SR ^c	1 160 97 59	PR SR
Standpipes	5-18 yrs	AM n= GM	316 160 640	PR	46 160 470	PR	3 160 160	PR
	>18 yrs	n= AM n= GM n=	168 398 160 720 159	PR PR PR	48 45 160 430 41	SR PR SR	62 2 160 13 25	SR PR SR
	0-5 yrs	AM n= GM n=	46 89 340 87	NR ^a PR	10 89 310 74	PR SR	3 89 340 64	PR SR
Ground tanks (Cato Manor)	5-18 yrs	AM n= GM n=	40 89 380 85	NR PR	9 89 290 54	PR SR	2 89 320 51	PR SR
	>18 yrs	AM n= GM n=	50 89 350 78	NR PR	6 89 250 42	PR SR	1 89 2 14	PR PR
Ground tanks (Cato Manor)	0-5 yrs	AM n= GM n=	40 93 170 92	NR PR	32 93 120 41	PR SR	1 93 1 3	PR PR
Ground	5-18 yrs	AM n= GM n=	25 93 160 89	NR PR	20 93 99 32	PR PR	2 93 7 58	PR PR
tanks (Sawpits)	>18 yrs	AM n= GM n=	20 93 160 78	NR PR	20 93 110 43	PR SR	1 93 5 12	PR PR
Community tankers	0-5 yrs	AM n= GM n=	3 981 160 195 000 160	SR SR	50 160 230 64	PR SR	4 160 110 49	PR SR
	5-18 yrs	AM n= GM n=	5 011 160 100 000 154	SR SR	61 160 250 49	PR SR	2 160 77 32	PR SR

Sample group	Age		HPC (cfu/ml)	Risk	Total Coliforms (cfu/ 100mL)	Risk	E. Coli (cfu/100mL)	Risk
Community	5-18 yrs	AM n= GM n=	5 011 160 100 000 154	SR SR	61 160 250 49	PR SR	2 160 77 32	PR SR
tankers	>18 yrs	AM n= GM n=	3 981 160 48 000 123	SR SR	60 160 240 26	PR SR	3 160 93 12	PR SR

^aNR-negligible risk.

^bPR- potential risk.

^cSR- substantial risk).

^dn represents the number of samples used to derive mean values.

The pH, turbidity, conductivity, residual chlorine and total chlorine levels in water sampled from standpipe, ground tank and community tanker-supplied households including children aged 0-5 years, children aged >5-18 years and households comprising adults only (>18 years), are given below in Table 11. The standard deviations and means for all physico-chemical parameters are given in Table 11.

The pH of water for all demographic groups supplied by any of the three delivery systems was within the recommended levels according to WHO and DWAF guidelines (guidelines given in Chapter 2, Table 2). Point-of-use water from all delivery systems, except community tankers, had turbidity values which fell within the guideline values of one to five NTU. Community tanker water in all three demographic groups had turbidity values which exceeded guideline values, with households comprising adults only having the highest turbidity values.

Conductivity values for all demographic groups supplied by any of the delivery systems were all within the recommended guideline values of < 100 mS/m. Total and residual chlorine levels were within guideline values for standpipe and Cato Manor ground tank-supplied households for all demographic groups, however ground tank-supplied water in Sawpitts and Community tanker-supplied water had residual chlorine levels below recommended guideline levels for all demographic groups (Table 11).

Table 11 Average pH, residual Cl (mg/L), free Cl (mg/L), turbidity (NTU) and conductivity (mS/m) for all point-of-use water samples from households including children aged 0-5 years, children aged 5-18 years and households comprising occupants >18 years (adults).

Delivery systems	Demographics	рН	Residual Cl (mg/L)	Total Cl (mg/L)	Turbidity (NTU)	Conductivity (mS/m)
	0-5 years	7.9	0.2	0.4	1.4	15.5
	n =	160	160	160	160	160
	SD	0.16	0.15	0.21	3.3	2.8
Standpipes	>5-18 years	7.9	0.3	0.4	1.1	15.4
Standpipes	n =	160	160	160	160	160
	SD	0.11	0.15	0.20	1.18	2.8
	>18 years (Adults)	7.9	0.2	0.3	0.8	15.3
	n =	160	160	160	160	160
	SD	0.09	0.13	0.19	0.34	2.6
	0-5 years	8	0.4	0.6	0.6	14.9
	n =	89	89	89	89	89
Ground tanks	SD	0.15	0.18	0.20	0.43	1.7
(Cato Manor)	>5-18 years	8	0.4	0.5	0.6	14.9
	n =	89	89	89	89	89
	SD	0.15	0.17	0.19	0.38	1.6
	>18 years (Adults)	8	0.4	0.6	0.7	14.9
	n =	89	89	89	89	89
	SD	0.11	0.18	0.20	0.83	2.04
	0-5 years	8.1	0.1	0.1	0.5	12.5
	<u>n =</u>	93	93	93	93	93
~	SD	0.37	0.7	0.9	0.24	0.84
Ground tanks (Sawpitts)	>5-18 years	8.1	0.1	0.1	0.6	12.2
(Sawpitts)	n =	93	93	93	93	93
	SD 10	0.35	0.07	0.08	1.43	1.56
	>18 years (Adults)	8.1	0.1	0.1	0.5	12.6
	n =	93	93	93	93	93
	SD	0.34	0.07	0.08	1.43	1.56
	0-5 years	8	0.1	0.1	7	6.9
	n =	160	160	160	160	160
	SD	0.17	0.16	0.27	0.54	6.24
Community	>5-18 years	8	0.1	0.1	7	7
tankers	n =	160	160	160	160	160
	SD	0.14	0.17	0.70	5.4	6.2
	>18 years (Adults)	8	0	0.1	8.1	5.5
	n =	160	160	160	160	160
	SD	0.13	0.80	0.10	5.3	6.1

^an represents the sample size. ^bSD represents the standard deviation.

4.4. Links between microbial water quality, health outcomes, socio-economic status, household demographics and hygiene practices.

In order to establish if links existed between microbial drinking water quality, incidences of health outcomes and socio-economic status, data from an epidemiological study by Lutchminarayan, 2007 was merged with microbial data from the current study. These two data sets were statistically analysed to identify relationships between the parameters measured in that study and water quality as measured in the present study. Household demographics and hygiene practices of household members were also included in the analysis. It is acknowledged that a small sample size was used, but comparisons were made with respect to health outcomes despite this limitation because guideline values were phrased in terms of potential health impacts (microbial infection) and this comparison allowed contributing factors to health outcomes to be interrogated. However, the limitations posed by small sample size are acknowledged.

4.4.1. The relationship between point-of-use microbial drinking water quality, age distribution in households and health outcomes.

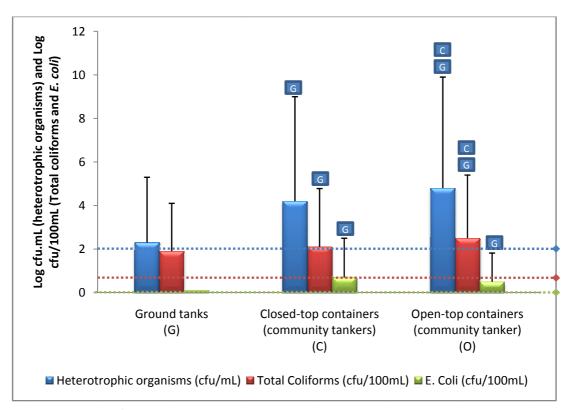
Microbial counts in point-of-use water, supplied by community tankers, stored in open-top and closed-top containers and microbial counts of point-of-use water sampled from ground tanks, are given in Figure 14. Data for standpipe-supplied households and Cato Manor community tanker-supplied households is not included in this section since the epidemiology study, by Lutchminarayan (2007), only made use of two areas used in the present study, namely Sawpitts (supplied water by ground tanks) and Mtamuntengayo (supplied water by community tankers).

There were significantly fewer *E. coli*, total coliforms and heterotrophic organisms in water from ground tanks when compared with that from closed-top and open-top storage containers supplied by community tankers (p=0.0005 for all) (Figure 14).

There were significantly more coliforms and heterotrophic organisms in water from open-top containers when compared with that from both closed-top containers (p=0.0005 and p=0.001, respectively) and ground tanks (p=0.0005 for both) (Figure 14). Even though community tanker source water for open-top containers was of a poor quality the deterioration of

microbial water quality at the point-of-use (open-top container) was significant in comparison to source water quality (Figure 12).

DWAF guideline values (Table 1), as marked on Figure 14, indicate that water from ground tanks, open-top containers and closed-top containers all have heterotrophic and total coliform counts which exceed the maximum allowed levels as per DWAF guidelines and hence pose a substantial risk of microbial infection to users. The upper limit of risk of the 95th percentile, for heterotrophic organisms and total coliforms, places water from all three storage systems (ground tanks, open-top containers and closed-top containers) in the category of posing a substantial risk of infection to users. Community tanker-supplied water from open-top and closed-top containers also posed a substantial risk of microbial infection to users with regard to the upper limit of risk presented by the 95th percentile for *E. coli*. Blue boxes on the graph indicate where microbial counts are significantly higher.

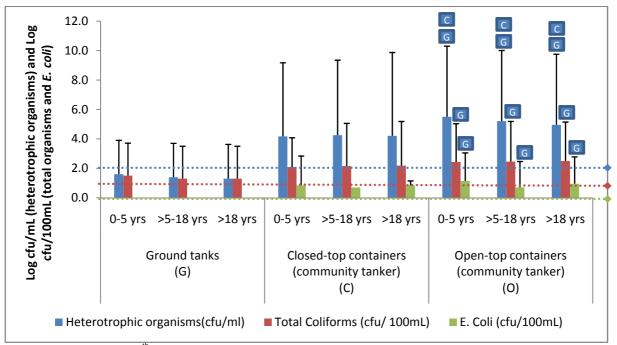


*Error bars represent the 95th percentile. Horizontal lines represent maximum values of heterotrophic organisms (blue), total coliforms (red) and *E. coli* (green) counts allowed in drinking water rendered safe for human consumption. Blue boxes represent significant differences in microbial counts.

Figure 14 Log arithmetic mean microbial counts at point-of-use for households using ground tanks as a water delivery and storage system and for community tanker-supplied households using closed-top and open-top containers as water storage systems.

The relationship between age distribution in households (households including children aged 0-5 years, households including children aged >5-18 years and households comprising adults only) and microbial quality of drinking water at the point-of-use is given below in Figure 15.

The highest heterotrophic organism and *E. coil* counts were observed in water from households including children aged 0-5 years and which used open-top containers for water storage. Heterotrophic organism counts were significantly higher in this group in comparison to all closed-top container and ground tank-supplied household groups. Households comprising adults only and which used open-top containers for water storage displayed the highest total coliform counts, but not significantly so (Figure 15). The upper risk limit of the 95th percentile indicates that heterotrophic organisms and total coliforms are present in high enough quantities in all three storage systems (ground tanks, open-top and closed-top containers) to pose substantial risk of microbial infection to users. For ground tank-supplied households *E. coli* levels were very low and according to upper limit of risk presented by the 95th percentile, no risk of microbial infection was posed to users. Blue boxes on the graph indicate where microbial counts are significantly higher.



*Error bars represent the 95th percentile. Horizontal lines represent maximum values of heterotrophic organisms (blue), total coliforms (red) and *E. coli* (green) counts allowed in drinking water rendered safe for human consumption. Blue boxes represent significant differences in microbial counts.

Figure 15 Arithmetic mean microbial counts for households including children aged 0-5 years or children aged >5-18 years and households comprising adults only (>18 yrs), which use ground tanks, open-top or closed-top containers for domestic water storage and use.

Water quality from open-top and closed-top containers, supplied by community tankers, to households comprising adults only and households including children aged 0-5 and >5-18 years (Table 12), was compared to the South African water quality guidelines for drinking water (Table 1) (DWAF, 1996).

Arithmetic mean microbial counts for ground tank water placed users at a negligible risk of microbial infection in one parameter (HPC) and at a potential risk of obtaining microbial infection in two parameters (total coliforms and *E. coli*). The geometric means however showed that water from ground tanks had HPC and *E. coli* counts which posed a potential risk of causing microbial infection to users. Geometric mean counts of total coliforms showed that water from households including children aged 0-5 years and households comprising adults only posed a substantial risk of microbial infection to users (Table 12).

All arithmetic and geometric mean counts of HPC and total coliforms in community tankersupplied water stored in open-top or closed-top containers posed a substantial risk of infection to users. Arithmetic and geometric mean microbial counts of *E. coli* from water stored in open-top containers, sampled from households including children aged 0-5 years posed a substantial risk of microbial infection to users. Water samples from households with children aged >5-18 years and households with adults only had *E. coli* counts which posed a potential risk of infection to users (Table 12). Therefore water from all three point-of use samples (open-top and closed-top storage containers and ground tanks) did not meet DWAF guidelines as being safe for drinking purposes.

The level of risk presented by the 95th percentile differed from the level of risk presented by the geometric and arithmetic mean in the following instances:

• 95th percentile values of heterotrophic bacteria and total coliforms placed ground tank water in the category of causing substantial risk of microbial infection to users. Arithmetic means placed the same water at a negligible risk of microbial infection for heterotrophic organisms and at a potential risk of causing microbial infection in users for total coliforms. The geometric mean microbial counts of heterotrophic organisms placed ground tank water in the category of causing potential risk of infection to users.

• The upper limit of risk presented by the 95th percentile for *E. coli* in water from opentop containers, supplied by community tankers, placed this water in the category of causing a substantial risk of infection to users in households including children aged >5-18 years and in households comprising adults only. This differed for the geometric and arithmetic mean which rendered this water as posing a potential risk of infection to users.

The arithmetic mean, geometric mean and upper limit of the 95th percentile for heterotrophic organisms and total coliforms provided the same result with regard to the level of risk associated with water from open-top and closed-top containers, which were sampled from households comprising any of the three age groups studied. The finding here showed that all three statistical measures rendered water as posing a substantial risk of microbial infection to users (Figure 15, Table 12).

Table 12 Arithmetic mean (AM) and geometric mean (GM) microbial counts from ground tank and community tanker-supplied point of-use water (water in open-top and closed-top containers) distinguished between by age distribution (heterotrophic plate counts (HPC) in cfu/mL, total coliforms and *E. coli* in cfu/100mL).

Point-of-use	Age		HPC (cfu/mL)	Risk	Total coliforms (cfu/100mL)	Risk	<i>E. coli</i> (cfu/100mL)	Risk
	0-5 yrs	AM n ^d =	40 93	NR ^a	32 93	PR	1 93	PR
		GM n=	170 92	PR ^b	120 41	SR ^c	1 3	PR
Ground	>5-18 yrs	AM n=	25 93	NR	20 93	PR	2 93	PR
tanks		GM n=	160 89	PR	99 32	PR	7 58	PR
	> 18 yrs	AM n=	20 93	NR	20 93	PR	1 93	PR
		GM n=	160 78	PR	110 43	SR	5 12	PR
	0-5 yrs	AM	14 858 80	SR	122 80	SR	7 80	PR
Closed-top containers		n= GM n=	15 683 80	SR	138 45	SR	9 42	PR
(Community tankers)	>5-18 yrs	AM	17 758 80	SR	142 80	SR	5 80	PR
		n= GM n=	18 598 78	SR	173 32	SR	8 12	PR

Point-of-use	Age		HPC (cfu/mL)	Risk	Total coliforms (cfu/100mL)	Risk	<i>E. coli</i> (cfu/100mL)	Risk
Closed-top containers (Community tankers)	> 18 yrs	AM n= GM n=	16 235 80 16 785 74	SR SR	153 80 159 28	SR SR	7 80 8 9	PR PR
Open-top	0-5 yrs	AM n= GM n=	318 025 80 327 654 79	SR SR	271 80 302 23	SR SR	14 80 16 43	SR SR
(Community tankers)	>5-18 yrs	AM n= GM n=	161 104 80 176 549 77	SR SR	287 80 305 22	SR SR	5 80 7 15	PR PR
	> 18 yrs	AM n= GM n=	90 125 80 102 692 72	SR SR	314 80 421 16	SR SR	9 80 10 14	PR PR

^aNR-negligible risk.

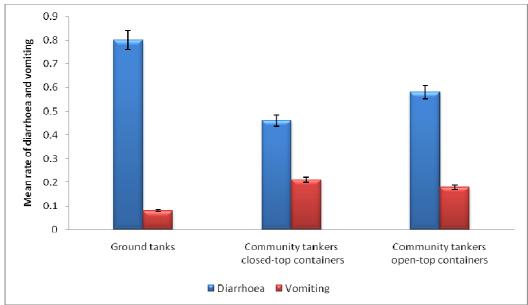
^bPR- potential risk.

^cSR- substantial risk).

^dn represents the number of samples used to derive mean values.

The relationship between water quality at the point-of-use (storage containers and ground tanks) and health outcomes (diarrhoea and vomiting) is presented in Figure 16. The rate of diarrhoea was highest among members of households using ground tanks at the point-of-use, followed by members of households using water from open-top containers and then water from closed-top containers, where rate is used as defined in the Methods chapter (Figure 16). However, these differences were not significant. The rate of diarrhoea showed an opposite trend to water quality and to risks according to guidelines. This therefore suggests that factors other than the microbial quality of water contribute to the rate of diarrhoea.

Rate of vomiting was lower in households using ground tanks for water storage than in those using open-top or closed-top storage containers. Again, this was not significant (Figure 16).



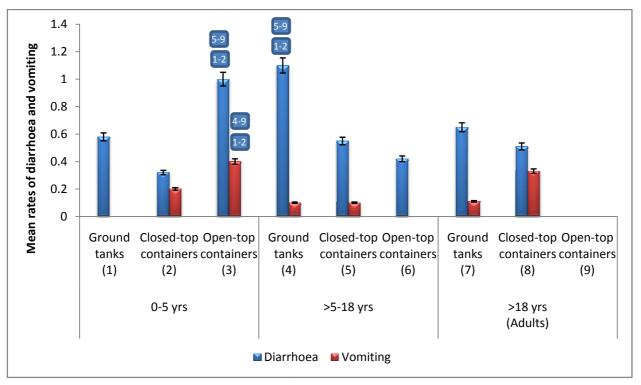
*Error bars represent the 95th percentile.

Figure 16 Mean rates of diarrhoea and vomiting in households using ground tanks, open-top or closed-top containers for water storage and supply.

The relationship between water quality at the point-of-use (storage containers and ground tanks), health outcomes (diarrhoea and vomiting) and age distribution in households was also investigated (Figure 17).

Figure 15 shows that generally, the microbial quality of drinking water in households including children aged 0-5 and >5-18 years of age was worse than in households comprising adults only. Households using ground tanks and which had a demographic distribution of children in the 0-5 year range had water with the lowest E. coli and total coliform counts (Figure 15). Members of these households presented with the highest rate of vomiting, this rate was significantly higher than all other demographic groups receiving drinking water from ground tanks, open-top or closed-top containers (p < 0.001) (Figure 17). Even though the E. coli counts in ground tank water was less than that in water from open-top or closed-top storage containers (Figure 15), the highest rate of diarrhoea still occurred amongst members aged >5-18 years in households using ground tanks for water storage, this rate was significantly higher than all other demographic groups using ground tanks, open-top or closed-top storage containers as point-of-use water supplies (p<0.001) (Figure 15 and Figure 17). This therefore suggests, again, that the rate of health outcomes is influenced by factors other than the microbial quality of drinking water. A possible explanation for the high rate of diarrhoea in this group could be the fact that this group comprises children of school-going age and hence infections may be picked up at school. On a qualitative basis, the rate of all

health outcomes was significantly higher in households with children aged 0-5 and >5-18 years using open-top containers or ground tanks, respectively, as the point-of-use, in comparison to all other demographic groups using ground tanks, open-top or closed-top containers as a point-of-use water supply (Figure 17).



*Error bars represents the 95th percentile. Blue boxes represent significant differences in mean rates of diarrhoea and vomiting amongst the various age groups analysed.

Figure 17 Mean rates of diarrhoea and vomiting in households using ground tanks, open-top or closed-top containers for domestic water storage at the point-of-use. Rate of diarrhoea and vomiting is also distinguished between by age distributions.

Overall no direct relationships were found between water quality and health with regard to vomiting and diarrhoea amongst all three age groups analysed.

4.4.2. Relationship between socio-economic factors and health

Although intervention areas (areas provided with UD toilets and ground tanks) were provided with UD toilets, not everyone used them. Some household members still preferred to practice open defaecation. Table 13 shows that, for those households where not everyone used the UD toilet, there was a higher rate of diarrhoea (p=0.036). If the outside toilet was used but not clean, significantly higher rates of diarrhoea were observed among toilet users (p=0.028). This indicates that the use of UD toilets and practice of good hygiene are closely related to the aversion of cases of diarrhoea.

Table 13 Association of sanitation with rate of diarrhoea. *P values lower than 0.05 indicates significance *i.e.* a significantly higher rate of diarrhoea associated with a specific variable.

In variable based on household and observational questionnaire		Result	Statistical test performed	Level of significance
Does everyone in the family use the UD toilet?	NO	significantly higher diarrhoea	Mann-Whitney	p=0.036
Is the outside toilet clean?	NO	significantly higher diarrhoea	Mann-Whitney	p=0.028

*P values lower than 0.05 indicates significance *i.e.* a significantly higher rate of diarrhoea associated with a specific variable.

4.4.3. Relationship between socio-economic factors and water quality

Table 14 represents results on the statistically significant relationships which were found between microbial parameters (such as *E. coli* counts) and sanitation practices. Sanitation practice data was captured in the household and an observational questionnaire study conducted by Lutchminarayan 2007. Statistical analysis of the combined microbial data and epidemiological data was conducted as part of the current study.

Table 14 shows that there were higher E. coli counts in drinking water if:

- children in the family did not use UD toilets most of the time as opposed to when children used UD toilets more often (p=0.034),
- when UD toilet seats were not clean (p=0.035) and
- when fewer books in the household was observed (p=0.028).

This indicates that contamination of drinking water by users is probably related to sanitation and hygiene practices and possibly a lack of knowledge of proper hygiene behaviour to prevent illness.

In variable based on household and observational questionnaire		Result	Statistical test performed	Level of significance
Do your children use UD toilets all the time?	NO	significantly higher <i>E. coli</i> counts	Mann-Whitney	p=0.034
Were UD toilet seats observed to be clean?	NO	significantly higher <i>E. coli</i> counts	Mann-Whitney	p=0.035

Table 14 Association of sanitation and *E. coli* counts in drinking water.

*P values lower than 0.05 indicates significance *i.e.* a significantly higher count of *E. coli* associated with sanitation.

Hands are more easily contaminated with faecal matter when toilet seats are not clean. The insertion of dirty hands into drinking water storage containers or the touching of ground tank taps with contaminated hands would result in faecal contamination of drinking water at the point-of-use. Table 15 shows that, there were higher levels of total coliforms and heterotrophic organisms in drinking water of those households where people practiced open defaecation as opposed to using outside toilets (p=0.034 and p=0.028). When open defaecation is practiced, people probably do not wash their hands after defaecation and this would result in contamination of drinking water supplies.

Table 15 also shows that, where households had radios and refrigerators (measures of socioeconomic status) heterotrophic organism counts in water at the point-of-use were significantly lower (p=0.046 and p=0.037 respectively). This suggests that more developed households or people with better socio-economic status, were more aware of good health and hygiene practices.

This could have resulted in reduced contamination of drinking water at the point-of-use. The absence of hand-washing facilities was related to higher coliform counts in water (p=0.0005) (Table 15). Therefore the lack of hygiene practices, such as washing of hands after using toilets, was linked to overall poorer water quality.

Table 15 Association of sanitation, education indicators and poverty indicators on microbial counts in drinking water.

In variable based on observational and household questionnaires	Result	Statistical test performed	Level of significance
Open defaecation practiced more than use of UD toilets	significantly higher total coliforms	Mann-Whitney	p=0.034
Open defaecation practiced more than use of UD toilets	significantly higher heterotrophic organisms	Mann-Whitney	p=0.028
Presence of radios and refrigerators in households	significantly lower heterotrophic organisms	Mann-Whitney	p=0.037 (fridge) p=0.046 (radio)
Absence of hand washing facilities	significantly higher total coliforms	Mann-Whitney	p=0.0005

*P Values lower than 0.05 indicate significance *i.e.* a significantly higher microbial count associated with sanitation, education and poverty indicators.

4.4.4. Relationship between age/place and socio-economic factors

Mtamuntengayo was compared to Sawpitts with regard to socio-economic factors and health (Table 16). Mann-Whitney tests showed that there were more households with books in Mtamuntengayo than in Sawpitts. Fischer's exact tests showed that

- significantly fewer hand-washing facilities
- significantly fewer responses to hygiene questions and
- significantly fewer cases of soap provided

were observed in households from Sawpitts than in households in Mtamuntengayo (p=0.0005, p=0.005 and p=0.001, respectively). A significantly higher number of people in Mtamuntengayo said they used soap for hand-washing, compared to Sawpitts (p=0.007). Together these results indicate that there was a lower standard of education and of hygiene practiced in Sawpitts than in Mtamuntengayo (Table 16). Households including children had higher unemployment than households with adults only (p=0.001). This was expected since caregivers (especially of younger children) would not be formally employed. No other significant differences were observed between age distributions and socio-economic factors. When place and age distributions were accounted for concurrently, cross tabulations showed that people in Mtamuntengayo practiced open defaecation more often than people in Sawpitts (Table 16). This could explain why Mtamuntengayo point-of-use (open-top and closed-top

storage containers supplied by community tankers) drinking water had higher overall microbial counts than water from ground tanks (Figure 15). Open defaecation allows an easier route for faecal contamination of hands. Hand-washing facilities are generally not provided in the vicinity of areas used for open defaecation. Insertion of hands into storage containers either to wash hands or to remove water for drinking purposes could result in contamination of drinking water and hence illness such as diarrhoea.

Table 16 Comparison of Sawpitts with Mtamuntengayo with regard to educational indicators,
hygiene indicators and sanitation indicators.

Sawpitts	Mtamuntengayo	Statistical test performed	Level of significance
	More books in Mtamuntengayo	Mann-Whitney	p=0.033
Fewer hand wash facilities		Fischer's Exact	p=0.0005
Fewer responses to hygiene questions		Fischer's Exact	p=0.005
Fewer observations of soap present		Fischer's Exact	p=0.001
	More people wash hands with soap	Mann-Whitney	p=0.007
	More people practice open- defaecation	Cross tabulation	

Overall, the poorest point-of-use microbial water quality occurred in households using opentop containers for water storage. However, these households did not always have the highest rates of health outcomes. Water from households including children collected in open-top water storage containers had the highest microbial counts, members from these households had highest rates of health outcomes. Therefore, when age and point-of-use water quality are analysed simultaneously, they can be linked to health outcomes. This is reviewed further in the following discussion chapter.

CHAPTER 5: DISCUSSION

The current study investigated the relationship between microbial quality of drinking water at the source and its corresponding point-of-use in three improved water delivery systems (standpipes, ground tanks and community tankers). Only households provided with ground tanks received the full three-pronged eThekwini intervention programme (Chapter 2). This intervention by eThekwini Municipality comprises provision of ground tanks, UD toilets and a hygiene program known as the WASH program

as a package to households. Individuals are educated in care and use of UD toilets and ground tanks, and on how they can become sick if they do not practice good personal hygiene such as washing of their hands with soap (Chapter 2).

5.1. Point-of-use versus source water quality from ground tanks, standpipes and community tankers

When comparing drinking water sources (standpipes, community water tankers and taps at eThekwini Municipality laboratory), results showed that the quality of drinking water from standpipes and community tankers was the poorest (Figure 11). Source water for ground tanks (water sampled from taps at eThekwini Municipality laboratory, representative of water from the high-pressure distribution system) was safe for human consumption whilst its point-of-use water (sampled from the low-pressure tap on the ground tank) posed a potential risk of microbial infection to users because of the elevated numbers of *E. coli* and total coliforms (Table 6). Semi-pressure standpipe and low-pressure community tanker point-of-use water (sampled from household containers) (Figure 9 and Figure 10) was not safe for human consumption because of the high counts of all three indicator microorganisms, with *E. coli* and total coliform counts posing a substantial risk of infection in users (Figure 11). The presence of *E. coli* in this water implies that there is a high risk of faecal contamination of water and that consequently users are at risk of becoming ill from microbial pathogens of faecal origin (DWAF, 2005; WHO, 2005).

Communal water sources (standpipes and community tankers) had high *E. coli*, total coliform and heterotrophic organisms counts (Figure 11) indicating that poor water quality at the point-of-use (open-top and closed-top storage containers) was also attributable to poor source

water quality and not necessarily only to storage, transportation or external factors. This provides grounds for concern as it has been well documented that water quality further deteriorates during storage and transport due to contamination induced by handling of water through dipping utensils, hand contact, airborne and insect components, and the use of dirty storage and transport containers, which are all at some stage in direct contact with drinking water (Jagals *et al.* 1997; Roberts *et al.* 2001; Momba and Notshe, 2003; Trevett *et al.* 2005; Trevett *et al.* 2008). If source water is of a poor microbial quality before collection, transportation and storage, then the quality of this water is likely to deteriorate even further due to external contaminant exposure.

Poor microbial quality of community tanker source water is likely to be due to biofilm presence, contamination of dispensing devices from tankers, contamination of water in the tanker by dust during transport and inefficient or inadequate water treatment. The hygiene of water tankers is also of concern. Communal tankers and their interiors are seldom washed and scrubbed due to their inaccessible nature. When questioned on how often tankers are washed, representatives of the municipality stated that they were only washed when water appeared to be dirty or if consumer complaints were noted. Water is also stored in the tanker for hours at a time. Although this allows a longer contact time for chlorine disinfection, it should be viewed as negative since biofilm formation and development are encouraged. It is well documented that microorganisms require a surface with nutrients and a flow of water to produce biofilm (LeChevallier et al. 1987; Block et al. 1993; Camper, 1994; Costerton et al. 1994). In this instance both are present and the interior surfaces of the tanker serve as an ideal environment for biofilm growth and attachment. Water from communal tankers also displayed low total and residual chlorine levels, indicating that microbial growth was offered almost no resistance. This however has not been confirmed in this study. Turbidity levels in water from communal tankers were above the recommended levels in WHO and DWAF guidelines (Table 2 and Table 7). High turbidity is typically associated with increased biofilm formation as a result of more particles being available to serve for attachment and nutritional purposes (Costerton et al. 1994). Therefore communal tankers need to be thoroughly washed on a regular basis, chlorine levels need to be monitored regularly and turbidity needs to be controlled. Such control could be implemented by increasing coagulant content or improving filtration techniques during water treatment, if the turbidity at the treatment plant is the source of turbidity in the tanker. If this is not the case, the source of turbidity needs to be identified and eliminated accordingly. Emphasis should also be placed on educating tanker operators and users on proper water hygiene practices.

Further microbial contamination of water from community tankers occurred at the point-ofuse (open-top or closed-top storage containers). Water collected from community tankers into open-top or closed-top storage containers had increased *E. coli*, total coliform and heterotrophic plate counts (Figure 11). This implies that water use practice at the point-of-use contributes further to decrease in water quality. The high total chlorine levels in community tanker-supplied point-of-use water could be attributed to Jik® or possibly some bleach based product used for cleaning of storage containers. This however has not been confirmed in this study.

Pilot studies in Cato Manor, in which one communal standpipe and five standpipe-supplied households, using either open-top or closed-top storage containers was sampled, found that water from standpipes was safe for drinking and did not pose any health risks (Arjun *et al.* 2004). This differed from findings in the current study, in which it was established that water from standpipes was generally not safe for drinking as it posed a potential risk of microbial infection in terms of one parameter (heterotrophic organisms) and a substantial risk of infection in terms of two parameters (total coliforms and *E. coli*), as per DWAF and WHO guidelines (Table 6).

The conflicting results of the pilot study could be attributed to the small number of samples (one standpipe and five standpipe-supplied households). During sampling of the 2005 study, it was observed that standpipe cleanliness and the surrounding environment were compromised as a result of people using the area immediately around standpipes for washing of clothes and vehicles and thereafter touching the tap nozzle with dirty hands. Cloths tied around a standpipe tap were also found. This allowed a mechanism or route of entrance for microbes into the point-of-use water as water from the standpipe taps passes along the dirty cloth and into the collection container. According to the study by Arjun *et al.* (2004) in the same area, the quality of water at the source for standpipes varied, with quality apparently linked to water usage at or near the standpipe. According to Carden *et al.* (2006) greywater management in the vicinity of communal water sources presents a hazard to water quality and hence to health. Even though sterile techniques were used during sampling of standpipe

water, the microbial quality of this water was still found to be poor. Therefore, treatment of water may have been inefficient or contamination of water in the distribution system may have occurred. Furthermore, education on hygiene practices and proper water use behavior to prevent additional contamination of the standpipe source water may prove beneficial in maintaining the microbial quality of drinking water.

With regard to ground tanks, source water (water sampled from taps at eThekwini Municipality labs, representative of treated water in the distribution system) was found to be of acceptable microbial quality with no health risk posed to consumers (Table 6). This demonstrated that water supplied from the distribution system had been adequately treated. However, the point-of-use water (water sampled from ground tanks) deteriorated relative to the source (Figure 11). This could be a result of biofilm growth in the ground tank. Studies by Arjun *et al.* (2004) confirmed the presence of biofilm in ground tanks sampled from the Cato Manor region in Durban.

Other factors that could have contributed to deterioration of ground tank water quality include poor hygiene practices of people resulting in contamination of ground tank taps during handling, and the introduction of foreign particles into ground tanks as a result of open or faulty ground tank covers. During sampling, it was observed that ground tank covers were not always tightly attached. This could result in the lids being blown off by wind and hence introduction of foreign particles into the water. Several ground tank users also said that people sabotaged their ground tank water by introducing litter into their water.

In Sawpitts and Cato Manor, water from ground tanks and its corresponding source showed residual chlorine levels below WHO and DWAF recommended levels (Table 2 and Table 7). This implies that the chlorine levels after treatment and transportation may not have been enough to destroy microorganisms in ground tanks that originate either from biofilm growth or from exterior induction into water.

Possible solutions to maintain the quality of drinking water in ground tanks as it is received from the distribution system, would be to implement further awareness programs on the importance of keeping ground tanks closed, washing of hands before handling areas close to the ground tank tap nozzle, cleaning ground tanks on a regular basis and possibly even addition of small amounts of chlorine into water stored in ground tanks.

5.2. Open-top storage containers versus closed-top storage containers

Standpipes and community tankers are communal water delivery systems. They require that water be collected, transported and stored within the household until required for use. In the current study, all households investigated used either open-top or closed-top containers for drinking water storage. Most households used dipping utensils to remove water from storage vessels. In contrast to standpipes and community tankers, ground tanks do not require collection, transport or storage of its water in open-top or closed-top containers. Ground tanks are situated on the property of individual households. They serve as a storage vessel for water. Water enters ground tanks from the distribution system. It is stored in these tanks for a maximum of 24 hours. During this time water is used as required. After 24 hours all tanks are refilled with water from the distribution system. Water is typically withdrawn via a tap at the base of the tank immediately prior to use. Hand-contact with water, unlike that which may occur in storage containers, is minimised.

Water from open-top and closed-top storage containers supplied by community tankers was of an overall poorer microbial quality than community tanker source water (Figure 12). This indicated that deterioration of microbial water quality occurred during storage. Water from open-top containers had higher microbial counts, lower residual chlorine levels and higher turbidity levels than water from closed-top containers (Figure 12 and Table 9). The high turbidity and low residual chlorine levels in water from open-top and closed-top storage containers could account for increased microbial counts in these samples since there are more suspended particles available for microorganisms to attach to and a higher nutrient supply to support growth and proliferation of microbes (Momba and Notshe, 2003). *E. coli* and heterotrophic organism levels in water sampled from open-top containers, supplied by community tankers, were significantly higher than in the corresponding source water from the tanker (Figure 12). This implies that microbial contamination of drinking water occurs predominantly at the point-of-use. Such contamination can be a result of the type of storage container (Trevett *et al.* 2004). However, the presence of *E. coli* suggests that there was also a source of faecal contamination.

Standpipe-supplied water stored in open-top containers had higher total coliform and heterotrophic organism counts than water stored in closed-top containers (Figure 12). Since no significant difference was detected between microbial quality of standpipe source and point-of-use water, contamination of water sampled from storage containers is probably due to poor source water quality.

Water stored in open-top containers was generally more prone to microbial contamination than water stored in closed-top containers. This agrees with studies by Hammad and Dirar (1982), Deb *et al.* (1986), Empereur-Bissonnet *et al.* (1992), Jagals *et al.* (1997), Mirza *et al.* (1997) and Roberts *et al.* (2001). It is probably associated with the easy access of dust, foreign particles, animal hairs and waste into water in these containers as has been reported by Shiffman *et al.* (1978), Jagals *et al.* (1997), Hoque *et al.* (1999), Momba and Notshe (2003). All standpipe sample groups posed a substantial risk of infection to users in at least two microbial parameters (Table 8). However, community tanker-supplied water at the point-of-use posed the highest risk of microbial infection to users, with all 3 microbial parameters exceeding the DWAF and WHO guideline limits (Table 2 and Table 8).

5.3. Distribution of household demographics and its impact on the microbial quality of drinking water

Drinking water in households including children aged 0-5 years or >5-18 years was at increased risk of faecal contamination in comparison to households comprising adults only (>18 years) (Figure 13 and Table 10). The relationship between the microbial deterioration of drinking water quality at the point-of-use and the presence of children aged 0-5 years in households has also been reported by Deb *et al.* (1986), Yeager *et al.* (1991) and Qadri *et al.* (1992). However, few studies have investigated the link between point-of-use water quality and the presence of older children (>5-18 years) in households. Children have been shown to contribute to faecal contamination of drinking water through the insertion of faecally contaminated hands into storage vessels when removing water for use. The dipping utensils used to remove drinking water could also be faecally contaminated by their hands and hence result in contamination of drinking water upon insertion into water containers (Hoque *et al.* 1995; Roberts *et al.* 2001; Trevett, 2003; Trevett *et al.* 2005). Children also contribute to microbial deterioration of drinking water by playing near water storage vessels which allow

for the entrance of dust and sand particles into drinking water (Trevett *et al.* 2004; Fewtrell *et al.* 2005).

Studies by Yeager *et al.* (1991), Qadri *et al.* (1992) and Mahmud *et al.* (2001) showed that the use of drinking water storage vessels which were devoid of taps and hence required dipping utensils for removal of water from storage vessels was associated with significantly increased microbial content and incidences of diarrhoeal disease in children less than 5 years of age. Jagals *et al.* (1997) and Mirza *et al.* (1997) demonstrated that microbial contamination of drinking water was most prevalent in households including children less than 5 years of age in which open-top containers were used for storage of drinking water at the point of use. It was therefore a surprising observation in the current study, to find that drinking water from ground tank-supplied households in Cato Manor including children aged 0-5 or >5-18 years were found to contain higher *E. coli* counts (in terms of the geometric mean) than water from standpipe, community tanker and Sawpitts ground tank-supplied households including children aged 0-5 or >5-18 years (Table 10).

During sampling it was observed that several ground tanks did not have their lids securely in place. Children often played in their yards close to where ground tanks were located. Several yards displayed unsanitary environments, in which excreta from humans or animals were present on the ground. This provided an easy route for faecal contamination of children's hands. The absence of lids on some ground tanks provides a route for microbial contamination by children. Children running in areas close to ground tanks could cause dust, sand and faecal particles to rise and contaminate drinking water further. Studies by Trevett et al. (2004) demonstrated that children play with drinking water and throw objects such as stones into water which could allow for the transfer of faecal pathogens from hands or the environment to water. This could have also occurred in the current study and thus provide an explanation for the high E. coli content found in ground tank drinking water samples taken from households including children. The efficiency of ground tanks in terms of reducing microbial deterioration of drinking water in households comprising children however, is not disputed, since ground tanks in Sawpitts had the lowest overall microbial content in comparison to community tanker, standpipe and Cato Manor ground tank-supplied households (Figure 11). The high faecal contamination of Cato Manor ground tank drinking water samples is most likely due to improper hygiene practices or lack of care of ground tanks, in terms of closure of tanks with lids and placement of containers high enough to prevent children from throwing objects into ground tank water supplies.

The higher content of faecal and environmental bacteria (*E. coli* and total coliforms) found in water, from standpipe and community tankers, which is stored in households including children aged 0-5 and >5-18 years in comparison to households comprising adults only (Figure 13) could also be attributed to unsanitary environments resulting in easy faecal contamination of children's hands and transfer of faecal pathogens from the hands of children into drinking water stores through either playing with water in storage containers, or insertion of hands or faecally contaminated dipping utensils into storage vessels to remove drinking water for use.

The faecal content present on the hands of children could be attributed to the lack of hygiene practices (such as washing of hands after using the toilet) or improper sanitation practices. Playing in areas where open defaecation is practiced would provide increased chances of picking up faecal pathogens and hence increased contamination of hands with faecal bacteria (VanDerslice and Briscoe, 1995; Kaltenthaler *et al.* 1996; Roberts *et al.* 2001; Root, 2001; Trevett *et al.* 2005). The increased risk of children coming into contact with faecal bacteria where open defaecation or improper hygiene are practiced provides an ideal route for contamination of standpipe and community tanker-supplied water with *E. coli* and total coliforms in households including children aged 0-5 or >5-18 years.

The microbial deterioration of drinking water in households including children aged 0-5 and >5-18 years, as demonstrated in this study and others, can be reduced through the promotion of hand washing in children by their elders, use of safe areas free from faecal waste for recreational activities and the maintenance of point-of-use water quality by including taps and tightly fitting lids into the design of storage vessels and by placing storage vessels on higher surfaces thus preventing easy access for children to play with drinking water.

5.4. Links between microbial water quality, health outcomes, socio-economic status, household demographics and hygiene practices.

Microbial contamination of drinking water has been shown, in this study, to occur at the point-of-use for all water supplies studied and at the source for communal water supplies

(Figure 11). The consumption of microbially contaminated drinking water has been associated with increased rates in health outcomes such as diarrhoea and vomiting in studies by Trevett *et al.* (2004) and Clasen *et al.* (2005). The present study showed that whilst vomiting may be linked to high *E. coli* counts in households using open-top containers for water storage, the high rate of diarrhoea, as experienced by members dwelling in households supplied drinking water by ground tanks was not linked to microbial content of water (Figure 15 and Figure 17). Therefore there are factors other than microbial quality of drinking water which contribute to the rate of diarrhoea.

Provision of ground tanks, UD toilets (which incorporate a hand-washing area in the design of the unit to promote hand-washing after toilet use) and hygiene education as a package, as practiced by eThekwini Municipality to some low-income areas, was aimed at promoting proper and safe sanitation practices, hand washing and supply of microbially safe drinking water for users at the point-of-use. It was surprising to find in this study that, even though people were given this package, household members still practiced open defaecation (Table13). Urine Diversion toilets were not kept clean and hand-washing amongst members from ground tank-supplied households was lower than in members from households supplied by community tankers (Table 13, Table 14, Table 15 and Table 16).

Ground tank water showed significant microbial deterioration at the point-of-use, with heterotrophic organisms and total coliforms being present in counts that posed a potential risk of infection to users (Figure 11 and Table 6). Ground tank water however, was the only point-of-use water that had *E. coli* counts that posed a negligible risk of infection to users (Table 7). Members of ground tank-supplied households had the highest rates of diarrhoea, in comparison to standpipe and community tanker-supplied households (Figure 16). This was an unexpected finding since, standpipe and community tanker-supplied households had higher faecal bacteria counts in drinking water than ground tank-supplied households (Figure 11). This suggests that diarrhoea in this case was associated with factors other than drinking water.

The high rate of diarrhoea amongst ground tank-supplied household members could be due to direct self-infection with faecal bacteria via the hand to mouth route rather than through consumption of faecally contaminated drinking water since open-defaecation was a common practice, for ground tank and community tanker users, and hand-washing was minimal.

The low *E. coli* counts in ground tank drinking water could be explained by the presence of taps on ground tanks even in cases where open defaecation was practiced. Taps minimise the direct contact of drinking water with faecally contaminated hands or dipping utensils. The high total coliform counts could be as a result of contamination of drinking water with soil or leaf matter. During sampling it was observed that lids of several ground tanks were off, or loosely placed over tanks. Animals and children playing near ground tanks were also observed. These factors could account for increased total coliform counts.

The lack of relationship between rates of diarrhoea and microbial quality of drinking water, in ground tank-supplied households, could also be explained by low numbers of indicator organisms in drinking water samples collected as a result of sampling techniques used. During sampling ground tank taps were wiped down with 90% ethanol and taps were flushed for one minute before samples were collected, to ensure that water collected was not contaminated by the tap. This is not typical of water use practice by ground tank users. As a result of limited quantity of water that is available to household members, taps are not usually flushed before water is used. Several households were observed to have containers below ground tank taps to collect water that is lost by dripping taps or by water passing the initial collection vessel. When asked what this water was used for, responses included cooking, washing of dishes or nappies. Therefore little or no water from ground tanks is wasted or subject to loss through actions such as flushing of taps. Water that comes out of ground tank taps may be subjected to stagnation and hence potential growth of bacterial contaminants. Bacteria growing at this point may be responsible for some of the health effects observed in ground tank-supplied households which failed to be shown through laboratory analysis, probably as a result of loss of microorganisms through the flushing of ground tank taps. It can therefore be argued that an increase in the amount of water which is supplied to households may be accompanied by better hygiene practices and thereby reduce health outcomes caused by poor microbial quality of drinking water. If the amount of water supplied to households is increased, the washing of ground tanks may be more frequent. Rinsing of cups before drinking water and rinsing or flushing of taps before using water may even increase. Such flushing of taps will allow for stagnant water containing bacteria to be removed instead of being consumed.

The indicator organisms used to investigate diarrhoeal incidence and its association to drinking water microbial quality could also have contributed to the lack of relationship between rates of diarrhoea and microbial quality of ground tank-supplied drinking water. Studies by Meara (1989) showed that the presence of total coliform and *E. coli* bacteria in drinking water did not have a relationship to rates of diarrhoeal disease. However, the presence of faecal streptococci in the same water samples was associated with increased rates of diarrhoea.

Drinking water at point-of-use in households using open-top storage containers had the highest faecal contamination of water, in terms of *E. coli*, in comparison to drinking water in households using closed-top storage containers and ground tanks (Figure 14). Drinking water at point-of-use from households using open-top storage containers and including children aged 0-5 years had higher heterotrophic organisms, total coliform and *E. coli* counts than that in households including children aged >5-18 years and households comprising adults only (>18 years) (Figure 15). These households also presented with the highest rates of diarrhoea and vomiting (Figure 16). Here a link between microbial quality of drinking water, age distribution in households and rates of health outcomes is clearly shown. However, hygiene and sanitation practices are likely to have made contributions to microbial drinking water quality at point-of-use and rates of health outcomes.

The deterioration in microbial quality of drinking water at the source and point-of-use and its relation to hygiene practices has been demonstrated in studies by Echeverria *et al.* (1987), Pinfold, (1990), Hoque *et al.* (1995), Islam *et al.* (2001) and Trevett *et al.* (2005). Lack of proper hygiene practices, such as cleaning of drinking water storage vessels and dipping utensils used to remove drinking water from storage vessels and washing of hands, has been shown to contribute to decreased microbial quality of drinking water in this study as well as in studies by Feachem *et al.* (1983), Blum *et al.* (1990), Pinfold, (1990), Swerdlow *et al.* (1992), Hoque *et al.* (1995), Tuttle *et al.* (1995), Roberts *et al.* (2001), and Trevett *et al.* (2005).

The type of sanitation used *i.e.* use of UD toilets, VIP toilets or open defaecation, also contributes to poor microbial quality of drinking water and increase in health outcomes such as diarrhoea (VanDerslice and Briscoe, 1995; Kaltenthaler *et al.* 1996; Roberts *et al.* 2001; Root, 2001; Trevett *et al.* 2005; Trevett *et al.* 2008). In this study it was demonstrated that

households which practiced open defaecation had the highest levels of *E. coli* in drinking water at the point-of-use and the highest rates of diarrhoea and vomiting. Conversely households that kept UD toilets clean were associated with decreased microbial contamination of drinking water and decreased rates of health outcomes.

Studies have shown that when open defaecation is practiced people are less likely to wash their hands (Trevett *et al.* 2004). Even if they do, washing of hands would require the removal of water from storage vessels using dipping utensils or the touching of taps on delivery systems such as ground tanks. This demonstrates a route for transfer of faecal contaminants from contaminated hands to point-of-use drinking water (Swerdlow *et al.* 1992; Quick *et al.* 1999).

When open defaecation is practiced, the overall sanitary quality of the environment deteriorates (Trevett *et al.* 2005). This affects people at a domestic or community level depending on the following:

- location or area used for open-defaecation
- proximity of this area to households and
- use of this area for recreational activity such as playing in the case of children.

If children play in unsanitary environments used by members of several different households for defaecation, they are exposed to a variety of external pathogens (*i.e.* pathogens from people with whom they are not usually in contact with). This increases their risk of contracting faecal-orally transmitted diseases such as gastroenteritis (Feachem *et al.* 1983; VanDerslice and Briscoe, 1993). The risk of transmitting faecal-oral disease to other household members is also increased. If hand washing is not practiced, children may dip faecally contaminated hands or dipping utensils into drinking water when removing water and thus cause contamination of drinking water consumed by all household members (as discussed previously). The risk in this instance is that the possibility of contracting faecal-oral disease will be heightened since household members are less likely to be immune to these external pathogens and hence would be more susceptible to infection (Feachem *et al.* 1983; VanDerslice and Briscoe, 1993).

Households using open-top containers, for storage of drinking water, used open defaecation more commonly. Children played in these areas thereby providing a possible route for high faecal microbe content and associated increase in faecal-oral disease. Hand-washing with soap was more prevalent in households using open-top and closed-top containers for drinking water storage than households using ground tanks. This might be expected to have reduced microbial counts in point-of-use drinking water and rates of health outcomes. However, community tanker-supplied households had higher rates of vomiting than did ground tanksupplied households. This could be attributed to the actual process of hand-washing. In the case of households using containers devoid of a tap for water storage, water needed for washing of hands would be removed from storage vessels either by using a dipping utensil or through pouring out of water. Contact of faecally contaminated hands with dipping utensils or water can result in faecal contamination of drinking water which could then result in increased rates of associated health outcomes (as discussed previously). A study by Swerdlow et al. (1997) demonstrated a similar occurrence, in which refugees using containers for water storage reported that they either inserted their hands into water storage containers to remove water for washing hands or that they rinsed their hands directly in the container containing water used for drinking purposes.

The current study also investigated the impact of socio-economic status of household members on the microbial quality of drinking water and rates of health outcomes (diarrhoea and vomiting). It was demonstrated that households with a higher poverty index (measured by amount of income per household, with a household having a high poverty index being in a lower income bracket than households with low poverty index) and lower education level (measured by presence of books, newspapers or educational material) had higher heterotrophic organisms, total coliform and *E. coli* counts in their point-of-use drinking water than households with lower poverty index and higher education levels. Households of high poverty index and low socio-economic status also had the highest rates of diarrhoea and vomiting. Studies by Ettner, (1996) and Feinstein, (1999) have demonstrated that poor socio-economic status is associated with increased rates of health outcomes. Such studies show that people in low-income brackets spend most of their income on basic food, leaving them with little to spend on health care and vaccinations. As a result individuals are more likely to develop full blown illness than sub-clinical illness due to a lack of medical attention. This also results in increased mortality.

The impact of education and income on health outcomes has also been demonstrated by Becker *et al.* (1986), Manun'Ebo *et al.* (1994), Mahmud *et al.* (2001) and Trevett *et al.* (2005). Becker *et al.* (1986) showed that the duration for which diarrhoea lasts in children in low-income households was significantly longer than in higher income households. Manun'Ebo *et al.* (1994) found that a decrease in parental education was related to increased rates of diarrhoea amongst children aged 0-3 years. Illiteracy was associated with increased rates of persistent diarrhoea (Mahmud *et al.* 2001). Trevett, (2003) and Trevett *et al.* (2005) demonstrated that making use of knowledge of good hygiene practice is to some extent upon household income. He suggested that poor households had limited ability to improve or maintain sanitary environments.

The importance of providing low-income areas with basic facilities cannot be emphasised enough. Children, pregnant women, the elderly and immuno-compromised individuals are at the highest risk of infections caused by poor drinking water quality, poor handling of drinking water, improper sanitation and lack of hygiene (Eisenberg *et al.* 2001; Fewtrell *et al.* 2005; Trevett *et al.* 2005; Trevett *et al.* 2008). In the communities represented in this study children below the age of 5 years were observed to have the highest rates of diarrhoea and vomiting, compared with children aged >5-18 years and adults. The immune system of a child is still developing until the age of 5 years. It is only once the child is over 5 years of age that the immune system is fully functional. Therefore children under 5 years of age are more susceptible to infection (Guyton and Hall, 2006).

Contamination routes of drinking water stored in open-top containers by children has been demonstrated. Studies by Hammad and Dirar (1982), Deb *et al.* (1986), Empereur-Bissonnet *et al.* (1992) and Roberts *et al.* 2001 demonstrated that using improved drinking water storage containers, defined as containers which minimise hand-contact with drinking water and which offer protection from microbial contamination of drinking water through the presence of lids, results in a significant decrease in microbial contamination of drinking water storage containers (as defined above), resulted in a significant decrease of diarrhoeal disease in children less than 5 years of age. Studies by Yeager *et al.* (1991), Qadri *et al.* (1992) and Mahmud *et al.* (2001) demonstrated a significant association between incidences of diarrhoeal disease in children younger than 5 years of age and the type of drinking water storage container used. They showed that the use of drinking water storage vessels which were devoid of taps and hence

required dipping utensils for removal of water from storage vessels was significantly associated with increased rates of diarrhoeal disease in children younger than 5 years of age.

The current study and studies by Jagals *et al.* (1997) and Mirza *et al.* (1997) demonstrated that microbial contamination and rates of diarrhoeal disease were most prevalent in households including children younger than 5 years of age in which open-top containers were used for storage of drinking water at the point of use. Therefore the importance of good water handling practices, proper sanitation practices, maintenance of a sanitary environment, practice of proper hygiene and good microbial quality of drinking water are critical to the health of immuno-compromised people and children younger than 5 years of age due to their immunity status.

Overall this study demonstrated relationships between deterioration of microbial drinking water quality, type of sanitation practices, hygiene practices, socio-economic factors and rates of health outcomes (diarrhoea and vomiting). This is in accordance with studies by Esrey *et al.* (1985), Esrey *et al.* (1991), Curtis *et al.* (2000), Feachem *et al.* (1983), VanDerslice and Briscoe (1993) and Trevett *et al.* (2005), amongst others, which demonstrated that the reduction in health outcomes in low-income areas cannot be reduced only by providing safe drinking water at the point-of-use but rather can be most successfully achieved by providing proper sanitation, safe drinking water at the source and point-of-use and hygiene education in conjunction with each other.

5.5. Overall discussion

The current study, as well as those by Pinfold, (1990), Jones, (1999), Pinfold *et al.* (1991), Payment *et al.* (1991), Payment *et al.* (1993), Chidavaenzi *et al.* (1998), Trevett, (2003), Trevett *et al.* (2004), Clasen *et al.* (2005) Trevett *et al.* (2005) and Trevett *et al* (2008), all indicated that microbial quality of drinking water at the point-of-use is associated with increased incidences of health outcomes. However, Feachem *et al.* (1978) and VanDerslice and Briscoe (1993) claim that point-of-use water quality is not as important as the quality of source water and that point-of-use water does not represent a serious risk of faecal-oral disease. They claimed that prevention of microbial contamination of source water is more important and effective in controlling disease transmission than prevention of microbial contamination of drinking water at the point-of-use. Feachem states that contamination of drinking water at the point-of-use results in the presence of pathogens that would only affect household members who would be exposed to the same pathogens by other routes. He argued that source water contamination could introduce external pathogens in point-of-use drinking water and cause inter-family disease. VanDerslice and Briscoe (1993) added to his argument by suggesting that microbial contamination of point-of-use drinking water does not present a significant risk for diarrhoeal disease as household members develop immunity to pathogens spread among them. They also supported Feachem's argument that households in which faecal contamination of drinking water occurred at the point-of-use were unlikely to display an increase in incidences of diarrhoea since these household members would be at risk of diarrhoeal disease through other routes such as hand to mouth routes, anyway. Both studies concluded that if source drinking water was contaminated, pathogens from a variety of people or sources would be present. Water users would therefore be more susceptible and hence risk of infections at the point-of-use would be greater.

Whilst studies by Feachem et al. (1983) and VanDerslice and Briscoe (1993) highlight the importance of maintaining microbial quality of source water, other authors highlight the importance of maintaining microbial quality of point-of-use drinking water in order to reduce incidences of health outcomes. The current study, as also a collection of related studies (Maraj et al. 2005; Lutchminarayan, 2007), showed that households supplied by standpipes and community tankers had microbially contaminated drinking water at the point-of-use and source. These households also had higher incidences of overall health outcomes (diarrhoea and vomiting) when children younger than 5 years were included in households, as compared to households comprising adults only (occupants aged >18 years) and households including children aged >5-18 years. It was also demonstrated that ground tank source water was microbially safest for human consumption of the studied water supply routes with no faecal contamination at the source. However, faecal contamination at the point-of-use did occur and these ground tank supplied households had the highest incidences of diarrhoea in comparison to standpipe and community tanker supplied households. Therefore it is clear that the microbial quality of drinking water at both the source and point-of-use are of importance with regard to reducing health outcomes, and that both these water supplies should be treated with equal importance and rigor.

A study by Clasen *et al.* (2005) showed a relationship between the presence of coliforms in drinking water and incidences of diarrhoea, with an increase in coliform count correlated with

an increase in incidences of diarrhoea. This implied a dose-dependant relationship between microbial contamination of drinking water and diarrhoeal disease. On the other hand a systematic review by Gundry *et al.* (2004) found that studies in general failed to demonstrate a clear relationship between disease and microbial water quality at the point-of-use with regard to incidences of general diarrhoea. In the current study, microbial drinking water quality was shown to influence the health outcomes, diarrhoea and vomiting. However, it was demonstrated that microbial quality of drinking water at the point-of-use was not solely responsible for occurrence of such health outcomes and that other factors contribute to incidences of health outcomes studied.

Trevett *et al.* (2005) developed a conceptual framework of primary and secondary factors that determine potential pathogen loads in household stored drinking water and the final barrier that prevents disease. He states that the primary factors leading to deterioration in microbial drinking water quality at a household level are:

- handling of water, which includes collection, transport and storage of drinking water;
- hygiene practices represented by hand washing with soap; and
- environmental factors that include type of sanitation used, the sanitary quality of the environment and surroundings of communities and climate changes.

The secondary factors which contribute to further microbial deterioration of stored drinking water are:

- socio-economic factors, namely poverty index and education levels of household members;
- type and characteristics of pathogens that are present in drinking water; and
- anthropology of communities which includes cultural values and norms of societies and social interaction between families, neighbors and strangers.

The barrier that influences risk of disease, brought about by poor microbial quality of drinking water at the point-of-use, is health and immunity of household members. Trevett showed that if primary and secondary factors causing deterioration of microbial quality of drinking water are not managed or controlled, and if household members are unhealthy, younger than 5 years old or immuno-compromised, they are at increased risk of contracting

diseases associated with microbial contamination of drinking water (Trevett *et al.* 2005). The findings of the current study appeared to fit this conceptual framework as explained below.

It has been clearly demonstrated in this study, as well as in others, that water at the point-ofuse is subject to microbial deterioration relative to its source (Hammad and Dirar, 1982; Blum et al. 1990; Empereur-Bissonnet et al. 1992; Swerdlow et al. 1992; Kalthenthaler et al. 1996; Genthe et al. 1997; Hoque et al. 1999; Roberts et al. 2001; Clasen and Bastable, 2003; Trevett et al. 2004; Wright et al. 2004; Maraj et al. 2005; Trevett et al. 2005). Such microbial contamination has also been showed to lead to increased health outcomes such as diarrhoea and vomiting. In this study and others, microbial deterioration of point-of-use water has been shown to be related to handling practices (Yeager et al. 1991; Jagals et al. 1997; Mirza et al. 1997; Trevett, 2003; Trevett et al. 2005; Trevett et al. 2008), hygiene practices (Echeverria et al. 1987; Pinfold, 1990; Swerdlow et al. 1992; Hoque et al. 1995; Tuttle et al. 1995; Roberts et al. 2001; Trevett et al. 2004) and sanitation practices (VanDerslice and Briscoe, 1995; Kaltenthaler et al. 1996; Root, 2001; Trevett, 2003) Collection, transport and storage of drinking water in open-top containers, closed-top containers and ground tanks was associated with increased microbial content in drinking water at the point-of-use. Households which used open-top or closed-top containers for storage of drinking water and in which members practiced hand washing after defaecation had higher faecal contamination of drinking water compared to ground tank-supplied households. This was possibly as a result of dipping faecally contaminated hands or utensils into water storage containers to remove water required for hand washing. This was also demonstrated in a study by Swerdlow et al. (1992). When UD toilet seats were not clean and when open defaecation was used rather than UD or VIP toilets, faecal contamination of point-of-use water was increased and the incidences of diarrhoea were also increased. The increase in diarrhoeal disease as a result of open defaecation practices demonstrated in this study was also shown by Trevett et al. (2005), Kaltenthaler et al. (1996), VanDerslice and Briscoe (1993), Roberts et al. (2001) and Root, (2001). Therefore the primary factors (handling of drinking water, hygiene practices and sanitation practices) contributing to poor point-of-use microbial drinking water quality, as listed in Trevett's conceptual framework, have also been demonstrated to contribute to poor microbial quality of drinking water at the point-of-use in the current study.

Some secondary factors from Trevett's conceptual framework, such as socio-economic factors and pathogen characteristics have also been shown to contribute to microbial

contamination of drinking water in this study. It was found that households with higher poverty indices and lower educational indicators (as measured by presence of books and newspapers in households) had higher total coliform, heterotrophic organism and *E. coli* counts in drinking water at the point-of-use. Members of these households also had higher overall incidences of diarrhoea and vomiting in comparison to households with lower poverty indices and better educational indicators. The type of pathogen and its associated characteristics were also found to be of importance in determining disease risks. The occurrence of total coliforms in conjunction with *E. coli* suggests that total coliforms were from human faecal origin and that other gastrointestinal pathogenic microorganisms may have been present in drinking water (Payment *et al.* 1993).

Age distribution in households also influenced microbial content of drinking water at the point-of-use, with households comprising children aged 0-5 or >5-18 years having the highest faecal contamination of drinking water. Households that stored drinking water in open-top storage containers and which included children aged 0-5 years had the highest *E. coli* counts and highest incidences of diarrhoea and vomiting. This relationship between age distribution, microbial water quality and occurrences of health outcomes is consistent with Trevett's conceptual framework which states that households comprising children aged 0-5 years or immuno-compromised individuals are at a higher risk of microbial infection through consumption of drinking water containing pathogenic microorganisms.

In summary, this study has shown that primary factors, namely handling of drinking water via collection, transport and storage, washing of hands as a measure of hygiene practices and sanitation practices, are major contributors to the deterioration of point-of-use drinking water and increased incidences of health outcomes. However secondary factors such as socio-economic factors and type of pathogens present, also contribute to further deterioration of microbial quality of drinking water at the point-of-use. Occurrences of health outcomes are not solely caused by microbial contamination of point-of-use drinking water and cannot be addressed or corrected only by improving drinking water quality at the point-of-use. Instead, occurrences of health outcomes have been shown to be linked to deterioration of drinking water as a result of storage and transport of drinking water, hygiene practices and sanitation practices. Therefore in order to reduce incidences of diarrhoea and vomiting, all primary factors (storage of drinking water, sanitation and hygiene practices) need to be targeted in a co-coordinated manner.

CHAPTER 6: CONCLUSIONS

6.1 Specific Conclusion

- Microbial deterioration of drinking water, supplied by improved sources, does occur at the point-of-use relative to the improved water sources, as supplied by eThekwini Municipality, for ground tanks and community tankers.
- Only ground tank source provided microbially safe drinking water to the corresponding point-of-use, whilst source water from community tankers and standpipes was unsafe for human consumption and resulted in the corresponding point-of-use water also being unsafe for human consumption.
- All point-of-use water was found to be unsafe for human consumption.
- Ground tank point-of-use water deteriorated significantly relative to its source with regard to microbial content. Ground tank point-of-use water was rendered unsafe for human consumption due to high overall microbial counts.
- With regard to standpipe-supplied households, there was no significant increase in microbial counts between standpipe source water and the corresponding point-of-use water samples. Therefore poor microbial quality of drinking water sampled from open-top and closed-top containers can be attributed to the poor source water quality. Since sterile techniques were used to sample drinking water from the standpipe, it is possible that the source water was contaminated at some point in the distribution system or within the mechanism of the standpipe.
- Point-of-use water from community tankers deteriorated even further relative to its source. These results confirm findings in other studies, which also suggest that water quality deteriorates relative to its source at the point-of-use (Clasen and Bastable, 2003; Trevett *et al.* 2004; Wright *et al.* 2004; Maraj *et al.* 2005; Trevett *et al.* 2005).

- Drinking water samples from closed-top containers supplied by community tankers displayed no further microbial deterioration relative to its source water (community tankers). This implies that using a closed-top container for water storage does decrease the chances of further microbial deterioration at the point-of-use. This is further supported by the findings of microbial water quality from open-top containers supplied by community tankers. In this instance there was a significant increase in microbial content in the point-of-use water samples when compared to the source water samples. Thereby implying that water in open-top storage containers are more prone to microbial contamination than water in closed-top containers.
- Findings on water stored in open-top and closed-top containers, supplied by communal sources, therefore show that the microbial quality of point-of-use water is affected not only by the type of storage vessel used to store drinking water at the point-of-use, but is also affected by the microbial quality of the source water. Therefore treatment of drinking water at the point-of-use should be considered in order to ensure that water reaching users homes are free of external pathogens. The point of contamination of source water needs to be identified through further studies.
- The presence of children aged 0-5 or >5-18 years in households was associated with higher, (but not significantly higher), faecal contamination of point-of-use drinking water in comparison to households comprising adults only (>18 years). This is an indication of two possibilities. It could be an indication that the presence of children in a household could be linked to increased faecal contamination of drinking water as a result of poor hygiene and sanitation practices. Alternatively it is a possibility that the faecal contamination noted at the point-of-use was partially contributed to by poor source water quality as shown in the case of standpipes and community tankers. This would explain why water from households comprising adults only also displays faecal contaminants.
- When using open-top containers for water storage in households including children as occupants, the rate of diarrhoea was higher in comparison to households using closed-top containers or ground tanks. These households also had the highest *E. coli* counts therefore; microbial quality of drinking water, demographic distribution and health outcomes can be linked.

- When demographic distribution is not taken into consideration, occupants in households supplied drinking water by ground tanks had the highest rate of diarrhoea in comparison to households receiving water at the point-of-use in open-top or closed-top storage containers. Since ground tank water was of a better overall microbial quality than water from open-top and closed-top containers, this was unexpected. Such a finding suggests that occurrences of health outcomes are linked to factors other than microbial quality of drinking water. In this case, high occurrences of health outcomes amongst users of ground tank supplied water could be due to hygiene and sanitation practices, since people in ground tank-supplied households were found to use open defaecation even though UD toilets were provided and they also did not practice hand washing after using the toilet.
- The practice of open defaecation, lack of cleanliness of UD toilets and the lack of hand wash practice all contribute to increase in microbial content in drinking water from ground tanks and open-top and closed-top containers supplied by community tankers. A lower level of education and higher poverty indices were also shown to cause increase in microbial content of drinking water.
- Poor sanitation practices, low socio-economic status, demographic distribution, and poor personal hygiene collectively contributed to overall deterioration of microbial quality of drinking water at the point-of-use, and to increased rates of health outcomes such as diarrhoea and vomiting. This complex interaction of factors shows that improving water quality in isolation is unlikely to improve health outcomes significantly.
- It has been shown that the relationship between point-of-use water quality and health effects is not as direct or simple as postulated in the development of the project objectives. This is clear from results of the present study, when analysed in conjunction with results from the epidemiological study by Lutchminarayan (2007).

6.2 Overall Conclusion

The microbial quality of point-of-use drinking water supplied by ground tanks, standpipes and community tankers was shown to be unsafe for human consumption. In this study this has been shown to be the result of the following:

- Poor source water quality with regard to community tankers and standpipes.
- The collection, transport and storage of drinking water in open top containers and other easily accessible storage containers.
- Poor personal hygiene and sanitation practices, improper water-use practices, low socio-economic status and low levels of education also contributed to the microbial deterioration of drinking water at the point-of-use.
- The presence of children aged 0-5 and >5-18 years in households storing drinking water in ground tanks, open-top or closed-top storage containers was associated with increased faecal contamination of drinking water.

The microbial deterioration of point-of-use drinking water, and poor hygiene and sanitation practices, resulted in increased incidences of health outcomes such as diarrhoea and vomiting in households including children younger than 5 years of age and in which open-top containers are used for water storage.

In order to improve the microbial quality of point-of-use drinking water and reduce the incidences of health outcomes, a holistic approach needs to be taken. This approach should include collectively the provision of adequate quantities of water which is safe for human consumption together with proper sanitation, hygiene and water-use behaviour. This will allow for provision of microbially safe point-of-use drinking water and probably a subsequent reduction in incidences of water-associated health outcomes.

6.3 Operational recommendations

Operational recommendations for improvement and maintenance of microbial quality of drinking water at the point-of-use and source are given below.

Maintenance of good microbial drinking water quality at the source (standpipes and community tankers) and point-of-use (ground tanks, open-top containers and closed-top containers) can be achieved through:

- Increase in disinfectant level during treatment in order to increase residual disinfectant levels so that water stays pathogen free for longer and more bacteria at the source are targeted and destroyed.
- Point-of-use water treatment. Water delivery systems such as community tankers and ground tanks, requires that water be stored in the tanker until collection and subsequent use. These vessels are subject to microorganism regrowth and biofilm development on surfaces due to their difficulty to be cleaned as a result of their inaccessible nature. Therefore disinfectant use at the point-of-use, within recommended guideline values, is suggested to target bacteria that may be present or regrow in these tanks.
- Thorough and regular cleaning of community tankers, ground tanks and storage vessels is also recommended to remove particulate matter and microorganisms that may be present on inner surfaces as a result of biofilm formation or microorganism regrowth.
- Visits to communal water source sites and water dispensing sites of communal tankers by municipal workers should be implemented to inspect surrounding environments of water sources.

Maintenance and provision of microbially safe drinking water at the point-of-use and hence reduction in incidences of health outcomes in children younger than 5 years of age can be achieved through:

- Promoting use of closed-top containers rather than open-top containers for the storage of drinking water. Household members should be encouraged to pour out drinking water from storage vessels rather than dip utensils or hands into vessels to remove water. If this is not feasible, promotion of the use of a designated dipping utensil, preferably with a long handle to prevent direct hand-to-water contact should be considered. Household members should also be made aware of the importance of cleaning dipping utensils. Treatment of drinking water in storage vessels at the point-of-use through either chlorination or the introduction of interventions such as ceramic household filters as shown by Clasen *et al.* (2005).
- Promotion of proper handling of drinking water, hygiene and sanitation practices collectively, in low-income areas, by further targeting education provision could present a possible route to improve poor microbial quality of water at the source and point-of-use. Health and hygiene education needs to be broadened to include certain elements which have arisen from this study, such as a lack of hand washing and practice of open defaecation.
- A possible recommendation to broaden hygiene education to include alleviation of poor education and socio-economic status in communities could be to directly involve community members in the delivery of education workshops.

During education of community members, emphasis should be placed on the following:

- Hand washing after using toilets and before cooking and eating to prevent microbial contamination of drinking water and food and subsequent transmission of waterborne disease.
- Washing of storage containers and ground tanks on a regular basis to avoid contamination of drinking water by microbes present in storage vessels either as a result of regrowth or from biofilms on the walls of the container.
- Keeping lids of ground tanks tightly in place to minimise the introduction of foreign particles and microbes into water.

- Avoiding the use of open-top storage containers for water collection, transport and storage because of their association with high microbe content and disease occurrences in users.
- Placing storage containers on high surfaces, away from animals and small children to prevent contamination of drinking water.
- Encouraging users to pour water out of storage containers when required for use rather than to dip hands or utensils into storage containers since microbial contamination of hands and utensils could result in contamination of drinking water.
- Emphasising the need for cleanliness of environments surrounding standpipes and community tankers.
- Flushing of taps for a few seconds should be encouraged to allow microbes that may have grown in stagnant water at the tap to be removed instead of consumed. This however would work best if the quantity of water supplied to households is increased.
- Educating communities on the dangers associated with open defaecation in areas where children play should be a priority as these areas represent unsanitary environments. People who have access to UD or VIP toilets but do not use these toilets should be interviewed to determine reasons why they prefer not to use toilets provided.

Lastly, the determination of the exact point at which microbial contamination of water supplied by standpipes and community tankers occurs needs to be located and dealt with. If microbially contaminated at the source then point-of-use water will also be contaminated and good hygiene and sanitation practices at the point-of-use will be of limited use in the deterioration of the microbial quality of drinking water.

6.4 Research recommendations

- Investigation of the microbial quality of drinking water at the treatment plant, various points in the distribution and delivery system, and a representative number of samples from standpipes and community tankers, supplied by these distribution routes, is recommended in order to determine the exact point of microbial contamination of drinking water and hence to correct it accordingly. This would be particularly useful to explain why water from standpipes and community tankers which are supplied directly from the municipal treatment plant was of such poor microbial quality.
- Future studies considering the use of other indicator microorganisms or possible use of molecular techniques to identify a broader range of microbial contaminants in drinking water that could contribute to incidences of health outcomes would be beneficial.
- A study involving focus groups of community members which allow them to bring their problems relating to water uses to the fore, and indicating which they experience as the most effective means of community health and hygiene education. This study could also be used to investigate why some communities are not using toilets which are provided to them and also to gain insight as to why health and hygiene education appears to be failing in some areas.
- Investigation of the presence of biofilm and its constituents, in ground tanks, water storage vessels and points along the distribution system, using molecular techniques such as fluorescent *in situ* hybridization and polymerase chain reaction is also recommended. This would be helpful in isolating and identifying an array of biofilm constituents and hence determining the contamination risks associated with the presence of biofilm constituents in drinking water. This study could be used to determine methods to reduce occurrences of biofilm. Such a study is of particular importance in rural and peri-urban areas since the rate of HIV/AIDS is highest in these areas and the implications of biofilm constituents on immune-compromised individuals in terms of opportunistic infections is yet to be conducted. Such areas will be using ground tanks with automatic bailiff systems to a greater extent in the coming years, hence a study of this nature is viewed as being beneficial.

References

Arjun N., Joubert J.P., Rodda N., Smith M and Buckley C., 2004. Water quality at point-ofuse for users supplied by standpipes and ground tanks in a peri-urban community. Paper presented at IWA Specialist Group Conference on Water and Wastewater Management for Developing Countries, Victoria Falls, Zimbabwe.

Allen J.M., Brecher R.W., Copes R., Hrudey S.E., Payment P., 2008. Turbidity and microbial risk in drinking water. Prepared for the Minister of Health, Province of British Columbia. Pursuant to section 5 of the Drinking Water Act (S.B.C. 2001).

American Society for Microbiology, 2002. Water-related health illness. <u>http://www.microbeworld.org/html/cissues/water/wqual_1htm</u> (accessed: 24-05-06)

APHA, WERF, AWWA, 2005. Standard methods for the examination of water and wastewater. 21st edition. Washington DC.

Bacchieri A., Cioppa G., 2007. Statistics for biology and health, Experimental design: the randomised blinded study as an instrument to reduce bias. Springer Milan. 200-227.

Bailey I.W., and Archer L., 2004. The impact of introducing treated water on aspects of community health in a rural community in KwaZulu-Natal. Water Science and Technology. **50** (1): 105-110.

Baqui A.H., Black R.E., Yunus M., Hoque A.R., Chowdhury H.R., Sack R.B., 1991. Methodological issues in diarrhoeal diseases epidemiology: definition of diarrhoeal episodes. International Journal of Epidemiology. **20**: 1057-1063.

Becker S., Black R.E., Brown K.H., Nahar S., 1986. Relations between socio-economic status and morbidity, food intake and growth in young children in Bangladesh. Ecology Food and Nutrition. **18**: 251–264.

Block J.C., Haudidier K., Paquin J.L., Miazga J., and Levi Y., 1993. Biofilm accumulation in drinking water distribution systems. Biofouling. **6**: 333-343.

Blum D., Emeh R.N., Huttly S.R.A., Dosunmu-Ogunbi O., Okeke N., Ajala M., Okoro J.I., Akujobi C., Kirkwood B.R., Feachem R.G., 1990. The Imo State (Nigeria) drinking water supply and sanitation project, 1. Description of the project, evaluation methods, and impact on intervening variables. Tropical Medicine and Hygiene. **84**: 309–315.

Bradley D., 1977. Health aspects of water supplies in tropical countries. In Feachem R, McGarry M and Mara D, Water Wastes and Health in hot climates. John Wiley and Sons. Chichester, UK. 3-17.

Bradshaw D., Groenewald P., Laubscher R., Nannan N., Nojilana B., Norman R., Pieterse D., Schneider M., 2003. Initial Burden of Disease Estimates for South Africa. Cape Town: South African Medical Research Council.

Brocklehurst C., 2005. Durban Metro Water: private sector partnerships to serve the poor. Water and sanitation program – Africa. <u>http://www.wsp.org/publications/af_durban.pdf</u> (Accessed 10 March 2005).

Buckley C.A., Foxon K.M., Rodda N., Nwaneri C., Balboni E., Couderc A., Magagna D., 2008. Scientific support for the design and operation of Ventilated Improved Pit (VIP) latrines and efficacy of pit latrine additives. Water Research Commission. Pretoria, report no. TT357/08.

Carden K., Armitage N., Winter K., Sichone O., Rivett U., 2006. Understanding the use and disposal of greywater in the non-sewered areas in South Africa. Final report WRC Projects K5/1524, Pretoria, South Africa.

Camper A. K., 1994. Coliform regrowth and biofilm accumulation in drinking water systems: a review, *In* G. G. Geesey, Z. Lewandowski, and H.-C. Flemming (ed.), Biofouling and Biocorrosion in Industrial Water Systems. Lewis Publishers, Boca Raton. P91-P105. Centre for Disease Control, 2005. Typhoid fever. www.cdc.gov (Accessed 07 July 2008).

Centre for Disease Control, 2008a. *Escherichia Coli*. <u>www.cdc.gov</u> (Accessed 03 March 2009).

Centre for Disease Control, 2008b. Cholera. www.cdc.gov (Accessed 07 July 2008).

Centre for Disease Control, 2008c. Shigellosis. www.cdc.gov (Accessed 07 July 2008).

Centre for Disease Control, 2008d. Giardiasis (Giardia infection). www.cdc.gov (Accessed 03 March 2009).

Centre for Disease Control, 2008e. Cryptosporidiosis. www.cdc.gov (Accessed 07 July 2008).

Centre for Disease Control, 2008f. Ascaris Infection. www.cdc.gov (Accessed 03 March 2009).

Chanlett E.T., 1992. Environnent protection. MacGraw-Hill Book Company. New York.

Chidavaenzi M.T., Jere M., Nhandara C., Chingundury D., Bradley M., 1998. An evaluation of water urns to maintain domestic water quality. Paper presented at the 24th WEDC conference, Islamabad, Pakistan.

Clasen T.F., Bastable A., 2003. Faecal contamination of drinking water during collection and household storage: the need to extend protection to the point-of-use. Journal of Water and Health. **1** (2): 1-7.

Clasen T., Roberts I., Rabie T., Schmidt W., Cairneross S., 2005. Interventions to improve water quality for preventing infectious diarrhoea. The Cochrane Library, Issue 4. Oxford: Update Software.

Costerton J. W., Lewandowski Z., de Beer D., Caldwell D., Korber D., and James G., 1994. Biofilms, the customised microniche. Journal of Bacteriology. **176**: 2137-2142.

Corrigan J.J. Jr., Boineau F.G., (2001). "Hemolytic-uremic syndrome". Paediatric Review 22 (11): 365–9.

Coulson N., 2000. Breaking the rules: New approaches to promoting health through water and sanitation in South Africa. Soul City, The Institute for Health and development Communication.

Crompton D.W.T., 1988. The prevalence of ascariasis. Parasitology Today. 4: 162-8.

Curtis V., Cairneross S., Yonli R., 2000. Domestic hygiene and diarrhoea: pinpointing the problem. Tropical Medicine and Health. **5** (1): 22–32.

Davies D.G., Mcfeters A., 1988. Growth and comparative physiology of *Klebsiella oxytoca* attached to granular carbon particles and in liquid media. Microbial Ecology.. **15**: 165-175.

Deb B.C., Sircar B.K., Sengupta P.G., De S.P., Mondal S.K., Gupta D.N., Saha N.C., Mitra G.U., Pal S.C., 1986. Studies on interventions to prevent eltor cholera transmission in urban slums. Bulletin World Health Organisation. **64:** 127–131.

Department of Water Affairs and Forestry (DWAF), 1994. White paper on Water and Sanitation. Department of Water Affairs and Forestry, Pretoria, South Africa.

Department of Water Affairs and Forestry (DWAF), 1996. South African Water Quality Guidelines. Vol 1 (Domestic Use), Pretoria.

Department of Water Affairs and Forestry (DWAF), 2001. Informational Kit for the Free Basic Water Implementation in South Africa. Department of Water Affairs and Forestry, Pretoria, South Africa.

Department of Water Affairs and Forestry (DWAF), 2002. Free basic water: tap into life, regulations and guidelines, guidelines for compulsory national standards and norms and

standards for water services tariffs and water services provider contract regulations. Department of Water Affairs and Forestry, Pretoria, South Africa.

Department of Water Affairs and Forestry (DWAF), 2003. Water is Life, Sanitation is Dignity: Strategic Framework for Water Services. Department of Water Affairs and Forestry, Pretoria, South Africa.

Department of Water Affairs and Forestry (DWAF), 2005. Drinking water quality management guide for water services authorities. Department of Water Affairs and Forestry, Pretoria, South Africa.

Department of Water Affairs and Forestry (DWAF), 2008. Water services national information system. <u>www.dwaf.gov.za</u> (Accessed 2 April 2009).

Donlan R.M., Pipes M.O., 1988. Selected drinking water characteristics and attached microbial population density. Journal of American Water Works Association. 80: 70 - 76.

Donlan R.M., Pipes W.O., Yohe T.L., 1994. Biofilm formation on cast iron substrata in water distribution system. Water Research. **28**: 1497 – 1503.

Durban Metro Water, 2002. Affordable Water Supply to the Poor. eThekwini Water Services, Durban.

Eaton A.D., Clesceri L.S., Greenberg A.E., 1995. Standard Methods for the examination of water and wastewater. American Public Health Association, Washington DC. 19th addition.

Echeverria P., Taylor D.N., Seriwatana J., Leksomboon U., Chaicumpa W., Tirapat C., Rowe B., 1987. Potential sources of enterotoxigenic *Escherichia coli* in homes of children with diarrhoea in Thailand. Bulletin World Health Organisation. **65**: 207–215.

Edberg S.C., Rice E.W., Karlin R.J., Allen M.J., 2000. *Escherichia Coli*: the best biological drinking water indicator for public health protection. Journal of Applied Microbiology. **88**: 106-116.

Eisenberg J.N.S., Bartram J., Hunter P.R., 2001. A public health perspective for establishing water-related guidelines and standards. In: Water Quality: Guidelines, standards and health. Assessment of risk and risk management for water-related infectious disease (edited, L. Fewtrell and J. Bartram), International Water Association Publishing, London. P229–P256.

Empereur-Bissonnet P., Salzman V., Monjour L., 1992. Application d'un nouveau mate' riel de transport et de stockage pour l'ame' lioration de la qualite' de l'eau de boisson en milieu rural africain. Bulletin Social Pathology. **85**: 390–394.

Esener A.A., Roels J.A., Kosen N.W.F., 1981. The influence of temperature on the maximum specific growth rate of *Klebsiella pneumoniae*. Biotechnology and Bioengineering. **23**: 1401 – 1405.

Esrey S.A., Feachem R.G., Hughes J.M., 1985. Interventions for the control of diarrhoeal diseases among young children: improving water supplies and excreta disposal facilities. Bulletin World Health Organisation. **64:** 776-72.

Esrey S.A., Potash J.B., Roberts L., Shiffman C., 1991. Effects of improved water supply and sanitation on ascariasis, diarrhoea, dracunculiasis, hookworm infection, schistosomiasis, and trachoma. Bulletin World Health Organisation. **69**: 609-21.

Esrey S.A., Gough J., Rapaport D., Sawyer R., Simpson-Herbert M., Vargas J., Winblad U., 1998. Ecological sanitation. Swedish International Development Cooperation Agency, Stockholm, Sweden.

eThekwini Municipality Test Method No MM002, 2004. Analysis of total coliform and *E. coli* on chromocult medium using the membrane filter procedure, eThekwini Water Services Waste Water Laboratory.

eThekwini Municipality Test Method No. MM007, 2004. Heterotrophic plate count, eThekwini Water Services Waste Water Laboratory.

eThekwini Municipality. 2005. Provision of potable water. http://www.durban.gov.za (Accessed 25 May 2006). Ettner S.L., 1996. New evidence on the relationship between income and health. Journal of Health Economics. **15**: 67-85.

Feachem R.G., Burns E., Cairneross S., Cronin A., Cross P., Curtis D., Khalid Khan M., Lamb D., Southall H., 1978. Water, Health and Development: An Interdisciplinary Evaluation. Tri-med books Ltd., London.

Feachem R.G., Bradley D.J., Garelick H., Mara D.D., 1983. Sanitation and disease: Health aspects of excreta and wastewater management. World Bank studies in water supply and sanitation 3. John Wiley & Sons, New York.

Feinstein J.S., 1999. The relationship between socio-economic status and health: A review of the literature. The Millbank Quarterly. **71** (2): 279-322.

Feng P., Weagent S., Grant M., 2002. Enumeration of *Escherichia Coli* and coliform bacteria. Bacteriological Analytical manual 8th edition. Centre for Food Safety and Applied Nutrition.

Http://www.csfan.fda.gov. (Accessed 27 January 2007).

Fewtrell L., Kaufmann R., Kay D., Enanoria W., Haller L., Colford J., 2005. Water, sanitation, and hygiene interventions to reduce diarrhoea in less developed countries: a systematic review and meta-analysis. Lancet Infectious Diseases. **5**: 42-52.

Ford T.E., 1999. Microbiological safety of drinking water: United States and global perspectives. Environmental Health Perspectives. **107** (S1): 191-206.

Freund J.F., Moore D.S., McCabe G.P., 1988. Introduction to the practice of statistics. Modern elementary statistics. D.H. Sanders. 21:25.

Genthe B., Strauss N., Seager J., Vundule C., Maforah F., Kfir R., 1997. The effect of type of water supply on water quality in a developing community in South Africa. Water Science and Technology. **35**: 35–40.

Gorchev H.G., 1996. Chlorine in water disinfection. Pure and Applied Chemistry. **68** (9): 1731-1735.

Guffanti A.A., Mann M., Sherman T.L., and Krulwhich T.A., 1984. Patterns of electrochemical proton gradient formation by membrane vesicles from an obligatory acidophilic bacterium. Journal of Bacteriology. **159**: 448 – 452.

Gundry S.W., Wright J., Conroy R., 2004. A systematic review of the health outcomes related to household water quality in developing countries. Journal of Water Health. **2**: 1–13.

Gundry S.W., Wright J.A., Conroy R., Du Preez M., Genthe B., Moyo S., Ndamba J., Potgieter N., 2006. Contamination of drinking water between source and point-of-use in rural households of South Africa and Zimbabwe: implications for monitoring the Millennium Development Goal for Water. Water Practice and Technology. **1** (2): 1-9.

Guyton A.C., Hall J.E., 2006. Textbook of Medical Physiology. Elsevier publishing. Eleventh edition: 439-441.

Hammad Z.H., Dirar H.A., 1982. Microbiological examination of 'sebeel' water. Applied Environmental Microbiology. **43**: 1238–1243.

Heinanen A., Chandiwana S.K., Makura O., Chimbari M., Bradley M., 1988. Faecal contamination of rural drinking water in a commercial farming area in Zimbabwe. Central African Journal of Medicine. **34**: 253-259.

Henley M., 1995. USP 23 changes include addition of TOC and conductivity testing. Ultrapure Water. **12** (6): 14-18.

Hoque B.A., Mahalanabis D., Pelto B., Alam M.J., 1995. Research methodology for developing efficient hand washing options: an example from Bangladesh. Journal of Tropical Medicine and Hygiene. **98:** 469–475.

Hoque B.A., Chakraborty J., Chowdhury J.T.A., Chowdhury U.K., Ali M., El Arifeen S., Sack R.B., 1999. Effects of environmental factors on child survival in Bangladesh: a case control study. Public Health. **113**, 57-64.

Islam M.S., Hossain M.A., Khan S.I., Khan M.N.H., Sack R.B., Albert M.J., Huq A., Colwell R.R., 2001. Survival of Shigella dysenteriae Type 1 on fomites. Journal of Health and Population Nutrition. **19** (3): 177–182.

Jagals P., Grabow W.O.K., Williams E., 1997. The effects of supplied water quality on human health in an urban development with limited basic subsistence facilities. Water South Africa. **4**: 372-379.

Jagals P., Bokako T.C., Grabow W.O.K., 1999. Changing consumer water use patterns and their effect on microbiological water quality as a result of an engineering intervention. Water South Africa. **25** (3): 297-300.

Jagals P., Nala N.P., Tsubane T.J., Moabi M., Motaung K.C., 2004. Measuring changes in water-related and hygiene practices by developing community households. Water Science and Technology. **50** (1): 91-97.

Jones D.L., 1999. Potential health risks associated with the persistence of *Escherichia coli* O157 in agricultural environments. Soil Use Management. **15**: 76-83.

Kaltenthaler E.C., Drasar B.S., Potter C.W., 1996. The use of microbiology in the study of hygiene behaviour. Microbiology. **88**: 35–43.

Kaptchuk T.J., 2000. The double-blind, randomised, placebo-controlled trial: Gold standard or golden calf. Journal of Clinical Epidemiology. **54**: 541-549.

Kay D., Bartram J., Pruss A., Ashbolt N., Wyer M.D., Fleisher J.M., Fewtrell L., Rogers A., Rees G., 2003. Derivation of numerical values for World Health Organisation guidelines for recreational waters. Water research. **38**: 1296-1304.

Khairy A.E.M., El Sebaie O., Gawad A.A., El Attar L., 1982. The sanitary condition of rural drinking water in a Nile Delta village. A parasitological assessment of 'zir' stored and direct tap water. Journal of hygiene. Cambridge. **88**: 57-61.

KwaZulu-Natal Municipalities. Drinking water supply in Durban, methods of delivery. www.kznmunicipalities.co.za (Accessed 13 April 2008).

Langmark J., Swedish institute for infectious disease control. Department of water microbiology. Jonas.langmark@smi.ki.se.

LeChevallier M.W., Seidler R.J., Evans T.M., 1980. Enumeration and characterization of standard plate count bacteria in chlorinated and raw water supplies. Applied and Environmental Microbiology. **40**: 922-930.

LeChevallier M.W., Babcock T.M., Lee R.G., 1987. Examination and characterization of distribution system biofilms. Applied and Environmental Microbiology. **53**: 2714-2724.

LeChevallier M.W., Lowry C.D., Lee R.G., 1990. Disinfecting biofilms in a model distribution system. Applied and Environmental Microbiology. **82**: 87-99.

LeChevallier M.W., Schultz W., Lee R.G., 1991. Bacterial nutrients in drinking water. Applied Environmental Microbiology. **57**: 857-862.

LeChevellier M.W., Welch N.J., Smith D.B., 1996. Full scale studies of factors related to coliform regrowth in drinking water. Applied Environmental Microbiology. **62**: 2201 – 2211.

Lehohla P., 2007. Community survey shows progress on millennium goals. Statistics South Africa. 08 November 2007.

Lehtola M.J., Talis J., Miettinen I.M., Vartiainen T., Martikainen P.J., 2004. Formation of biofilms in drinking water distribution networks, a case study in two cities in Finland and Latvia. Journal of Microbiology and Biotechnology. **31**: 489-494.

Lund V., Omerod K., 1995. The influence of disinfection processes on biofilm formation in water distribution systems. Water Research. **29**: 1013 – 1021.

Lutchminarayan R.D., 2007. A comparative study evaluating the health impacts of ecological sanitation interventions, water services and hygiene education programmes individually and in combinations, in eThekwini district, Durban, South Africa. Masters dissertation, University of KwaZulu-Natal, Durban, passed.

Mahmud M.A., Hossain M.M., Huang D.B., Habib M., DuPont H.L., 2001. Sociodemographic, environmental and clinical risk factors for developing persistent diarrhoea among infants in a rural community of Egypt. Journal of Health Population and Nutrition. **19** (4): 313–319.

Manun'Ebo M.N., Haggerty P.A., Kalengaie M., Ashworth A., Kirkwood B.R., 1994. Influence of demographic, socio-economic and environmental variables on childhood diarrhoea in a rural area of Zaire. Journal of Tropical Medicine and Hygiene. **97**: 31–38.

Mara D., 2006. Appropriate technologies. National Water Summit, South Africa.

Maraj S; Rodda N., Jackson S., Buckley C., Macleod N., 2005. Microbial deterioration of stored water for users supplied by standpipes and ground tanks in a peri-urban community. Water South Africa. 1: 222-230.

Mayo A.W., Noike T., 1994. Response of mixed culture of *Chlorella vulgaricus* and heterotrophic organisms to variation of pH. Water Science and Technology. **30** (8/9): 283 – 289.

McCoy W.F., Olson B.H., 1986. Relationship among turbidity, particle counts and bacteriological quality within water distribution lines. Water Resources. **20**: 1023-1029.

McFeters G.A., 1990. Drinking water microbiology. Springer-verlag, New York.

Meara J.L., 1989. An investigation of health and lifestyle in people who have private water supplies at home. Community Medicine. **11**: 131-139.

Medical Research Council., 1999. Household environment and health in Port Elizabeth, South Africa. Executive summary. Urban Health and Development Bulletin. **2** (4): 44-45.

Miettinen I.T., Vartiainen T., Martikainen P.J., 1997. Phosphorous and bacterial growth in drinking water. Applied Environmental Microbiology. **63**: 3242-3245.

Mittleman M.W., 1995. Biofilm development in purified water systems. Microbial Biofilms. 133-141.

Mirza N.M., Caulfield L.E., Black R.E., Macharia W.M., 1997. Risk factors for diarrhoeal duration. American Journal of Epidemiology. **146** (9): 776–785.

Momba M.N.B., Mnquemvu B.M., 2000. Detection of faecal coliform bacteria and heterotrophic plate count bacteria attached to household containers during the storage of drinking groundwater in a rural community. WISA Biennial Conference and Exhibition, Sun City.

Momba M.N.B., Kaleni P., 2002. Regrowth and survival of indicator microorganisms on the surfaces of household containers used for the storage of drinking water in rural communities of South Africa. Water Research. **36**: 3023-3028.

Momba M.N.B., Notshe T.L., 2003. The microbiological quality of groundwater derived drinking water after long storage in household containers in a rural community of South Africa. Journal of Water Supply: Research and Technology AQUA. **52** (1): 66-67.

Morgan P., 1999. Ecological sanitation in Zimbabwe – A compilation of manuals and experiences. Conlon Printers, Harare Zimbabwe.

Morris J.C., 1982. Health perspective in the oxidative treatment of water for potable supply. National institute for Water Supply. Leidschendam, the Netherlands.

Mosdell T., Leatt A., 2005. On tap: A review of the Free Basic Water Policy. In: Leatt A and Rosa S (eds) *towards a means to live: targeting poverty alleviation to make children's rights real*. Cape Town: Children's Institute (UCT).

Moyo S., Wright J., Ndamba J., Gundry S., 2004. Realising the maximum health benefits from water quality improvements in the home: a case from Zaka district, Zimbabwe. Physics and Chemistry of the Earth. **29**: 1295-1299.

Murray P.R., Rosenthal K.S., Pfaller M.A., 2005. Parasites. Medical Microbiology, Fifth Edition. United States: Elsevier Mosby.

Muyima NYO., Ngcakani F., 1997. Indicator bacteria and regrowth potential of the drinking water in Alice, Eastern Cape. Water South Africa. **24** (1): 194-202.

Nala N.P., Jagals P., Joubert G., 2003. The effect of a water-hygiene educational program on the microbiological quality of container-stored water in households. Water SA. **29** (2): 171-176.

Norby J., 2000. The origin and meaning of the little p in pH. Trends in Biochemical sciences. **25**: 36-37.

Novak J.T., 1974. Temperature-sub interactions in biological treatment. Journal of Water Pollution Control. **46**: 1984–1994.

Payment P., Richardson L., Siemiatychi J., Dewar R., Edwards M., Franco E., 1991. A randomized trial to evaluate the risk of gastrointestinal disease due to the consumption of drinking water meeting current microbiological standards. American Journal of Public Health. **81**: 703-708.

Payment P., Franco E., Siemiatycki J., 1993. Absence of relationship between health effects due to tap water consumption and drinking water quality parameters. Journal of Water Science and technology. **27** (3): 137-143.

Payment P., Hunter P.R., 2001. Endemic and epidemic infectious intestinal disease and its relationship to drinking water. In Fewtrell L and Bartram J (eds.) Water quality: guidelines, standards and health. IWA, London. 61-88.

Percival S.L., Walker J.T., Hunter P.R., 2000. Microbiological aspects of biofilms and drinking water. CRC Press, Boca raton, Florida.

Petersen E.A., Roberts L., Toole M.J., Peterson D.E., 1998. The effect of soap distribution on diarrhoea: Nyamithuthu refugee camp. International Journal of Epidemiology. **27**: 520-524.

Pinfold J.V., 1990. Faecal contamination of water and fingertip rinses as a method for evaluating the effect of low-cost water supply and sanitation activities on faeco-oral disease transmission. A case study in rural north-east Thailand. Epidemiological Infection. **105**: 363–375.

Pinfold J.V., Horan N.J., Mara D.D., 1991. Seasonal effects on the reported incidence of acute diarrhoeal disease in northeast Thailand. International Journal of Epidemiology. **20**: 778-786.

Pirt S.J., 1971. Principles of Microbe and Cell Cultivation. Blackwell. London.

Power K.N., Naggy L.A., 1999. Relationships between bacterial regrowth and some physical and chemical parameters within Sydney's drinking water distribution system. Water Research. **33**: 741 – 750.

Prescott L.M., Harley J.P., Klein D.A., 1993. Microbiology. 2nd edition. W.C Brown Publishers, Dubuque IA.

Qadri M., H., Al-Ghamdi M.A., Musharaf A.Y., Haq M.I., 1992. A study on diarrhoeal diseases in children under five years of age. Saudi Medicine. **12**: 459–462.

Quick R.E., 1997. A new strategy for waterborne disease prevention. Paper presented at the 23rd WEDC conference, Durban, South Africa.

Quick R.E., Venczel L.V., Mintz E.D., Soleto L., Aparicio J., Gironaz M., Hutwagner L., Greene K., Bopp C., Maloney K., Chavez D., Sobsey M., Tauxe R.V., 1999. Diarrhoea prevention in Bolivia through point-of-use water treatment and safe storage: a promising new strategy. Epidemiological Infections. **122**: 83–90.

Roberts L., Chartier O., Malenga G., Toole M., Rodka H., 2001. Keeping clean water in a Malawi refugee camp: a randomized intervention trial. Bulletin World Health Organisation. **79**: 280-287.

Rompre A., Servais P., Baudart J., Renee de-Roubin A., Laurent P., 2002. Detection and enumeration of coliforms in drinking water; current methods and emerging approaches. Journal of microbiological Methods. **49** (1): 31-54

Root G.P.M., 2001. Sanitation, community environments, and childhood diarrhoea in rural Zimbabwe. Journal of Health and Population Nutrition. **19** (2): 73–82.

Sack R.B., 1997. Antimicrobial resistance in organisms causing diarrhoeal disease. Clinical Infectious Disease. **24** (1): S102-S105.

Sack R., Sack G., Nair A., Siddique A., 2003. Cholera. The lancet. 363 (9404): 223-233D.

Sathasivan A., Ohgaki S., Yamamoto K., Kamiko N., 1997. Role of inorganic phosphorous in controlling regrowth in water distribution system. Water Science and Technology. **35**: 37-44.

Sartory D.P., Howard L., 2008. A medium detecting β -glucuonidase for the simultaneous membrane filtration enumeration of *E. coli* and coliforms from drinking water. Letters in Applied Microbiology. Volume 15 (6): 273-276.

Shiffman M.A., Schneider R., Faigenblum J., Helms R., Turner A., 1978. Field studies on water, sanitation and health education in relation to health status in Central America. Water Technology. **11**: 143-150.

SUDEA., 2000. Ecosan – Economy, ecology, sanitation. Society for Urban Development in East Africa. <u>http://user.tninet.se</u> (Accessed 25 June 2007).

Swerdlow D.L., Mintz E.D., Rodriguez M., Tejada E., Ocampo C., Espejo L., Greene K.D., Saldana W., Seminario L., Tauxe R.V., Wells J.G., Bean N.H., Ries A.A., Pollack M., Vertiz B., Blake P.A., 1992. Waterborne transmission of epidemic cholera in Trujillo, Peru: lessons for a continent at risk. Lancet. **340**: 28–32.

Swerdlow D.L., Malenga G., Begkoyian G., Nyangulu D., Toole M., Waldman R.J., Puhr D.N.D., Tauxe R.V., 1997. Epidemic cholera among refugees in Malawi, Africa: treatment and transmission. Epidemiology of Infections. **118**, 207–214.

Tokajian S., Sunna N., Evison L. Hahwa F., 2000. Microbial regrowth in household water storage tanks. First World Water Congress of the International Water Association (IWA). Health-related water microbiology. Paris. 94-95.

Trevett A.F., 2003. The public health significance of drinking water quality deterioration in rural Honduran communities. PhD thesis. Cranfield University, UK.

Trevett A.F., Carter R.C., Tyrrel S.F., 2004. Water quality deterioration: a study of household drinking-water quality in rural Honduras. International Journal of Environmental Health. **14**: 273-283.

Trevett A.F., Carter R.C., Tyrrel S.F., 2005. The importance of domestic water quality management in the context of faecal-oral disease transmission. Journal of Water and Health. **3** (3): 221-228.

Trevett A.F., Carter R., 2008. Targeting appropriate interventions to minimise deterioration of drinking-water quality in developing countries. Journal of Health and Population Nutrition. **26** (2): 125-138.

Tuttle J., Ries A.A., Chimba R.M., Perera C.U., Bean N.H., Griffin P.M., 1995. Antimicrobial resistant epidemic *Shigella dysenteriae* Type 1 in Zambia: modes of transmission. Journal of Infectious Disease. **171**: 371–375.

United Nations Children Foundation (UNICEF), 2007. Millennium goal tracker. Indicator 30. Proportion of population with sustainable access to an improved water source, urban and rural (UNICEF, WHO)'In United Nations Statistics Division, Millennium indicators database.

http://unstats.un.org/unsd/mi/mi_indicator_xrxx.asp?ind_code=30 (Accessed 13 April 2008).

VanDerslice J., Briscoe J., 1993. All coliforms are not created equal: A comparison of the effects of water source and in-house water contamination on infantile diarrhoeal disease. Water Resource Research. **29**: 1983–1985.

VanDerslice J., Briscoe J., 1995. Environmental interventions in developing countries: Interactions and their implications. American Journal of Epidemiology. **141** (2): 135–144.

Verweij, P.E., van Egmond, M., Bac, D.J., van der Schroeff, J.G., Mouton, R.P., 1991. Hygiene, skin infections and types of water supply in Venda, South Africa. Tropical Medicine and Hygiene. **85**: 681-684.

Warby V., 2007. South Africa's water saving campaign. Statistics South Africa.

Water Aid., 2008. Diarrhoeal Disease. www.wateraid.org. 2008. (Accessed 13 April 2008).

Water Research Commission., 1998. The National Water Act of 1998. www.wrc.co.za (Accessed 12 February 2009).

Wilson M.E., 2005. Diarrhoea in non-travellers: risk and aetiology. Clinical Infectious Disease. **41** (8): 541-546.

Winblad U., 1996. Towards an ecological approach to sanitation. In: Simpson-Hébert, M. and Wood, S. (eds.) Sanitation promotion. World Health Organization, Geneva, Switzerland. 203-214.

Wright JA., Gundry SW., Conroy R., 2004. Household drinking water in developing countries: a systematic review of microbiological contamination between source and point-of-use. Tropical Medicine and International Health. **9**: 106-117.

World Health Organisation., 1979. Sodium chloride and conductivity in drinking water. World Health Organisation, Copenhagen.

www.who.int (Accessed 12 August 2006).

World Health Organisation (WHO), 1993. World Health Organization Guidelines for Drinking Water Quality. Second edition. World Health Organization, Geneva.

World Health Organisation., 1996. Guidelines for drinking water quality. 2nd edition, Volume 2-Health criteria and other supporting information. World health Organisation, Geneva.

World Health Organisation., 1997. Environmental Health newsletter, no 27. October 27.World Health Organisation.www.who.int (Accessed 19 August 2007)

World Health Organisation., 2001. Water Quality: Guidelines, Standards and Health. Edited by Lorna Fewtrell and Jamie Bartram. IWA publishing, London, UK. ISBN: 1900222280.

World Health Organization (WHO), 2005. World Health Organization Guidelines for Drinking Water Quality, recommendations. Volume 1. Third edition. World Health Organization, Geneva.

Xavier J.B., Picioreanu C., Rani S.A., Loosdrecht M.C.M.V., Stewart P.S., 2005. Biofilm control strategies based on enzymatic disruption of the extracellular polymeric substance matrix – a modelling structure. **151**: 3817-3832.

Yeager B.A., C., Lanata C.F., Lazo F., Verastegui H., Black R.E., 1991. Transmission factors and socio-economic status as determinants of diarrhoeal incidence in Lima, Peru. Journal of Diarrhoeal Disease. **9**: 186–193.

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Appendix

Appendix A

An investigation into the effect of season on the quality of point-of-use drinking water from ground tanks in peri urban Cato Manor and Sawpitts: Durban, South Africa

Abstract

The aim of this study was to evaluate the effect of season on the quality of drinking water supplied to ground tanks from the peri-urban settlements; Cato Manor, which was sampled during winter months and Sawpitts, which was sampled during summer months. The effect of season was expected to be mediated predominantly via lower ambient temperatures in winter than in summer. Ground tank-supplied point-of-use drinking water sampled during winter (from Cato Manor)) and summer (from Sawpitts), had significantly higher heterotrophic organism, total coliform and E. coli counts than its corresponding source drinking water samples. An increase in air temperature resulted in increase in water temperature at the point-of-use. Increased water temperature was correlated with a reduction in point-of-use residual chlorine levels and an increase in heterotrophic organism and total coliform counts. Overall, point-of-use drinking water sampled in summer (Sawpitts) had higher heterotrophic organism and total coliform counts than point-of-use drinking water sampled in winter (Cato Manor). Therefore, season appears to affect point-of-use water quality. However, point-of-use water in winter had higher *E. coli* counts than point-of-use water in summer. This could be due to improper hygiene, sanitation and water-use practices at the point-of-use in Cato Manor (winter), or to longer survival of E. coli introduced to point-of-use water at lower temperature.

Introduction

One of the most important factors influencing bacterial growth in water is temperature (LeChevallier *et al.* 1980; Donlan and Pipes *et al.* 1988; LeChevallier *et al.* 1996; Donlan

et al. 1994; Momba and Notshe, 2003). Temperature has been reported to be one of the major factors controlling the regrowth of microorganisms in household tanks (Tokajian *et al.* 2000). The biomass composition of microorganisms requires external assistance, such as high temperatures, in order to be regulated. Nutrient requirements and rate of metabolism are also governed by temperature (Pirt, 1971; Novak, 1974; Esener *et al.*, 1981). Microorganisms are incapable of regulating their own internal temperature and are dependant on temperature to influence their biomass composition, nutrient requirements, nature of metabolism and their rate of metabolic reactions (Pirt, 1971, Novak, 1974 and Esener *et al.*, 1981). It may therefore be expected that at lower temperatures *i.e.* in winter, the rate at which organisms utilize substrates would be lower, and hence microbial water quality at the point-of-use would be better in winter than in summer. The current study aimed to investigate the effect of season on water quality from ground tanks in rural and peri-urban areas in Durban, South Africa since both average summer and winter temperatures are above the required temperatures to increase microbial growth and regrowth.

This study made use of data from an associated study by Maraj (2005), who investigated the microbial deterioration of stored water supplied by standpipes and ground tanks in winter in Cato Manor, a peri-urban settlement in Durban, South Africa. Maraj also made use of a questionnaire survey to gain insight of community water-use behaviour and practices. All data in this study pertaining to the Cato Manor study is used here with permission from the author. Cato Manor ground tanks were sampled during the winter months of July and August in 2005. In this study, winter data for microbial and physico-chemical attributes of ground tank drinking water from Cato Manor was compared to summer data for microbial and physico-chemical attributes of ground tank drinking water from Sawpitts, another peri-urban settlement in Durban, South Africa. The same communities could not be used in this study due to the mobility of households and household members in Cato Manor. Such mobility meant that communities in Cato Manor sampled by Maraj (2005) had changed significantly by the time of the 2007 study. Since considerable background data had been collected for the Sawpitts community and since each house could be located precisely on the basis of recent GPS readings it was

decided that the study of Lutchminarayan (2007) would be used to source summer data instead. The same source was used for the multifactorial analysis of contributors to point-of-use water quality presented in Chapter 3. Even though different communities were used, they appeared to be similar to each other in terms of infrastructure and general living conditions. However, it is acknowledged that the comparison of different communities is a weakness of the study presented in this chapter.

Material and methods

Study site

Two sites in the Durban region, supplied by eThekwini Municipality, were sampled. These are:

- The peri-urban settlement, Cato-Manor, supplied drinking water via ground tanks, sampled and analysed in 2005 (Maraj *et al.* 2005).
- Sawpitts, a peri-urban settlement, supplied drinking water by ground tanks, sampled and analysed in 2007 from households surveyed in 2006 for social and health effects data by Lutchminarayan.

Areas were chosen on the basis that they are both supplied with water by ground tanks provided by eThekwini Municipality. Houses were randomly selected on the basis of location (peri-urban Sawpitts and peri-urban Cato Manor) and water delivery supply (ground tanks). Water samples were collected in duplicate from 27 ground tanks in Cato Manor and 28 ground tanks in Sawpitts over a 10 day period and were analysed daily. Turbidity, pH, conductivity, water temperature, air temperature, residual and total chlorine levels were determined. Microbiological analyses included heterotrophic plate counts, total coliforms and *E. coli* counts. Summer sampling in Sawpitts was conducted in March 2007 and winter sampling in Cato Manor was conducted in July 2004 according to eThekwini Municipality standard methods. eThekwini Municipality Laboratory tap water was assumed to be representative of water in the distribution system, hence representative of the quality of source water for ground tanks from Cato Manor and Sawpitts.

Statistics

Data were statistically analysed using SPSS statistical package, version 15. Three microbial parameters were measured in drinking water (*E. coli*, total coliforms and heterotrophic organisms). Because distribution of microbiological data is typically skewed, log_{10} of all microbiological data was calculated. A value of 1 was added to all data before taking the log_{10} to account for zero values. This enabled generation of Log graphs which presented data in a more logical manner due to data variability.

All physico-chemical and microbiological data were tested for normality of distribution using the Kolmogrov-Smirnov test. Transforming of microbial data was unsuccessful in obtaining normal distribution due to the extreme variability amongst microbial counts. Microbial data was therefore ranked and subject to non parametric analysis of variance using the Kruskall-Wallis tests with post hoc testing using the Tukey test in order to determine the relationship between seasons (winter and summer) and the microbial quality of drinking water at the point-of-use.

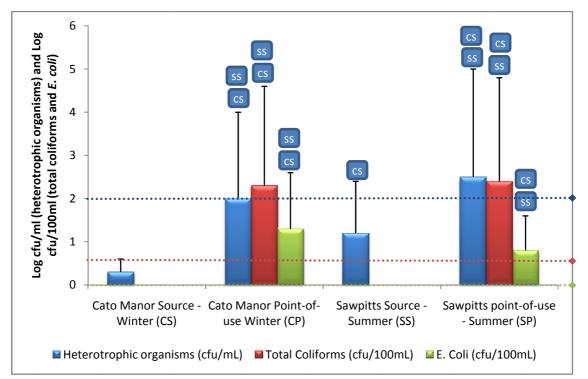
Physico-chemical data were tested for significance of differences by one-way-Analysis of variance (ANOVA) coupled with least significant difference (LSD) post-hoc tests. A 95% confidence interval was used. Spearman's rank correlation was also performed on all microbial and physico-chemical data to determine which parameters showed associations.

Results

The arithmetic mean microbial levels of heterotrophic organisms, total coliforms and *E. coli* from ground tank point-of-use (water from ground tanks) and source (water from tap at eThekwini Municipality Laboratories) drinking water, sampled in winter and summer, are shown in Figure 1.

Ground tank source water had significantly lower microbial counts (heterotrophic organisms, total coliforms and *E. coli*) than its corresponding point-of-use drinking water sampled in summer from Sawpitts and in winter from Cato Manor (p=0.035). Source water in summer had significantly higher heterotrophic organism counts than source water in winter (p=0.013).

Point-of-use water in summer had slightly higher heterotrophic organism and total coliform counts than point-of-use water in winter. Point-of-use water in winter had higher, but not significant, *E. coli* counts than point-of-use water in summer. Season could be associated to the microbial quality of drinking water in two parameters (heterotrophic organisms and total coliforms). The greater *E. coli* contamination of point-of-use water in winter is more likely to be due to faecal contamination of drinking water as a result of poor hygiene or sanitation practices as described in Chapter 5. Blue boxes indicate where microbial counts are significantly higher in comparison to microbial counts in other source and point-of-use water.



*Error bars represent a 95% confidence interval. Blue boxes represent significantly higher microbial counts than the sampling point referred to in the box.

Figure 1. Arithmetic mean microbial levels in source and point-of-use water sampled from ground tanks in Cato manor, during winter, and Sawpitts, during summer.

Source water in winter (Cato Manor) had a significantly higher residual chlorine and conductivity levels than source water in summer (Sawpitts) (p=0.025). Turbidity levels were higher in winter (Cato Manor) source water than in summer (Sawpitts), but not

significantly, whilst total chlorine levels were slightly, but not significantly, higher in Sawpitts source water (p=0.059). Point-of-use water in summer (Sawpitts) showed significantly higher pH than point-of-use water in winter (Cato Manor). Point-of-use water in winter displayed significantly higher total chlorine, residual chlorine, turbidity and conductivity than point-of-use water in summer (p=0.023) (Table 1). Point-of-use water in summer (Sawpitts) and winter (Cato Manor) had higher water temperatures than air temperatures. The converse applies for temperature of source water (sampled from a tap at eThekwini Municipality laboratories).

Table 1. Mean air and water temperature (°C), pH, residual Cl (mg/L), total Cl (mg/L), turbidity (NTU) and conductivity (mS/m) for all source and point-of-use water samples. Standard deviation and number of samples are given.

Area		Air temp	Water temp	рН	Residual Cl	Total Cl	Turbidity	Conductivity
Winter	Mean	23	23	8.13	0.2	0.4	0.5	25.84
Source	n	9	9	8	9	9	8	8
	SD	0.44	0.37	0.16	0.11	0.22	0.59	6.43
Winter	Mean	23	25	7.99	0.3	0.5	0.66	14.91
point-of- use	n	267	267	267	267	267	267	267
use	SD	0.68	0.85	0.14	0.17	0.19	0.6	1.82
Summer source	Mean	30	26	7.56	0.1	0.5	0.02	10.24
Source	n	40	40	40	40	40	40	40
	SD	0.15	0.14	0.25	0.04	0.35	0.02	1.23
Summer	Mean	30	32	8.08	0.07	0.1	0.5	12.41
point-of- use	n	280	280	183	183	183	183	183
ube	SD	0.59	0.65	0.35	0.08	0.09	0.99	1.27

^an represents the sample size.

^bSD represents the standard deviation.

The following positive correlations were observed using Spearman's rank correlation test:

- An increase in water temperature correlated positively with an increase in corresponding air temperatures (p<0.001).
- Increases in air and water temperature correlated positively with decreases in free and total chlorine levels in drinking water samples (p<0.001). Therefore temperature affected physico-chemical characteristics of drinking water by reducing the disinfectant levels which probably allows for proliferation and survival of microorganisms since limited disinfectant to destroy them is present.
- The increase of heterotrophic organisms and total coliforms correlated positively with corresponding increase in water temperature (p<0.001 for both, heterotrophic organisms and total coliforms).
- Increased total coliform counts correlated positively with increase in *E. coli* counts (p<0.001). This suggests that total coliforms are most likely of faecal origin.
- A decrease in residual and total chlorine content correlated positively with an increase in total coliform and heterotrophic organism counts (p<0.001). This suggests that proliferation, growth, regrowth and survival of these microbes are favoured when the disinfectant levels are low.

Therefore temperature shows a positive association with microbial content and physicochemical parameters of drinking water in terms of heterotrophic organisms, total coliforms, residual and total chlorine levels. Since *E. coli* counts in this study showed no association with water temperature, its presence is more likely to be the result of faecal contamination of ground tanks associated with poor personal hygiene and sanitation practices than through a change in physico-chemical parameters.

Discussion

Studies in the past demonstrated that for season (via ambient temperature) to affect the microbial quality of drinking water, temperatures should be above 15°C (LeChevallier,

1996; Power and Naggy, 1999). Cato Manor (sampled in winter) and Sawpitts (sampled in summer) showed temperatures that were above 15°C in both seasons. The maximum winter air temperatures ranged between 21-26°C whilst maximum summer air temperatures ranged between 29-32°C. This indicated a very mild winter since the mean temperature over the last 5 years was 18-21°C. An increase in air temperature correlated positively with an increase in water temperature. This could be due to the location of ground tanks, which are placed outside dwellings and are commonly exposed to sunlight. Therefore temperature of water inside tanks is likely to fluctuate with changes in ambient temperatures, as was observed in the current study.

An increase in air and water temperature also correlates positively with an increase in heterotrophic organism and total coliform counts. This was an expected result, since previous studies showed that increases in temperature favour microbial growth and proliferation (LeChevallier *et al.* 1980; Donlan and Pipes, 1988; Donlan *et al.* 1994; LeChevallier *et al.* 1996; Momba and Notshe, 2003). Bacteria are unable to regulate their own internal temperature and require high temperature to facilitate substrate utilization, nutrient uptake, metabolism and biomass composition (Pirt *et al.* 1971; Novak *et al.* 1974; Esener *et al.* 1981). The high heterotrophic organism and total coliform counts observed in point-of-use water in summer is therefore likely to be due to high temperatures which facilitated and favoured metabolic processes that increased growth and proliferation of microorganisms.

Microbial counts of point-of-use water in winter were also relatively high when compared to point-of-use water in summer. This could also be attributed to increase in water temperatures as a result of relatively high ambient air temperatures in winter. Even though temperatures in summer were higher than that in winter, winter ambient temperatures and hence ground tank water temperature were all above 21°C. Microbial growth and proliferation has been shown to increase at temperatures above 15°C (LeChevallier *et al.* 1980; Donlan and Pipes, 1988; Donlan *et al.* 1994; LeChevallier *et al.* 1996; Momba and Notshe, 2003). Therefore point-of-use water in winter would have also experienced an increase in microbial counts, as has been demonstrated. Overall, high air and thus water temperatures, affected the microbial quality of point-of-use drinking

water in both summer and winter. However, point-of-use water in summer was affected to a greater extent since temperatures were higher than in winter.

Besides the direct effect of air and water temperature on heterotrophic organism and total coliform growth and proliferation, another cause of lower point-of-use water quality in summer than winter was low residual chlorine levels in summer. Residual chlorine levels in point-of-use water in winter were just above the recommended level of 0.2-1.5mg/l (as per DWAF and WHO guidelines, Chapter 2). However, point-of-use water in summer had residual chlorine levels below the recommended levels. Low residual chlorine levels have also been shown to contribute to microbial growth and proliferation of drinking water that is stored (LeChevallier *et al.* 1996; Donlan *et al.* 1994; Momba and Notshe, 2003). It has been demonstrated that residual chlorine levels in point-of-use water increases. At high temperatures, chlorine volatilizes more readily as free Cl₂ gas. It dissociates faster and lasts for a shorter period of time in water systems and storage vessels (such as ground tanks in this instance). Therefore fewer microorganisms are destroyed and microbial contamination of drinking water at the point-of-use is not efficiently controlled. For chlorine to function optimally, temperatures below 25°C are recommended by WHO and DWAF guidelines (Chapter 2).

Ground tanks, and their associated higher water temperatures and lower residual chlorine levels, in comparison to high pressure water delivery systems, provides an ideal environment for biofilm development since water is stored in ground tanks for a 24 hour period. The growth of biofilm in tanks has been demonstrated by Tokajian *et al.* (2000) and Arjun *et al.* (2004). The growth of biofilm on ground tank walls would increase the likelihood of detachment of organisms from the biofilm into water as the tank is emptied. Since the organisms found in biofilm are unlikely to solely of specifically faecal origin, it may be expected that water drawn from ground tanks in summer would have higher heterotrophic organism and possibly total coliform counts, but similar *E. coli* counts in winter. Should an additional increase in *E. coli* be noted, this would suggest an additional source of faecal contamination from the environment. In summer heterotrophic organism and total coliform counts were high, although *E. coli* counts were minimal. It is therefore a possibility that biofilm was present in these ground tanks and that the high microbial

counts observed in point-of-use samples was due to high water temperatures and low residual chlorine levels which favoured microbial growth and proliferation in the water (planktonic) and in biofilm.

Point-of-use water in Cato Manor (winter) had higher *E. coli* counts than point-of-use water in Sawpitts (summer). This can be attributed to faecal contamination of ground tank point-of-use water in Cato Manor (winter) from the environment which can be linked to poor personal hygiene and sanitation practices or to poor water-use behaviour practices. Thus even though Sawpitts (summer) had higher total coliform and heterotrophic bacteria, as a result of high water temperatures, low residual chlorine levels and possibly biofilm formation, Cato Manor (winter) had drinking water that posed a higher risk of infection to users due to the presence of *E*. coli which is an indication of faecal contamination and the potential presence of faecal pathogens.

The use of two different locations, even though both were peri-urban, both used ground tanks as a water delivery system and both had the same infrastructure, could have impacted on results, in that point-of-use water (Cato Manor) in winter may have had significantly lower overall microbial counts and the effect of season on microbial quality of ground tank drinking water may have been more pronounced if Sawpitts was used as a sample site for both seasons. The high *E. coli* counts in point-of-use water in Cato Manor could be a possible indication of poorer hygiene and sanitation practices and poorer water-use behaviour of household members in this area.

Conclusion

- For ground tanks, the water delivery system addressed in this chapter, an increase in air temperature resulted in a corresponding increase in point-of-use water temperature.
- High point-of-use water temperatures caused a reduction in residual chlorine levels, thus leaving point-of-use water susceptible to microbial contamination, regrowth and proliferation.

- Point-of-use water in summer showed higher temperatures, lower residual chlorine levels, and associated higher heterotrophic organism and total coliform levels than point-of-use water in winter.
- Therefore season does impact on the microbial quality of point-of-use drinking water supplied by ground tanks via ambient air temperature.
- Temperature can influence the microbial quality of point-of-use drinking water directly (through increasing microorganism metabolism) or indirectly (by reducing residual chlorine levels).
- High *E. coli* counts observed in Cato Manor point-of-use water (sampled in winter) could be attributed to poor water-use behaviour, poor hygiene and sanitation practices by household members in the study.

References

Arjun N., Joubert J.P., Rodda N., Smith M and Buckley C., 2004. Water quality at pointof-use for users supplied by standpipes and ground tanks in a peri-urban community. Paper presented at IWA Specialist Group Conference on Water and Wastewater Management for Developing Countries, Victoria Falls, Zimbabwe.

Donlan R.M., and Pipes M.O., 1988. Selected drinking water characteristics and attached microbial population density. Journal of American Water Works Association. **80**: 70 – 76.

Donlan R.M., Pipes W.O., Yohe T.L., 1994. Biofilm formation on cast iron substrata in water distribution system. Water Research. **28**: 1497 – 1503.

Esener A.A., Roels J.A., and Kosen N.W.F., 1981. The influence of temperature on the maximum specific growth rate of *Klebsiella pneumoniae*. Biotechnology and bioengineering. **23**: 1401 – 1405.

LeChevallier M.W., Seidler R.J., and Evans T.M., 1980. Enumeration and characterization of standard plate count bacteria in chlorinated and raw water supplies. Applied and Environmental Microbiology. **40**: 922-930.

LeChevellier M.W., Welch N.J., Smith D.B., 1996. Full scale studies of factors related to coliform regrowth in drinking water. Applied Environmental Microbiology. **62**: 2201 – 2211.

Lutchminarayan R.D., 2007. A comparative study evaluating the health impacts of ecological sanitation interventions, water services and hygiene education programmes individually and in combinations, in eThekwini district, Durban, South Africa. Masters thesis, University of KwaZulu-Natal, Durban, submitted.

Maraj S; Rodda N., Jackson S., Buckley C., Macleod N., 2005. Microbial deterioration of stored water for users supplied by standpipes and ground tanks in a peri-urban community. Water South Africa. 1: 222-230.

Momba M.N.B., and Notshe T.L., 2003. The microbiological quality of groundwater derived drinking water after long storage in household containers in a rural community of South Africa. Journal of Water Supply: Research and Technology AQUA. **52** (1): 66-67.

Novak J.T., 1974. Temperature-sub interactions in biological treatment. Journal of Water Pollution Control. **46**: 1984–1994.

Pirt S.J., 1971. Principles of Microbe and Cell Cultivation. Blackwell. London.

Power K.N., and Naggy L.A., 1999. Relationships between bacterial regrowth and some physical and chemical parameters within Sydney's drinking water distribution system. Water Research. **33**: 741 – 750.

Tokajian S., Sunna N., Evison L. and Hahwa F., 2000. Microbial regrowth in household water storage tanks. First World Water Congress of the International Water Association (IWA). Health-related water microbiology. Paris. 94-95.

Permission for use of epidemiology database

TO: Dr Nicola Rodda,	FROM: Dr Stephen Knight
School of Biological & Conservation Sciences	Department of Public Health Medicine
Date: 01 June 2009	
Re: Permission to for Miss Urisha Singh to Survey Database for Academic Purposes.	Use the Ecological Sanitation (EcoSAN)
Dear Dr Rodda, Thank you for requesting permission for Miss Sanitation (EcoSAN) Survey Database belong Medicine for academic purposes and specific of Science Research Project entitled; "Evalua water in Peri urban settlements in Durban". T utilised for academic purposes the better, but needs to be appropriately consulted when it is that may ensue its use. Thank you for your co-operation.	ging to the Department of Public Health ally to achieve the objectives of her Masters tion of microbiological quality of drinking he more that this extensive database can be the Department of Public Health Medicine,
I agree that the database can be used for acad- appropriately acknowledged.	emic purposes as long as its use is
KNI 6HT SE Stephen Knight partment of Public Health Medicine iversity of KwaZulu-Natal Bag 7 CONGELLA 4013	Dr Nicola Rodda School of Biological & Conservation Sciences University of KwaZulu-Natal
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ETHEKWINI WATER



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University Kwa Zulu Natal School of Biological and Conservation Studies Higher Degree Committee

TO WHOM IT MAY CONCERN

This serves to confirm that Miss Urisha Singh, MSc student at the University of KwaZulu-Natal (student number 202501371), is conducting a survey of the microbiological quality of drinking water at source and at point of use in eThekwini, specifically in the areas Mntamntengayo and Sawpitts.

This study is being conducted with the consent and co-operation of eThekwini Municipality under an existing Memorandum of Agreement between eThekwini Municipality and the University of KwaZulu-Natal.

Yours sincerely

den T GOUNDEN

